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Development of a Chemical Source Apportionment Decision Support Framework for Catchment Management

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Development of a Chemical Source Apportionment Decision Support Framework for Catchment Management Sean D W Comber, Russell Smith,¹ Peter Daldorph², Michael J Gardner², Carlos Constantino², Brian Ellor³ corresponding author: B525 Portland Square, Plymouth University, Drake Circus, Plymouth, PL4 8AA, UK. Tel: +44(0)1752858974. Fax: +44(0)1752584710. Email: sean.comber@plymouth.ac.uk ¹ Westcountry Rivers Trust, Rain-Charm House, Stoke Climsland, Cornwall, PL17 8PH. ² Atkins Limited, The Hub, 500 Park Avenue, Aztec West, Bristol, BS32 4RZ, UK. ³ UK Water Industry Research, 1 Queen Anne's Gate, London SW1H 9BT, UK. **Research Article – Environmental Modeling** Abstract Art

Key words

Chemical, source apportionment, nutrients, metals, environmental modeling

34 Abstract

EU legislation, including the Water Framework Directive, has led to the application of increasingly stringent quality standards for a wide range of chemical contaminants in surface waters. This has raised the question of how to determine and to quantify the sources of such substances so that measures can be taken to address breaches of these quality standards using the polluter pays principle. Contaminants enter surface waters via a number of diffuse and point sources. Decision support tools are required to assess the relative magnitudes of these sources and to estimate the impacts of any programmes of measures. This paper describes the development and testing of a modeling framework, the Source Apportionment Geographical Information System

(SAGIS). The model uses readily available national data sets to estimate contributions of a number of nutrients (nitrogen and phosphorus), metals (copper, zinc, cadmium, lead, mercury and nickel) and organic chemicals (a phthalate and a number of polynuclear aromatic hydrocarbons) from multiple sector sources. Such a tool has not been available on a national scale previously for such a wide range of chemicals. It is intended to provide a common platform to assist stakeholders in future catchment management.

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54 Introduction

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The European Union Water Framework Directive (WFD)¹ sets criteria for water bodies to meet a defined status categorised as 'good', which requires chemical standards for 33 priority and priority hazardous substances and groups of substances to be achieved, in addition to standards for ecology, hydrology and hydromorphology. The Environmental Quality Standards (EQSs) set for these substances are generally more stringent than existing EQSs that were under other previous Directives, including the Dangerous Substances Directive.²

63

64 Historically, the principal method for improving river water quality has been to place 65 restrictive discharge consents on point source discharges, including those from 66 wastewater treatment works. In the UK, a combination of the development of a less 67 polluting industrial base as a response to more stringent regulation in the form of EU 68 Directives and downward pressure on limit values from other point sources, means 69 that for a number of substances, diffuse sources from agriculture, urban runoff, soil 70 erosion and discharges from abandoned mines make an increasingly important contribution to exceedances of EQSs.³ The WFD advocates the application of the 71 72 'polluter pays principle'¹ to ensure that any one sector is not unduly burdened with 73 the requirement to reduce discharges to meet an EQS. To plan to meet the stringent 74 standards set by the Directive it is therefore necessary to quantify the significance of 75 all sources to an EQS exceedance in any given water catchment. In many instances 76 it is likely that mitigation measures targeting multiple sources / sectors will be needed 77 to achieve good ecological status, so a clear appreciation of relative contributions is 78 essential.

79

80 In order to establish plans to comply with new EQSs it is necessary for regulators 81 and regulated alike to have tools to test and support planning decisions. A number of 82 models have been developed to predict inputs of chemicals from agricultural diffuse 83 sources; notable amongst these in the UK are PSYCHIC for soil and phosphorus and 84 NEAP-N for nitrogen. NEAP-N is a simple model created by ADAS that looks at 85 leachate from different land uses.⁴ It incorporates details down to for example a 86 livestock or fertilizer management level. Output is a visual representation of grid 87 squares over the catchment of interest, representing predicted change in 88 concentrations per hectare. The model does not include a within stream fate 89 component, so it is primarily used as a source of information on agricultural inputs of 90 nitrogen into a catchment based water quality model. PSYCHIC, The Phosphorus

91 and Sediment Yield CHaracterisation In Catchments model has been developed by 92 ADAS, NSRI, CEH-Wallingford, and the Universities of Exeter, Reading and Sheffield 93 through a Defra funded research programme.⁵ PSYCHIC was developed to model 94 phosphorus (P) and sediment transfers in agricultural systems to investigate options 95 to reduce P delivery from land to rivers, particularly through identification of hotspots 96 and an associated field-scale risk assessment exercise.

97

98 Once discharged to a watercourse, any given chemical will be subject to dilution and 99 undergo various biogeochemical processes that might be incorporated into a model. 100 Water Quality models include the United States Environmental Protection Agency 101 (USEPA) model QUAL2E;⁶ the MIKE series of models that is developed by the 102 Danish Hydraulics Institute, and the Systeme Hydrologique European (SHE)⁷; The 103 most used models by the Environment Agency of England and Wales are SIMCAT 104 and TOMCAT. SIMCAT is able to simulate a statistical distribution of discharge and 105 water quality data for multiple effluent inputs within a catchment. It is capable of 106 simulating up to 2500 random boundary conditions (also known as the Monte Carlo 107 approach), based on the input data, SIMCAT produces a distribution of results from 108 which an assessment of the impact can be made on the predicted mean and ninety-109 five percentile concentrations.⁹ SIMCAT allows for inputting decay constants based 110 on a first order decay rate and provides options for point and diffuse source inputs. In 111 addition the SIMCAT source code is sufficiently flexible to allow upgrades to include 112 such options as inputting partition coefficients to allow concentrations of, for example, 113 metals, to be split into dissolved and particulate bound concentrations.

114

115 Regulators have also sought to develop screening models for assessing pollutant 116 pressure in order to plan relevant measures on a national scale. The Scottish 117 Environmental Protection Agency (SEPA) have developed such a screening tool; the 118 Diffuse Pollution Screening Tool (DPST)¹⁰ which has drawn together large national 119 datasets for a number of different types of chemicals including metals, nutrients, 120 pesticides and sanitary determinands (BOD, ammonia). However, the focus of the 121 model is on source apportionment alone, rather than predicting in-river 122 concentrations, partly owing to the fact that there is no national water quality model 123 for Scotland, akin to that of SIMCAT for England and Wales.

124

125 The difference between source apportionment and water quality prediction is a key 126 distinction. Source apportionment models have value in risk assessment of 127 determining input loads and to some degree locations, but their application to water 128 quality modeling is not straightforward. Modeling of the mixing of inputs, taking 129 account of the variability of flow and load and incorporating appropriate processes of 130 chemical behaviour add a considerable level of complexity. There are currently few 131 models which can do this on a local or regional scale and none on a national scale. 132 One model developed to achieve catchment management of a regional scale is the 133 GREAT-ER model (Geographically-referenced Regional Exposure Assessment Tool 134 for European Rivers). The model was originally developed to assess the exposure 135 risk of new substances discharged to sewer from predominantly domestic sources, 136 but has been expanded to include diffuse and point sources within a number of 137 German catchments. GREAT-ER is a hybrid Monte-Carlo deterministic model which 138 allows a user to calculate the predicted environmental concentration (PEC) for a 139 substance, taking into account its geographical and temporal distribution, to produce 140 a statistical output and perform 'what-if' scenarios Data entry and selection is via a 141 GIS interface and In the receiving watercourse, chemical processes are represented 142 by conservative dilution, first order decay (similar to SIMCAT) or more complex 143 processes (similar to QUAL2E). Output is provided as annual statistics produced 144 longitudinally downstream and converted into PEC values, which are displayed via 145 the GIS interface. The modeling system is open-source (http://www.great-146 er.org/pages/home.cfm) with the aim of creating a live development framework. The 147 GREAT-ER model has been recently applied to the Ruhr catchment for the source 148 apportionment of zinc which includes background inputs, discharges from mining 149 activities, runoff, sewage treatment plant sources and diffuse agricultural 150 contributions.¹¹

151

152 The development of river basin management plans to meet WFD objectives requires 153 the assessment of a synthesis of local and national, ppoint and diffuse source 154 measures, this paper describes a modeling framework developed to utilise for the 155 first time, national datasets for multiple parameters including hydrology, rainfall, 156 modelled discharges of chemicals, reported discharge loads, and spatial datasets 157 including the locations of wastewater treatment works and smaller on-site works 158 often septic tanks, combined sewer overflow locations, output from diffuse pollution 159 risk models, road and river system networks combined within a GIS based modeling 160 framework to provide estimates of pollutant loads and in-river concentrations of 161 chemicals at water body scale for the whole of England and Wales.

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165 Methods

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167 The Source Apportionment-GIS (SAGIS) modeling framework integrates information 168 from multiple sources. Loads from different source sectors are derived from 169 coefficients expressed as an annual or monthly mean load with corresponding 170 standard deviation. Point sources are represented as mean and standard deviations 171 of concentrations and flow with the option of breaking down inputs into monthly 172 values to allow simulation of seasonal effects. Diffuse sources are represented as 173 mass per year, or month, per km². All loads are routed into associated river reaches 174 using one of 18 regional SIMCAT models covering England and Wales. Simcat 175 models are being developed for Scotland and so currently only loads can be derived. 176 To cover the whole of Great Britain (England, Wales and Scotland) on this basis, the 177 use of datasets with national coverage was imperative to provide both consistency of 178 approach and the ability to manage and update data. A common map projection was 179 used for all databases and mapping based on a 1km² grid of England, Wales and 180 Scotland. Such GIS mapping calibration and validation had previously been 181 undertaken as part of previous projects associated with the hydrological and diffuse 182 source components. Detailed information regarding the methodologies used to 183 calculate loads for each source is provided elsewhere.¹² However, a brief description 184 of the data and method used to derived load estimates is provided in the following 185 section.

186

Table 1 summarises the type of source (both diffuse and point) and notes whether loads to surface waterbodies were either derived via an established model or were calculated as part of this research. Figure 1 provides a schematic for the structure and key components of the SAGIS decision support framework.

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194				
Category	Source	Metals	Nutrients	Organics
Agriculture – Arable and Livestock	Diffuse	n/a ¹	PSYCHIC (P), NEAP-N models	n/a
Highway (non urban runoff)	Diffuse	Highway Agency HAWRAT model	Highway Agency HAWRAT model	Highway Agency HAWRAT model
Urban runoff	Diffuse	Calculated	Calculated	Calculated
Background erosion	Diffuse	PSYCHIC model + calculated	n/a	PSYCHIC model + calculated
Onsite wastewater treatment systems	Diffuse	Environment Agency model ² + calculated	Calculated	Calculated
Atmospheric deposition	Diffuse	Calculated	P n/a N within NEAP-N models	Calculated
Treated wastewater effluent	Point	Measured ³ and defaults	Measured and defaults	Measured and defaults
Storm tanks/combined sewer overflows	Point	Calculated	Calculated	Calculated
Industrial discharges	Point	Environment Agency Measured/reported	Measured/reported	Measured/reported
Mine water discharges	Point	Environment Agency Measured/reported	n/a	n/a

192 Table 1 Summary of methodology used to calculate loads discharged to 193 waterbodies

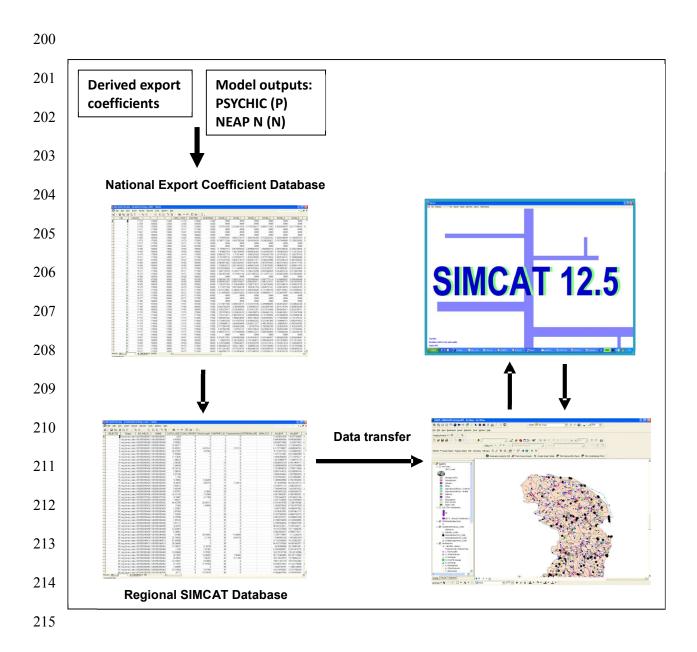
195 ¹ Included in erosion category; ² EA model used to predict locations of onsite wastewater

196 197

treatment systems ³ based on EA and Water Company data (flow and concentration);

198 199

ACS Paragon Plus Environment



216 Figure 1 Schematic diagram for SAGIS tool structure

217

Tables which list the key datasets used to derive the exported loads from each source and information as to how the datasets were used to derive the calculated loads to waterbodies are provided in Table 1 and 2 of the Supporting Information respectively.

222 Owing to WwTW inputs being significant for many substances of interest, it was 223 critical to utilise as much monitoring data as possible to derive accurate loads 224 entering receiving waters. In the absence of effluent data for WwTW, ie for WwTW 225 where there is not a requirement to determine the chemicals of interest in their 226 effluent owing to the absence of a consent, then default values were required. All 227 available concentration data were collected and collated for Environment Agency 228 monitoring between 2007 and 2009 inclusive, which added up to a maximum of over 229 2,000 results for phosphorus and nitrogen, several hundred for the metals to very few 230 for the organic determinands. In most cases a mean concentration was chosen as 231 the default value. For phosphorus because the non-consented works without effluent 232 data would not have had phosphorus reduction measures installed, a default value 233 was generated from all monitoring data reported above 2 mg-P/l. Details of the 234 monitoring data used and the default values chosen are provided in Table 3 of the 235 Supporting Information.

- 236 Inputs to SIMCAT for point sources also require a flow for WwTW effluent discharges
- in order to generate a load.
- 238
- Flows for the works were based on a number of collated data in the followinghierarchy of available data:
- 1) Measured flows and standard deviations provided by the water companies

242 2) Consented DWF

- 243
 243
 244
 3) Populations multiplied by an average flow per capita per day (from all sources) assumed to be 250 l/capita/day.
- A summary of the number of WwTW applicable to each category is provided in Table4 of the Supporting Information.

Reported literature runoff data¹⁵ was used to generate concentrations of the substances of interest in road runoff expressed as an event mean concentration (which takes account of the 'first flush' effect and subsequent reduction in concentration with increasing rainfall) and in domestic dry weather flow (Table 2).

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Concentrations used	Event Mean Concentration for Road Runoff (µg/l)	Domestic raw sewage (µg/I)
Copper	34.7	186.1
Zinc	82.5	62.6
Total PAHs	0.68	0.76
Fluoranthene	0.06	0.093
Naphthalene	0.08	0.077
Benzo(a)pyrene	0.03	0.025
Benzo(b)fluoranthene	0.03	0.030
Benzo(k)fluoranthene	0.04	0.030
Benzo(ghi)pyrene	0.03	0.025
Indeno(123-cd)pyrene	0.03	0.025
DEHP	20	33.1
Total Nitrogen (mg/l)	1.93	39.7
Total Phosphorus (mg/l)	0.27	13.9

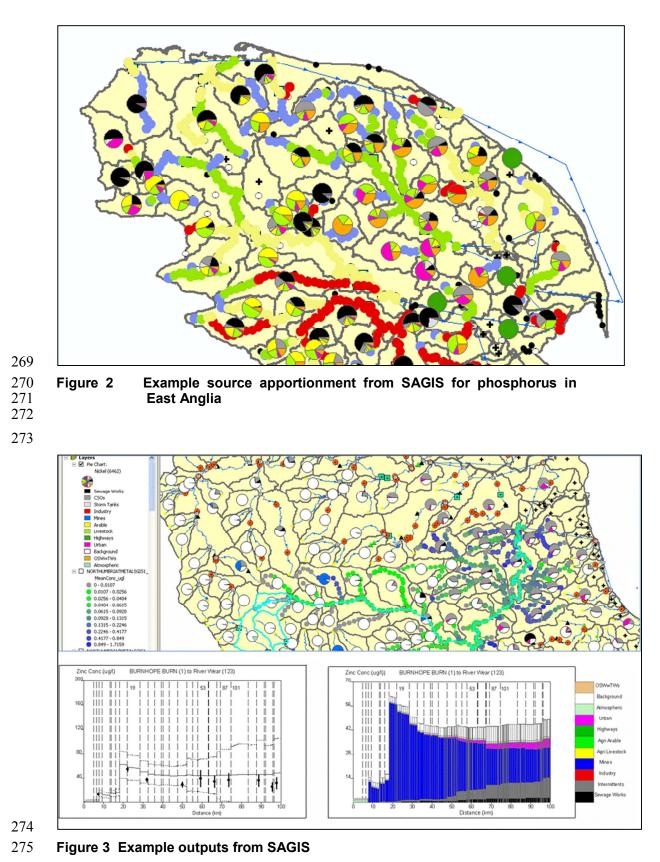
252 Table 2 Runoff and dry weather flow concentration data¹³

253

Water quality monitoring data provided by the Environment Agency and SEPA wasused for model testing purposes.

256

257 Each source was represented within a Microsoft Access™ database either as a point 258 source with an X and Y,UK national grid location coordinate or as an individual 1 km² 259 grid (approximately 150,000 for England and Wales) (Figure 1). The main database 260 is then split into 18 regional Access databases (see Figure 1 in the Supporting 261 Information) which form the attribute tables behind the features in ArcMAP 9.3™ GIS 262 software. The functionality with ArcMap and bespoke macros developed in Visual 263 Basic are then used to extract the necessary data and generate the text file required 264 to run SIMCAT, a stochastic water quality model. SIMCAT can be run from within 265 SAGIS and provide outputs (total and dissolved concentrations and loads for metals, 266 total concentrations and loads for nutrients and organics) which are fed back into 267 ArcGIS to provide cartographic, graphical and spreadsheet outputs (Figures 2 and 3). 268





277 Results

The model was initially populated with input data for total nitrogen, total phosphorus, total copper, zinc, nickel, lead, cadmium and lead, PAHs (naphthalaene, anthracene, fluoranthene, benzo-a-pyrene, benzo-b-fluoranthere, benzo-k-fluoranthene, benzoghi-perylene, indeno-123,cd-pyrene) and diethylhexylphthalate. Extension to other substances is under ongoing consideration.

283

284 The SAGIS model provides a number of outputs (Figures 2 and 3), including:

- Colour coded concentrations within the river system at 1km intervals which
 can be aligned with compliance assessment guidelines (e.g. EQS or
 ecological status under the WFD)
- Pie charts illustrating the relative contributions from all upstream sources to
 the load or concentration at the outflow from each waterbody
 (approximately 7,000 in England and Wales)
- Pie charts illustrating the relative contributions from different sources to the
 load on a waterbody basis
- Cumulative concentration from each source along a river length defined by
 the user
- Predicted versus observed concentrations where monitoring data is
 available
- 297

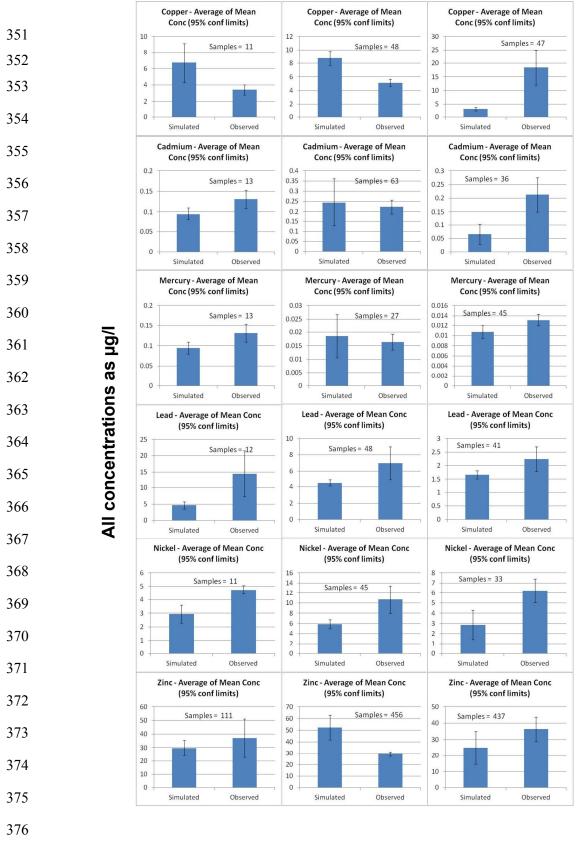
The ability to graphically present the percentage contribution from the different point and diffuse source sectors provides a very visual representation of the main contributors to loads or concentrations of a chemical to any given waterbody. Such outputs are vital in engaging stakeholders in the process of improving water quality under the Water Framework Directive.

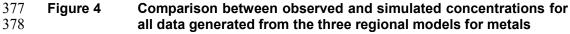
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304 When integrating models from many sources for different components of physic-305 chemical environments, with varying spatial and temporal specificity, the question of 306 parameter estimation (i.e. accuracy of export coefficients) and potential error 307 propagation becomes paramount. To test the predictive skill of SAGIS based on the 308 National / default data layers a comparison between predicted and measured 309 concentration data was undertaken for three selected catchments of differing 310 typology (mine dominated – river Wear in the NE of England, part of the 311 Northumbrian model and mixed, urban – river Tame Midlands, part of the river Trent 312 model and rural river Avon part of the SW England model) (see Figure 2 in the 313 Supporting Information).

314	
315	The process aimed to:
316	Identify any data transfer errors in the development of the export coefficient
317	databases and the input of these to SIMCAT
318	 Identify where the default export coefficients based on national datasets do
319	not provide a good representation in individual catchments and may require
320	modification; either to the methodologies, the associated assumptions or the
321	underlying data
322	To test the accuracy of the default partition coefficients in predicting dissolved
323	metal concentrations.
324	 Identify any underlying uncertainties that may affect the performance in the
325	model
326	Identify key improvements that might be made to the tool to improve the
327	predictive capability of the model at the National and catchment scale.
328	
329	The SAGIS model has been developed for use at a waterbody and catchment spatial
330	scale and consequently has utilised nationally derived default assumptions where
331	measured data is not available, Finer resolution may be possible, but the accuracy of
332	outputs would need to be tested at a local level, potentially requiring more detailed
333	local data.
334	
335	There are two key components of the model, the loads discharged to rivers,
336	waterbodies and catchments and their conversion into concentrations using the
337	SIMCAT water quality model. Although it would have been desirable to compare
338	measured and predicted loads entering the aquatic environment, there are no
339	national databases for measured loads for nay of the chemical parameters or input
340	sectors. There is however, an extensive national water quality monitoring database
341	held by the Environment Agency which was used for the purpose of testing the
342	SAGIS model. Furthermore, compliance testing is based on measurement of
343	concentrations against an EQS, and so it is vital that any model used by stakeholders
344	provides concentration data so that the impacts of any future measures applied to
345	improve water quality can be measured against EQS compliance.
346	
347	Figures 4 and 5 provide summary statistics for the output of the SAGIS tool
348	compared with observed data (means derived from all monitoring between 2007 and
349	2009 inclusive) for each of the three regional model areas for metals and nutrients

350 respectively.





379

380 The data in Figure 4 provide averages and 95% confidence intervals for a number of 381 metals for which concentrations can be simulated within the SAGIS programme. For 382 copper all predictions are of the same order, although simulations in the Northumbria 383 and Trent model tend to overestimate levels in the catchment slightly, with the 384 reverse for the SW model. It should be noted that for the SW model, observed values 385 were biased towards a number of highly mineralised sites in the west of the 386 catchment explaining the lower predicted value. Cadmium simulations were 387 comparable for Northumbrian and Trent regions but similar to copper, the model 388 underpredicted concentrations in the SW, for the same reasons. Mercury, being un-389 influenced by UK mineralogy, shows good comparability between predicted and 390 observed means, although it should be noted that mercury monitoring data is 391 somewhat limited owing to many reported concentrations being less than limit of 392 detection. Lead concentrations, too are generally comparable, although in this case, 393 the NE mineralogy is dominated by lead/zinc mines and so as for Cu/Cd in the SW, 394 the model underpredicts lead inputs from this region. Concentrations in the Trent and 395 SW model are comparable and low. For nickel a metal for which anthropogenic 396 inputs dominate sources, a good comparison between measured and predicted 397 means is observed. Finally, for zinc, levels are relatively high and variable owing to 398 both the ubiquitous nature of zinc and its presence in soil, minerals and 399 anthropogenic discharges. However, the comparability is still good even for the mine 400 dominated sites in the Northumbrian model.

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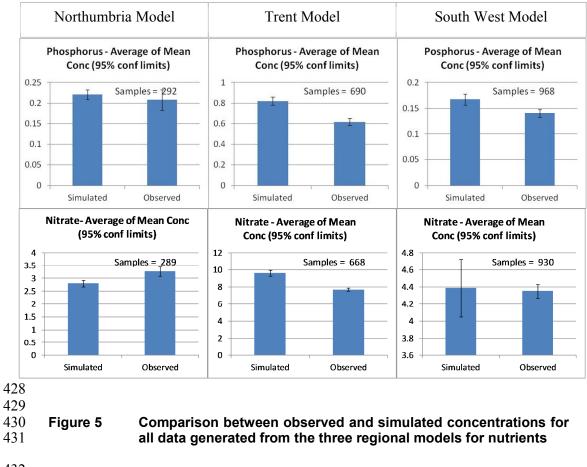
402 Overall it is clear from Figure 4 that given the nature of the generic datasets used 403 within the SAGIS model, comparisons with observed data may be considered very 404 good. Under estimates compared with measured data are associated with mineral-405 rich areas where although some point source loads are accounted for, ie major adits 406 which are monitored for flow and concentration, thus providing a load, there are 407 many diffuse inputs not accounted for from minor adits, leaching and soil loss from 408 spoil tips and old processing sites. Previous studies in the Tamar catchment, for 409 example, have highlighted the loads of metals that can potentially arise from diffuse sources, which may match or even exceed point source inputs.¹⁶ The model is 410 411 currently being updated to take account of diffuse mine inputs.

412

413 Data for nutrients (Figure 5) shows excellent agreement between predicted and 414 observed values for phosphorus and nitrate with significantly high levels in the urban Environmental Science & Technology

415 Trent region associated with discharges from wastewater treatment works (WwTW). 416 Both of these inputs are dominated by a combination of agricultural diffuse and 417 WwTW inputs. Given that agricultural inputs are derived from well developed and 418 tested models and WwTW inputs are derived from extensive monitoring data, it is 419 unsurprising that comparability is good. Furthermore, it needs to be noted that 420 although observed data are collected all year round, sampling cannot take place 421 during periods of heavy rain or flooding for safety and practicality reasons. 422 Consequently, significant loads entering the waterbodies during high rainfall events 423 particularly substances associated with particulate matter may be underestimated via 424 routine monitoring data, thus leading to an underestimation compared with predicted 425 data.

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Although in some cases there appears to be a degree of statistically significant bias
they don't indicate a bias that is likely to be important in terms of status assessment
or catchment management. This suggests that either estimates of variance in the
model are too low or that there is bias in monitoring; or both.

439

440 The model was further validated by comparing measured versus predicted 441 concentrations within the individual rivers (Wear, Avon and Tame). Examples of 442 simulated concentrations of the case study substances in the main river channel are 443 provided in the Supporting Information (Figures 3, 4 and 5). There was insufficient 444 observed information for the organic substances to allow a meaningful comparison 445 between observed data and model output. For the river Wear in Northumbria zinc 446 inputs are dominated in the upper catchment by a limited number of mine water 447 inputs from abandoned zinc/lead mines. The model predicts accurately the step 448 change in concentration 25 km downstream of the source as major mine inputs enter 449 the river. After a small amount of dilution, levels then remain at $\sim 32 \ \mu g/l$ to the tidal 450 limit. Comparison between predicted and measured values is excellent and further 451 supports the validity of the model outputs generated from summing a variety of point 452 and diffuse loads entering the river. Zinc levels in the river Tame are high owing to 453 inputs from historically contaminated land and resulting leachates (thought to be an 454 historic landfill site, EA data). Concentrations of several 100 µg/l are measured in the 455 upper catchment. These are not accurately predicted because no loads for inputs 456 from landfill or other contaminated land (with the exception of large mine sites) were 457 available on a nationwide basis. Further downstream, however, after substantial 458 dilution and where levels are influenced mainly from sewage effluent discharges (for 459 which adequate datasets are available), predictions match observed values. For the 460 agricultural catchment of the River Wylye, levels of zinc are substantially lower and 461 predictions mostly lie within the 95% confidence intervals of measured data.

462

463 A similar exercise was carried out for phosphorus (Supporting Information, Figure 4). 464 The river Wear is subject to inputs from upland low intensity livestock farming and 465 low population centres. Overall predicted concentrations track observed values from 466 low concentrations of phosphorus in the upper catchment, slowly rising downstream 467 as larger towns provide phosphorus inputs to the river via sewage works. In all cases 468 except one anomalously high observed concentration predicted values were slightly 469 in excess of measured values, potentially owing to the PSYCHIC model over 470 estimating agricultural loads or a slight bias in the relationship between loads and 471 river flow used in the model. In a catchment such as the Tame, where accurate data

for the main inputs (sewage effluents) are available, predictions versus observed values matched consistently down the catchment. A similar situation occurred for phosphorus in the Wylye/Avon, a predominantly chalk catchment dominated by arable farming, with good agreement between measured and predicted values throughout the river length. Observed and predicted concentrations of phosphorus show an increase at the 11km mark owing to a sewage treatment works input and remain relatively consistent thereafter..

479

Simulated and observed nitrate concentration in the river Wear showed excellent comparability with a gradual increase down the catchment as the contribution from sewage effluents slowly increases. A similarly close agreement was observed for the Tame, although much higher concentrations were observed and simulated (ca. 10 mg-N/I) owing to it being a sewage effluent-dominated catchment. The Wylye being a more rural catchment exhibits lower concentrations which again are influenced by an effluent discharge 11 km downstream of the source.

487

488 Overall, given the acceptable level of comparability between measured and predicted 489 concentrations it can be concluded that the model may be used for river planning 490 purposes with confidence, particularly when considering the impacts of applying 491 certain programmes of measures to meet environmental quality standards. This will 492 be of particular importance given mean catchment concentrations for Cd, Pb and Hg 493 will exceed the WFD EQS at certain sites with the potential of P, Zn and Cu also 494 exceeding UK derived limits in certain waterbodies. Cases where there is a 495 statistically significant bias between simulated and predicted data may require further 496 investigation at a local scale which is something regional Environment Agency staff 497 have been recommended to undertake.

498

499 **Discussion**

500

501 For the first time a model has be developed at a national scale which combines 502 predictions of input loads for all major sources of a contaminant with a water quality 503 model to predict in-stream concentrations for a number of determinands including 504 organics, metals and nutrients. Previously reported source apportionment models 505 have been developed at a catchment scale for single chemicals such as zinc¹¹ but 506 none has used national datasets to provide predictions over an area of approximately 507 150,000 Km² and including over 100 river catchments.

508

509 It is anticipated the framework could be used for:

- Water quality planning. SIMCAT has been widely used for water quality planning and setting of wastewater consents. All of the existing functionality of SIMCAT is retained in the tool and the diffuse and point source sector inputs can be 'switched off' and the model used in the normal way if required. By using the tool to estimate the contribution of the various diffuse and point sources, this traditional use of SIMCAT can be enhanced by providing a better indication on the sources of chemicals to improve planning of measures.
- Source control. SAGIS provides national and regional scale information on source apportionment which can be used to inform national policy on source control. For example, for organic chemicals the SAGIS could be used to indicate whether the main sources of chemicals are from controllable sources (e.g. rather than background).
- Reporting. SAGIS provides a range of visualisation options for chemical inputs
 and predicted within-stream concentrations loads which could be readily used
 for reporting of pressure characterisation and compliance for River Basin
 Management plans.
- Testing of measures. SAGIS provides a framework to test the effectiveness of
 measures related to each source sector and these can compared using the
 model output and visualisation tools.
- Catchment management stakeholder engagement. SAGIS provides an overview of the contribution of all sources of chemicals and, therefore, provides the 'big picture' for a catchment to identify the dominant sectors and sources and highlight where additional information is required or measures should be targeted. Presentation of this overview using the visualisation functionality with SAGIS provides a valuable starting point for stakeholder engagement through provide the context at the catchment scale.
- Identify further monitoring and research. By bringing together a wide range of national datasets, key areas of uncertainty in estimation of source apportionment have become clearer which could provide a focus for improving source data or the methodologies to create the export coefficient databases. By showing which sectors are likely to be important for each chemical, the tool provides a focus for where additional research and data collection would be beneficial. Without this overview, this effort might be misdirected.
- 543

544 It is important to understand that the current version of the SAGIS is based on 545 national datasets and so lacks refinement at a local level. It is estimated that for 546 catchments in excess of 50 km² confidence in the model outputs can be considered 547 good. The model provides an open framework derived from the best available 548 national data and knowledge which may easily be refined at a local level.

549

No calibration or model conditioning has been carried out on the tool at present and default values have been used in many cases; for example travel times and decay rates. For assessment of compliance with water quality standards, data refinement, calibration of decay rates and travel times (using the approach traditionally adopted for SIMCAT) or conditioning of the export coefficients need to be carried out or the assessment should take into account the difference between model output and observed data.

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558 The next step of its development is to provide it to local Environment Agency staff to 559 input data considered to be better than current default data currently in the 560 databases. Furthermore, additional substances are to be added over the coming 561 years and the databases maintained to ensure they are up to date. There will be a 562 particular focus on phosphorus in the next sets of river basin plans under the WFD 563 and so the interaction and data generated by PSYCHIC and used by SAGIS will be 564 more closely examined in the next phase of development, along with increasing the 565 coverage of the model from rivers to lakes, estuaries and coastal waters.

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567 The SAGIS model represents the first comprehensive source apportionment tool to 568 be developed on a national scale for such a wide variety of chemicals. To meet ever 569 more stringent standards multiple interventions will be required to reduce discharges 570 from point and diffuse sources. SAGIS will assist regulators in making effective 571 decisions regarding how best to meet challenging water quality targets by identifying 572 the predominant source of a chemical. It also allows practitioners from sectors such 573 as the water industry to plan future improvements within a catchment at wastewater 574 treatment works where the greatest benefits on receiving water quality will be 575 achieved. This is of particular importance for ubiquitous substances such as nutrients 576 and metals where the impact of different sources can vary throughout a catchment. 577

578 SAGIS has been shown to provide accurate predictions of in-river concentrations for 579 metals and nutrients at a catchment scale, providing a degree of confidence in the 580 predicted outputs. The Environment Agency of England and Wales in conjunction 581 with key stakeholders will be using and developing the model as part of the second 582 cycle of river basin planning from 2013 onwards.

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593	
594	Supporting Information
595	Tables of databases, default values and information used to develop the model are
596	provided in the supporting information in addition to comparison data for model
597	outputs versus observed data.
598	
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