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# Biodegraders of Large Woody Debris Across a Tidal Gradient in an Indonesian Mangrove Ecosystem

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There has been limited research on the breakdown, recycling, and flux of carbon from large woody detritus (LWD) in mangrove forests. The breakdown of LWD is caused by guilds of terrestrial and marine biodegrading organisms that degrade wood at a range of rates and efficiencies. Spatial variations in environmental factors within mangroves affect the distribution and community of biodegrading organisms, which, in turn, impacts carbon flow and sequestration. We reveal the role of biodegrading organisms in LWD breakdown and the environmental factors that influence the distribution of biodegrading guilds within a mangrove forest in South East Sulawesi that supports a diversity of mangrove species typical of Indonesian mangrove forests, which constitute 20% of Global mangrove cover. Within the high intertidal regions, terrestrial biodegradation processes dominated upon LWD. After 12 months exposure on the forest floor, experimental wooden panels in these areas remained unchanged in mass and condition. In the low intertidal region, marine wood-boring animals belonging to the family Teredinidae were the dominant biodegraders of LWD, and their activity reduces LWD volume and speeds up the loss of LWD volume. More than 50% of the experimental wooden panels' weight in these areas was lost after 12 months exposure on the forest floor. Although different biodegrading guilds occupy the same LWD niche, their distribution throughout the mangrove forest is influenced by inundation time. The change of biodegrading guilds within LWD between the terrestrial and the marine organisms was distinct, creating a biodegradation boundary in a distance as narrow as 1 m on the mangrove forest floor. These results are important, as rising sea levels have crucial implications for biodegrading guilds. A full understanding of factors affecting the biodegradation processes of LWD in mangrove forests is critical to accurately assess mangrove carbon stores and the fate of mangrove derived carbon.

**Keywords:** coastal wetlands, biodegradation, blue carbon, wood decay, fungi, shipworms, carbon cycling, sea level rise (SLR)

## INTRODUCTION

Mangrove forests form the characteristic vegetation of low wave energy tropical and subtropical shorelines. Due to their location they play an important role in land-to-sea transfer of organic carbon derived from the breakdown of vascular plant detritus (Cragg et al., 2020). Some of this detritus is brought by rivers, but much is derived from the high primary productivity of these forests (Duarte, 2017). Mangrove productivity is channeled into biomass compartments of leaves, branches, main stems and roots, with the proportion of woody tissue increasing as the tree grows, with over 70% of total biomass being composed of trunk wood in larger trees (Cintron and Schaeffer-Novelli, 1984; Ong et al., 2004). These biomass components eventually fall to the forest floor, where they are processed *in situ* or tidally exported. *In situ* processing drives the remarkably high capacity of mangrove forests to store carbon in their sediments (Donato et al., 2011; Breithaupt et al., 2012; Sanderman et al., 2018) where anoxic conditions in fine grained sediment tend to limit further breakdown (Jennerjahn, 2020; Kauffman et al., 2020). Crabs can play a major role in leaf breakdown (Thongtham et al., 2008). However, the fate of mangrove carbon derived from large woody detritus (LWD) is less understood.

Woody tissues can represent a high proportion of the above ground biomass of mangroves that increases as the tree grows (Rovai et al., 2021), but woody components are shed in increasing amounts throughout the life of the tree and eventually the tree falls to the forest floor where it will be partially submerged during the tidal cycle (Cragg et al., 2020). Measures of the woody component of litter fall at multiple sites in tropical Australia indicate that woody tissues represent 8–15% of total small litter fall (Duke et al., 1981). Litter traps are not designed to capture the rare events of falling trees or large branches, so the main input of wood onto the forest floor can only be measured by surveys of LWD, which show that quantities vary widely between sites (Allen et al., 2000; Donato et al., 2011). Kauffman et al. (2020) undertook a remarkable systematic study of organic carbon in mangrove forests at numerous sites around the world. They report that LWD (downed wood in their terminology) generally constitutes less than 10% of the above ground C store in live and dead tissue, with quantities ranging from 0 to 150 Mg C ha<sup>-1</sup>, with the upper values coming from forests where LWD quantities exceed standing biomass.

Large woody detritus enhances soil pedogenesis, provides habitat for both germinating seeds (Allen et al., 2000; Krauss et al., 2005) and animals (Hendy et al., 2013, 2014), and increases the mangrove forest nitrogen budget (Robertson and Daniel, 1989), thus contributing to ecosystem function and productivity. LWD also plays a crucial role in carbon flow and storage within mangrove ecosystems. Carbon fluxes originating from LWD within mangrove ecosystems are driven by fungi and woodborers, depending on the mangrove tidal zone (Kohlmeyer et al., 1995). These biodegrading organisms promote fragmentation of LWD (Filho et al., 2008), facilitating incorporation into anoxic sediments, where the majority of mangrove carbon typically resides (Kauffman et al., 2020) and is generally slow to degrade due to mostly anoxic conditions (Jennerjahn, 2020).

However, there are few studies on biodegradation of LWD in mangroves, and the processes of LWD biodegradation are complex and require the ability to unlock the enzyme-recalcitrant lignocellulose complex characteristic of woody plants (Cragg et al., 2020).

Rates of wood biodegradation vary between tidal zones within mangrove forests (Kohlmeyer et al., 1995). In the high intertidal region of a mangrove forest, decay of LWD is slow (Benner and Hodson, 1985) and caused mainly by white and brown rot fungi (basidiomycetes) (Kohlmeyer et al., 1995) and termites (Vane et al., 2013). In the mid to low intertidal regions, degradation of LWD is rapid (Middleton and McKee, 2001), driven mainly by teredinid bivalves commonly known as shipworms (Robertson and Daniel, 1989; Filho et al., 2008). Decay due to bacteria and ascomycete fungi also occurs in waterlogged wood, but at a much slower rate than that achieved by white and brown rot fungi (Singh et al., 2022).

Gradients of inundation and salinity may alter the distribution of biodegrading guilds as mangrove forests are habitats where terrestrial and marine influences interact. Teredinid wood-boring activity is particularly important in areas of mangrove forests with high levels of LWD (Robertson, 1990; Kohlmeyer et al., 1995). Teredinids convert fallen logs into fine fragments of fecal material (frass) that contribute to carbon cycling and reduce build-up of LWD in mangrove ecosystems (Filho et al., 2008). However, teredinids cannot tolerate the prolonged emersion that occurs in the high intertidal (Robertson, 1990). In the mid to low intertidal zones of a *Rhizophora*-dominated Australian mangrove forest where there is less organic content in the sediment, Robertson (1990) found that half of the original LWD was consumed by teredinids within 2 years, whereas in the high intertidal, where teredinids were absent, only 5% of the original mass of fallen logs was lost. Teredinids and isopod wood-borers are sensitive to salinity levels. In an extensive study of teredinids in LWD and wooden panels around Papua New Guinea, Rayner (1983) categorized the numerous teredinid species in these waters into stenohaline marine, euryhaline marine, euryhaline brackish water or stenohaline brackish water with the latter three categories occurring in mangrove ecosystems. Limnoriid crustaceans bore into wood in the intertidal zone, but are not tolerant of brackish conditions. They occur in the seaward regions of the forest investigated in this study (Cookson et al., 2012). Global-mean sea level (GMSL) has increased by approximately 1.5 mm yr<sup>-1</sup> over the 20th century (Hay et al., 2015; Dangendorf et al., 2019) and is predicted to increase by 65 ± 12 cm by 2100 (Nerem et al., 2018). How climate change and sea level rise affect the distribution and activity of LWD biodegrading organisms, and by proxy, rates of carbon flow and carbon storage in mangrove ecosystems, is an open and urgent question. In the West Pacific region, geostrophic forcing causes seawater to accumulate along the western margin of the Pacific Basin, resulting in sea levels and rates of sea level rise in Southeast Asia that are higher than the global average (Weller et al., 2016). For example, the rate of sea-level rise in coastal Indonesia from 1992 to 2015 was approximately 7 mm per year (Surya et al., 2019), and climate change models predict a sea level rise of 74 cm by 2100 (Karondia et al., 2019) which would inundate

large areas of coastal mangrove. Shifting tidal ranges will impact biodegradation of LWD through changes in the distribution of wood-boring guilds.

This study uses a mangrove ecosystem with a range of tree species and wood degrading organisms that are typical of the mesotidal mangrove forests of Indonesia (Cragg and Hendy, 2010), a country that contains 20% of the global mangrove forest area (Bunting et al., 2018). It aims to identify the organisms that initiate break down of woody detritus within four mangrove forests of Sulawesi, Indonesia, and to determine how tidal elevation, inundation and salinity affect their distribution and rates of biodegradation. Understanding the baseline structure of LWD biodegrading guilds and how they vary with tidal elevation, and how ecological conditions influence rates of biodegradation, is essential to understand carbon cycling and carbon sequestration in mangrove ecosystems.

## MATERIALS AND METHODS

### Site Description

This study was undertaken in the mangroves situated in the north of Kaledupa Island located in the Wakatobi National Park, the third largest marine park in Indonesia. Average annual air and water temperatures vary between 26 and 28°C, and salinity in the mangroves varies between 1 and 40 psu. The mangroves in this area grow in shallow water with thin carbonate-rich sediments, patches of raised fossil coral occur in the mangroves and these impact the hydrodynamics and limit the development of mangrove roots. Dominant tree species in the seaward regions of these mangrove forests are *Rhizophora stylosa*, *Sonneratia alba*, and *Bruguiera gymnorhiza*, while the landward portions of the forest were dominated by *Excoecaria agallocha*, *Xylocarpus* spp., *Rhizophora apiculata*, and *Heritiera littoralis* (see Cragg and Hendy, 2010 for detailed site descriptions).

Four mangrove forests were studied: Langira, which is situated in the low-mid intertidal zone; Sombano, which is located at the high intertidal zone and separated from Langira by a coral rock plateau; and Lulua and Galua, which both extend across the high and low intertidal zones within the mangroves (Figure 1).

Large woody detritus in the Sombano site supports guilds of terrestrial wood degraders – basidiomycete fungi, termites and beetles which are absent in LWD in Langira site, where marine wood degraders, principally teredinids, dominate. These sites were used for comparison with the Lulua and Galua sites are home to both terrestrial and marine biodegraders of LWD, making identification of the environmental variables that determine differences between distribution of biodegrading guilds possible.

### Wood Collection, Volume, and Wood Degradation Scale Classification

Five transects were established in the Sombano and Langira mangrove forests, extending from the low elevation to high elevation mangrove plant community fringes. Eight belt transects, each 20 m wide were used to characterize the biodegradation boundary in the Galua forest. Three belt transects, each 60 m wide were used for the Lulua forest.

The belt transects began in areas dominated by terrestrial wood-degrading organisms, well above the high intertidal zones and ended in areas where marine wood-degrading organisms were the only organisms present at the seaward edges (Figure 1). Transect widths were determined by operational considerations regarding limitations imposed in some areas by the density of vegetation, particularly root systems, with sampling of some areas.

The volume of LWD was measured in 22 and 19 plots in Sombano and Langira respectively. The volume of LWD in the Galua and Lulua mangrove forests was measured in 24 and 9 plots respectively. All plots measured 20 × 20 m. All pieces of LWD, defined as being more than 2 cm in diameter, were measured to estimate their volume. Large woody items were collected at low tide from the transects. Within the Sombano and Langira forests values were given describing levels of degradation as defined in Table 1.

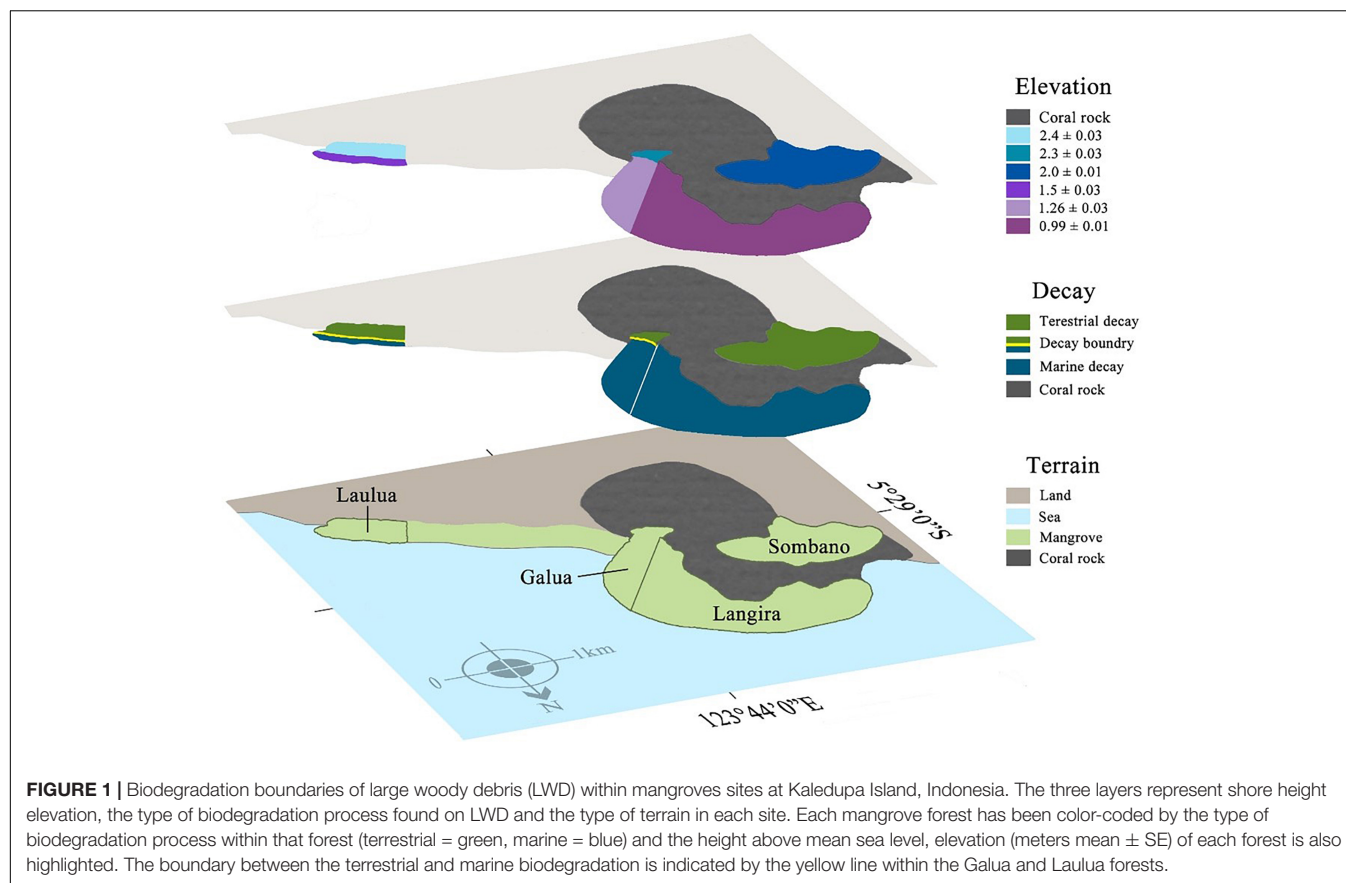
### Rate of Wood-Loss in Panels

To assess the degree of teredinid activity, 12.5 cm × 11 cm × 2 cm panels of *Bruguiera gymnorhiza* were exposed in three of the mangrove forests. Twenty-five panels were deployed within both the Sombano and Langira mangrove forests, and sixty panels within the Galua forest. Fifteen panels were also located at the sub-tidal Kal 1 reef site adjacent to the Langira forest. Before field exposure, all panels were oven dried for 48 h at 60°C, weighed and then placed in the mangrove forests for 12 months, the weight loss of wood was determined using the oven-dry weight after exposure (48 h at 60°C). Panels were split and then imaged. The percentage of the sectional area of the panel represented by the teredinid tunnels was calculated using the digital analysis package ImageTool Version 3.00 (The University of Texas Health Science Centre at San Antonio). After exposure, panels were carefully split open to allow collection of calcareous pellets, the key taxonomic characters of teredinids. Pellets were counted and used to identify the teredinid species using the key in Turner (1971).

Panels were randomly placed at 0, 20, 100, 240, 380, and 500 m from the lowest elevation in areas of the Sombano mangrove forest. Then randomly placed at 20, 120, 140, 280, and 370 m within Langira. Within Sombano, panels were located at elevations >2 m relative to mean sea level. At Langira, panels were located at >2 m, 1.99–1.9 m, and <1.9 m relative to mean sea level, and in Galua, panels were located at >2.1 m, 2.1–2.0 m, 1.99–1.90 m and <1.90 m relative to mean sea level. Panels at the Kal 1 reef site were placed at 18 m depth. To determine the elevation above sea level along each mangrove transect the level of high tide was marked on mangrove trees using high visibility string. At low tide the distance from the substrate to the mark on the tree was measured and then subtracted from the height of high tide as given in the local tide tables (Figure 1).

### Wood Degrading Guilds and Measurements of Environmental Variables

The Galua and Lulua mangrove forests were chosen to investigate the environmental variables that determine the



**TABLE 1 |** Scale describing levels of fungal wood decay observed in large woody debris (LWD) in the Sombano and Langira mangrove forests. Note, level colors correspond to **Figure 2A**.

Level	Condition and appearance of specimen
0	No visible signs of biodegradation inside or out.
1	The tough outer covering of bark reveals signs of flakiness. Some initial epiphytic fungal fruiting bodies are visible in areas.
2	Areas of bark can be pulled off by hand and are soft in areas. Fungal fruiting bodies increase. Some exposure of sapwood, with initial softening.
3	Bark is flaky, and in most areas can be pulled off with relative ease. Sapwood is easily penetrated to a depth of 2–4 cm using a knife. Sapwood begins to deteriorate revealing holes – sometimes with fungal growths.
4	Bark, if present, is soft/flaky. The sapwood has numerous holes that extend to the heartwood.
5	The bark, when visible, crumbles easily. The woody debris can appear to be ‘collapsed’. Both sapwood and heartwood are soft. Often, understory plants grow through the woody debris. In addition, basidiomycete activity upon the detrital material is common.

boundary between terrestrial and marine wood-degrading guilds at sediment level (**Figure 1**) (terrestrial degraders are found in wood out of reach of the tides even at the seaward fringe). A total of 583 and 332 pieces of LWD were examined from the Galua and Laulua forests respectively. Volumes of LWD were calculated for every piece investigated. Emersion times for the LWD were estimated by relating their tidal height to data in local tide tables.

The Galua forest has a gradually inclining shoreline extending 900 m from the seaward edge to the strandline with an elevation range of 1.36 m from the high to low intertidal. The Laulua forest extends 220 m from the strandline out to the seaward edge. From 0 to 170 m the elevation above sea level ranged from 2.57 to 2.55 m elevation above sea level. From 170 to

220 m at the seaward edge the shore height elevation was recorded at 1.32 m above sea level – a drop in shore elevation of 1.23 m in 50 m.

In areas where LWD was surveyed, ground water salinities at low tide were measured using a Bellingham and Stanley E-Line Aquatic hand-held Refractometer and expressed in Practical Salinity Units (psu). In areas without water, a small hole was dug to access the ground water. All wood-degrading organisms from the LWD samples were extracted and identified.

## Identification of Marine and Terrestrial Wood Degraders

Beetle (coleopteran) larvae, teredinid pallets and worker and soldier termites, were stored in absolute ethanol for later



identification. Using an Olympus E-510 SLR with a 50mm Macro-lens, digital images of all bracket fungi and termites, focusing on morphology and color characteristics were used for identification.

## Data Analysis

Differences in LWD volumes between mangrove forests at different tidal elevations were tested using a General Linear Model with site and tidal elevation as fixed factors. A one-way ANOVA was used to test for differences between the dry weights of panels and distance from the strandline. Regression analysis was used to determine relationships between the percentage of the cross-sectional area represented by teredinid tunnels and the level of fragmentation to each panel. Chi-Square tests were used to test for differences in teredinid settlement at different tidal elevations. A Kruskal–Wallis test was used to test for differences of wood-decay between mangrove sites. All percentage data were normalized using arcsin transformations, and count data were square root transformed. The suitability of transformations was scrutinized by examining residuals. *Post hoc* Tukey's pairwise comparison tests separated values into statistically distinct subsets in all ANOVAs using MINITAB (MINITAB Inc, version 13.20, Pennsylvania, United States).

The univariate and non-parametric multivariate techniques of the distance-based linear modeling package (DISTLM) contained in PRIMER 6.1 (PrimerE Ltd, Albany, Auckland: Plymouth Routines in Multivariate Ecological Research) was used to explore relationships between wood-degrading guilds and the environmental variables emersion time and salinity. The relationships between distributions of wood-degrading guilds and environmental variables in different forests were examined using PERMANOVA, based on Bray–Curtis matrices. The DISTLM model produces a sequential test, which assess the variation each environmental variable has on its own, and a conditional test, assessing the variation of all the variables (McArdle and Anderson, 2001). The most parsimonious model for predicting the distribution of biodegrading organisms affected by emersion time and salinity was identified using the Akaike information criterion (AIC) and  $R^2$ . AIC ranks models from all possible combinations of the environmental variables. The DISTLM was based on presence/absence data with 9999 permutations and a Sorensons transformation. Environmental variables were log transformed.

## RESULTS

### Wood Degradation Within the Sombano and Langira Mangrove Forests

Levels of fungal and bacterial decay, as indicated by softening of 168 LWD samples and other features summarized in **Table 1**, were greatest within the Sombano forest (**Figure 2A**). Woody debris beyond 400–500 m was harder. With greater distances from the landward edge, the levels of fungal wood degradation decreased (Kruskal–Wallis test, degradation rating vs. distance from land,  $P \leq 0.001$ ).

### Sombano and Langira Panel Analysis – Wood Loss

The percentage dry-weight loss of wood from panels was greatest in the Langira forest (**Figure 2B**). When emersion was more than 10 h per day, percent weight loss was low. When emersion was 10 h per day or less, weight loss was high (one-way ANOVA, distance from land vs. percentage dry weight;  $F_{11,48} = 178.75$ ,  $p \leq 0.001$ ). Wood consumption rates increased with decreasing emersion time [**Figure 2B**, one-way ANOVA emersion time (h) vs. wood loss ( $\text{g d}^{-1}$ ) per panel,  $F_{11,48} = 178.75$ ,  $p \leq 0.001$ ]. With increasing teredinid attack an increasing percentage of the whole panels broke up [Regression analysis, ( $y = 0.46x - 7.29$ )  $R^2 = 0.416$ ,  $F_{1,58} = 41.3$ ,  $p \leq 0.001$ ].

### Wood-Biodegrading Organisms on Large Woody Detritus

#### Bracket Fungi

Bracket fungi were found only in the Sombano and Galua forests (**Table 2**). Both forests had mutually exclusive mycoflora except for *Ganoderma* sp. and *Pycnoporus sanguineus*, which were found in both forests (**Figure 3A**). The mycoflora consisted of 13 species belonging to eight families. Almost all bracket fungi within the Sombano and Galua forests were found on fallen tree species that were specific hosts for particular fungal species. For example, *Schizophyllum commune* was found only in the landward regions in the Galua forest on dead *Xylocarpus granatum* trees. *Trametes elegans* was found in the Sombano forest on *Excoecaria agallocha* and *Bruguiera gymnorhiza* wood.

#### Beetle (Coleopteran) Larvae

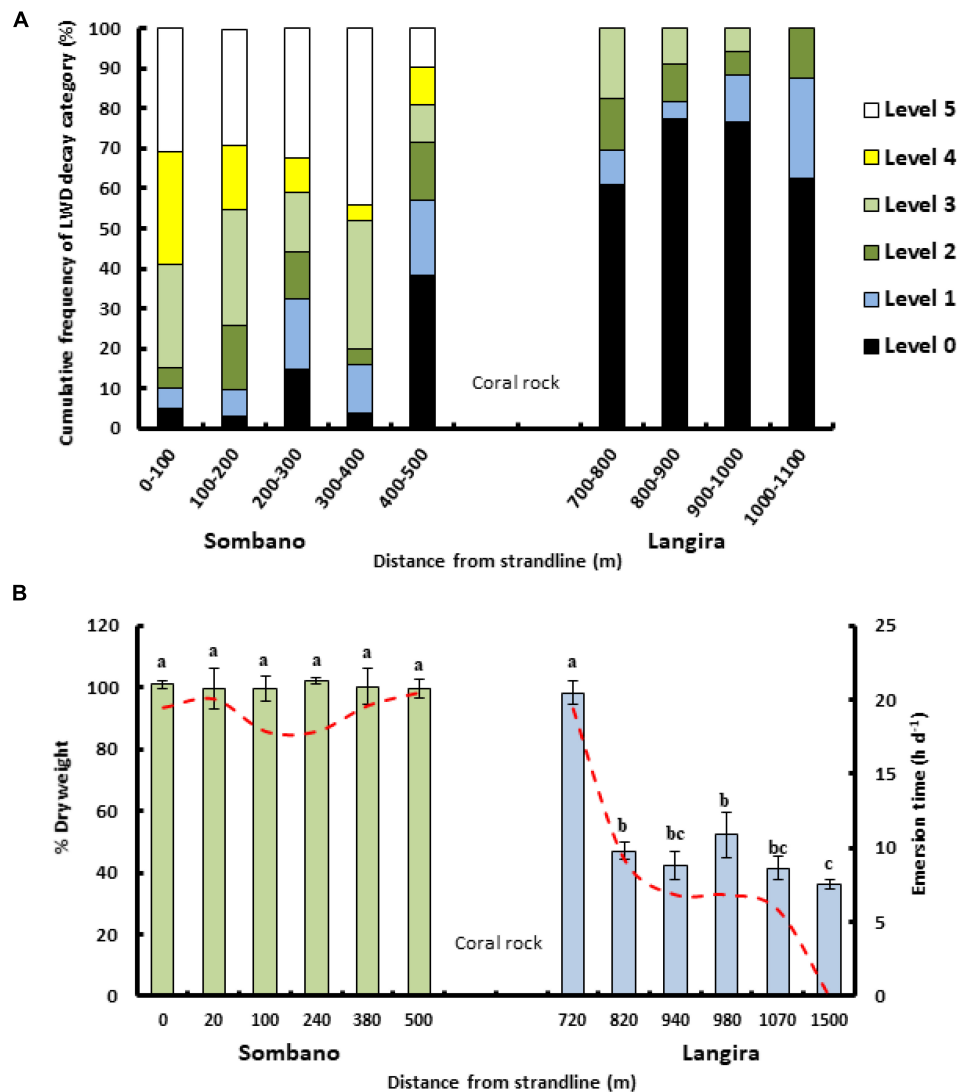
Larvae from four beetle families were found in the mid to high intertidal areas of the forests: the Cerambycidae, Elateridae, Oedemeridae, and Tenebrionidae (**Figure 3B** and **Table 2**), were found in the Sombano and Galua forests. The Oedemeridae were found in all sites except the Sombano forest. We found two items of LWD in the low intertidal containing oedemerid beetle larvae.

#### Termites

Five genera of termites from three families were found in the mid to high intertidal areas of the Sombano, Galua, and Laulua mangrove forests (**Figure 3C** and **Table 2**). Of the five genera, *Cryptotermes*, *Coptotermes*, *Microcerotermes*, and *Nasutitermes* were found in fallen wood in Galua and Laulua. *Prorhinotermes* were found only in the Galua forest. *Prorhinotermes*, *Coptotermes*, and *Nasutitermes* were found in arboreal nests above spring high tide.

#### Teredinids

Teredinids were recorded from all sites except Sombano (**Table 2**). Seven teredinid species were identified (**Figure 3D**): *Bactronophorus thoracites*, *Dicyathifer mannii*, *Teredo furcifera*, *T. fulleri*, *T. mindanensis*, *Lyrodus massa*, and *Spathoteredo obtusa*. *Teredo furcifera* were the most abundant teredinids found in the experimental panels and *S. obtusa* were the least abundant.



**FIGURE 2 |** Wood decay levels correspond to emersion time and increasing distance from the land. **(A)** The frequency (%) of fungal decay levels assigned to LWD ( $n = 168$ ) within Sombano and Langira. Decay levels correspond to the scale outlined in **Table 1**, with level 0 being the least decayed and level 5 the most decayed. **(B)** The percentage dry-weight of experimental wooden panels (mean  $\pm$  SE) exposed for 12 months within different areas of the Sombano and Langira forests. Emersion times are represented by dashed line. Wood loss is from teredinid activity. Letters above the bars = *post hoc* Tukey's pairwise test groupings.

## The Distribution of Wood-Degrading Guilds

Fallen wood in the Sombano forest was degraded by terrestrial organisms. Conversely, the Langira forest predominantly had marine wood-degraders with a few terrestrial wood degrading organisms found in the high intertidal areas. However, the Galua and Laulua mangrove forests contained both terrestrial and marine wood degrading organisms.

We found more than one type of wood-degrading organism processing wood in twenty-five items of LWD in the high intertidal region of Galua, with thirteen couplings of: termites and beetle larvae, eight items of LWD with termites and bracket fungi, and four items of LWD with beetle larvae and bracket fungi. In Laulua only one item of LWD in the high intertidal region

contained termites and beetle larvae. More than 97% of LWD items in the Galua and Laulua forests contained only a single type of biodegrading organism.

## The Volume of Large Woody Detritus in Mangrove Forests

The Sombano mangrove forest, in the high intertidal (>2 m above mean sea level) had the greatest volume of LWD per unit area (**Figure 4**, one-way ANOVA, LWD volume vs. site,  $F_{3,66} = 18.31, p \leq 0.001$ ). The volume of LWD from Sombano was  $102.9 \pm 19.5 \text{ m}^3$  (mean  $\pm$  SE) and the largest log recorded was  $9.6 \text{ m}^3$ . The volume of LWD varied at different tidal elevations (ANOVA, LWD volume vs. tidal elevation,  $F_{3,66} = 9.58, p \leq 0.001$ ). In the high intertidal areas of the Langira, Galua, and

**TABLE 2 |** Wood degrading organisms recorded within mangrove forests of Kaledupa Island, Indonesia.

Common name	Family	Species	Site	Substrate
Basidiomycetes	Auriculariidae	<i>Auricularia cornea</i>	G	Xg
	Entolomataceae	<i>Entoloma</i> sp.	S	Ea
	Ganodermataceae	<i>Ganoderma</i> sp.	S, G	Xg
	Panaeoleae	<i>Panaeolus</i> sp.	G	Xg
	Physalacriaceae	<i>Oudemansiella canari</i>	G	Xg
	Polyporaceae	<i>Favolus</i> sp.	G	Xg
		<i>Lenzites acuta</i>	S	Bg
		<i>Microporus xanthopus</i>	S	Ea
		<i>Pycnoporus sanguineus</i>	S, G	Xg
		<i>Trametes elegans</i>	S	Ea, Bg
		<i>Trametes versicolor</i>	G	Ct
	Schizophyllaceae	<i>Schizophyllum commune</i>	G	Xg
	Stereaceae	<i>Stereum ostrea</i>	G	Bg
Beetle larvae	Cerambycidae		S, G	Ea, decaying wood
	Elateridae		S	Decaying wood
	Oedemeridae		G, La, L	Decaying wood
	Tenebrionidae		S, G	Decaying wood
Termites	Kalotermitidae	<i>Cryptotermes</i> sp.	G, La	Decaying wood
	Rhinotermitidae	<i>Coptotermes</i> sp.	G, La	Arboreal
		<i>Prorhinotermes</i> sp.	G	Arboreal
		<i>Microcerotermes</i> sp.	G, La	Decaying wood
	Termitidae	<i>Nasutitermes</i> sp.	G, La	Arboreal
Limnoriids	Limnoriidae	<i>Limnoria insulae</i>	L, D	
		<i>L. pfefferi</i>	L, D	
		<i>L. sellifera</i>	L, D	
		<i>L. unicornis</i>	D	
Shipworms	Teredinidae	<i>Bactronophorus thoracites</i>	L, G	
		<i>Dicyathifer mannii</i>	G	
		<i>Lyrodus massa</i>	L, D	
		<i>Spathoteredo obtusa</i>	L, D	
		<i>Teredo fulleri</i>	L, D	
		<i>T. furcifera</i>	L, D	
		<i>T. mindanensis</i>	L, G, D	

Species of mangrove wood substrate on which organisms were found: Xg, *Xylocarpus granatum*; Ea, *Excoecaria agallocha*; Bg, *Bruguiera gymnorrhiza*; Ct, *Ceriops tagal*. Sites: G, Galua; S, Sombano; La, Laulua; L, Langira; D, Derawa.

Laulua mangrove forests, the volumes of LWD were significantly greater compared to those found toward the seaward edge. The largest logs located in the Galua and Laulua mangrove forests at >2.0 m above sea level were 1.83 m<sup>3</sup> and 0.5 m<sup>3</sup> respectively, while the largest logs found in the low intertidal from Galua and Laulua measured 0.32 m<sup>3</sup> and 0.05 m<sup>3</sup> respectively.

## Effect of Environmental Variables Upon the Distribution of Wood-Degrading Guilds

In Galua, bracket fungi were found only at the strandline, where the greatest emersion time was recorded (one-way ANOVA, emersion time vs. biodegradation process;  $F_{3,442} = 318.52$ ,  $p \leq 0.001$ ). Woody debris with mutually exclusive communities of either termites or beetle larvae were also found in these areas. There was no significant difference in emersion time of wood with termites or with beetle larvae. Similar results were

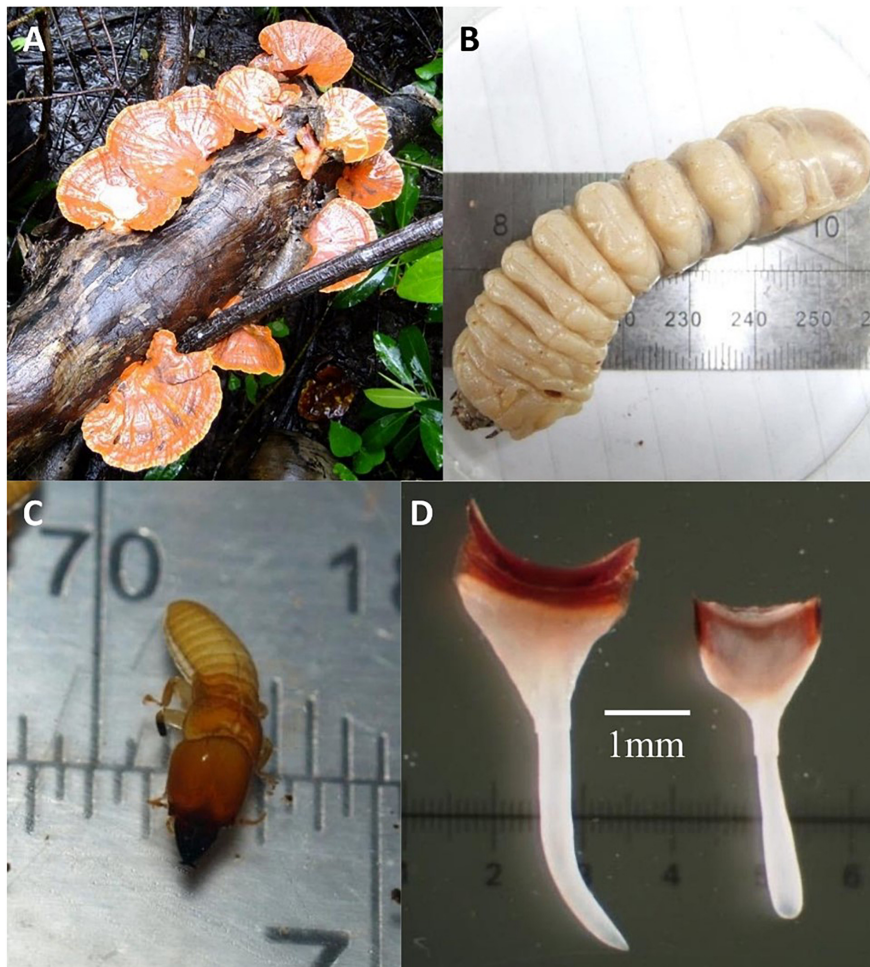
found in the Laulua forest, with LWD containing mutually exclusive communities of either termites or beetle larvae. Wood containing teredinids had significantly lower emersion times than wood containing terrestrial wood-degrading organisms (one-way ANOVA, emersion time vs. biodegradation process;  $F_{2,103} = 211.77$ ,  $p \leq 0.001$ ).

The elevation in the intertidal areas of each mangrove forest where terrestrial wood-degradation occurred was significantly higher than in areas of the forest exposed to marine biodegradation only (Figure 1, one-way ANOVA, elevation above sea level vs. intertidal zone,  $F_{5,644} = 236.84$ ,  $p \leq 0.001$ ).

## The Change Between Terrestrial and Marine Wood-Degrading Organisms

While the distance from the strandline of members of the wood degrading guilds varied between Galua and Laulua, a biodegradation-boundary marked by the sharp change between





**FIGURE 3 |** Wood degrading organisms found in mangrove forests of Kaledupa Island, Indonesia. **(A)** The bracket fungi, *Pycnoporus sanguineus* on a fallen *Xylocarpus* sp. tree in the Galua forest; **(B)** a beetle larva of the family Cerambycidae from the Galua forest; **(C)** frequently found in decaying mangrove wood, the termite *Cryptotermes* from the Galua and Laulua forests; **(D)** teredinid pallets belonging to *Lyrodus massa* removed from experimental wooden panels in the Langira mangrove forest.

terrestrial and marine wood-degrading organisms in LWD was recorded within an elevation of 0.04 m of each other in the Galua and Laulua forests. The change between the terrestrial and marine consortia in the Galua and Laulua forests was remarkably similar, which occurred at  $1.82 \pm 0.01$  (mean  $\pm$  SE) m elevation above sea level and  $1.78 \pm 0.02$  (mean  $\pm$  SE) m elevation above sea level respectively (**Figure 5**, one-way ANOVA, sea level boundary change vs. site,  $F_{1,27} = 3$ ,  $p \geq 0.05$ ).

### Comparisons of the Distribution of Wood Biodegrading Guilds Between Localities

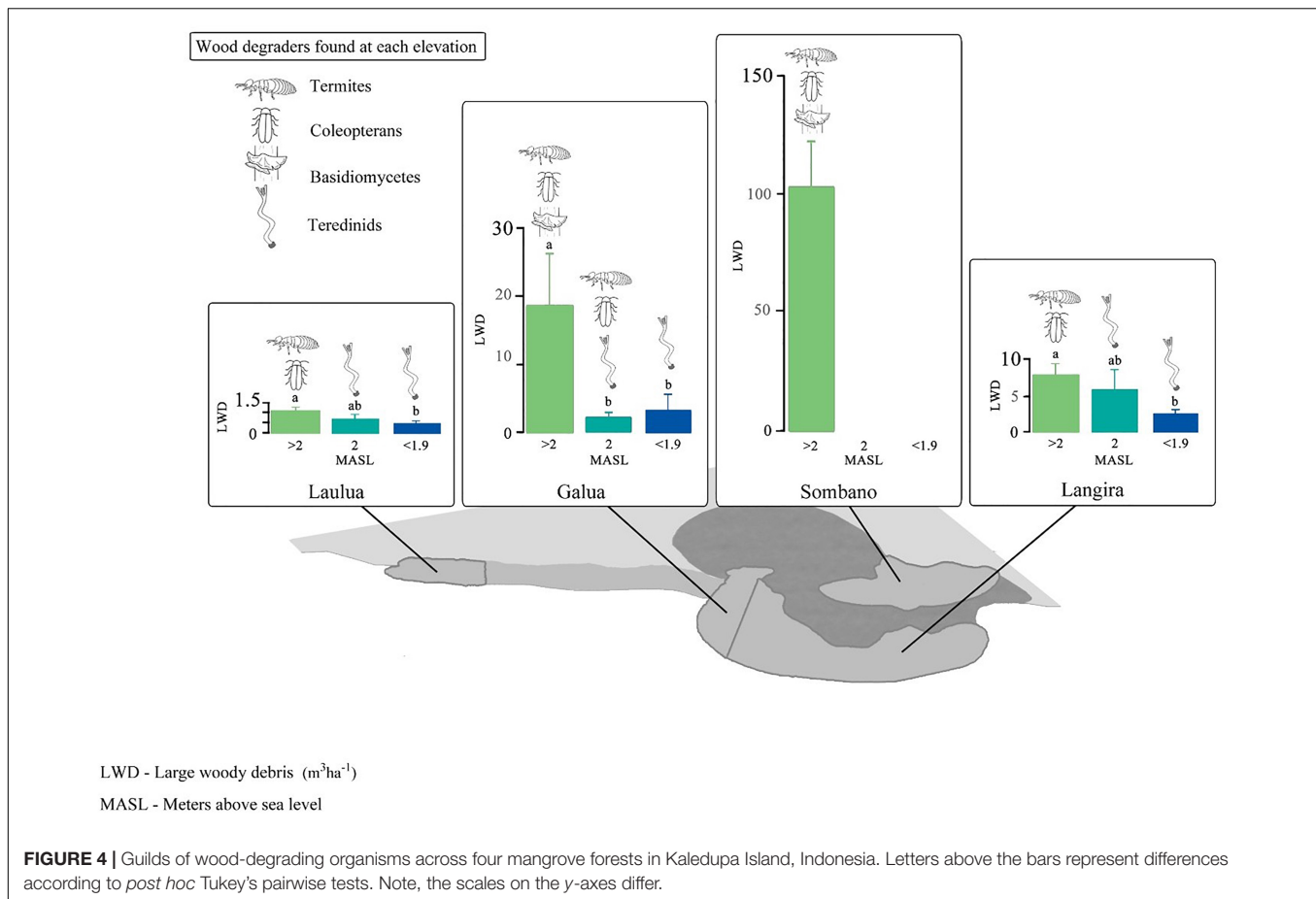
Spatial distributions of wood biodegrading guilds between the Laulua and Galua mangrove forests varied significantly within each forest (PERMANOVA Pseudo- $F_{1,550} = 56.9$ ,  $p \leq 0.001$ ). In the Laulua forest, the most parsimonious model (AIC = 785.9,  $R^2 = 0.49$ ) for the wood-degrading guilds explained 49% of the variation, with emersion explaining 42% and salinity explaining 7% (DistLM sequential test, Pseudo- $F$  ranging from 13.7 to 76;

in all cases  $P \leq 0.001$ ). Emersion explained 98.9% of the total variation, followed by salinity with 1.1%.

In the Galua forest, the most parsimonious model (AIC = 3393.5,  $R^2 = 0.36$ ) for the variation in wood-degrading consortia explained 36% of that variation, with emersion explaining 35% followed salinity explaining 1% (DistLM sequential test, Pseudo- $F$  ranging from 8.1 to 236; in all cases  $P \leq 0.001$ ). Emersion explained 98% of the total variation, followed by salinity with 2%.

### Teredinid Settlement in Panels

No differences were found in teredinid recruitment to panels placed at intertidal regions in the Galua mangrove forest (**Figure 6**, Chi-Squared,  $P \geq 0.05$ ). However, a greater proportion of teredinid activity was recorded in panels in areas below the mid intertidal. The degraded panels in the high intertidal had minimal teredinid activity and they each contained one teredinid tunnel measuring between 0.02 and 0.2 cm<sup>2</sup>. Teredinid tunnels in panels



in the mid to low intertidal regions occupied between 0.1 and 3.4  $\text{cm}^2$  of the surface area of the panels.

## DISCUSSION

### Spatial Variation in Large Woody Detritus Biodegrading Guilds

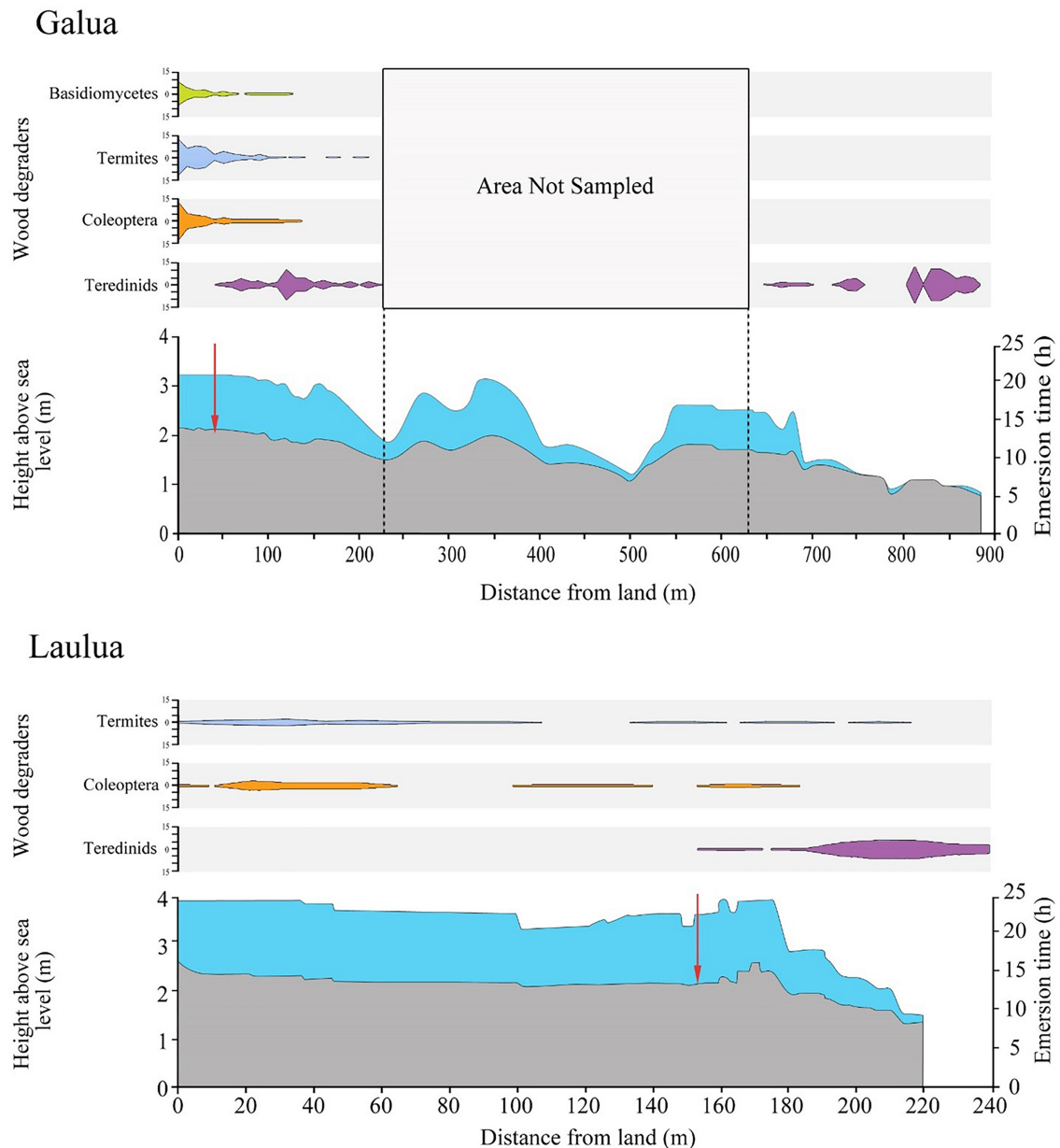
This study found significantly different volumes of LWD dependent upon the location within the mangrove forest (Figures 1, 2). Greater quantities of LWD were found in the high intertidal regions of the mangrove forests where the primary biodegradation process was microbial, and therefore slow. Further down the shore to the mid and low intertidal areas, LWD volume decreased due to rapid degradation by guilds of animals and to a lesser extent fungi and then, after fragmentation, removal by ebbing tides.

The Sombano mangrove forest does not extend below the high intertidal and maintains wood-degrading organisms that are found in terrestrial forests, including bracket fungi, beetles, and termites. In the adjacent Langira mangrove forest, the types of wood-degrading organisms were different and the elevation was lower, extending down to mean sea-level. Although both mangrove forests are adjacent to one another, salinity and immersion time differ considerably. These environmental

differences facilitated the contrast of wood-degrading guilds between both localities.

In the Sombano mangrove forest and the high intertidal region in the Galua mangrove forest, basidiomycetes were found on almost all LWD. Basidiomycetes provide habitat for many other organisms and enable the regeneration of forests and soil formation (Ananda and Sridhar, 2004; Lonsdale et al., 2008). In these areas of the forests the environmental conditions for basidiomycetes were favorable due to lower salinities and longer shorter inundation periods. In the mid to low intertidal areas were absent and the level of wood decay was decreased. In areas of the mid intertidal and below, higher salinity within the LWD and longer inundation periods may prevent establishment of basidiomycete fungi (Castillo and Demoulin, 1997; Ejechi, 2002), as found in this present study. At the sites examined in this study, there was a critical inundation level below which basidiomycete activity is negligible. Ascomycetes are not limited in the same way, they cause much less rapid decay. This type of decay cannot be readily assessed in the field.

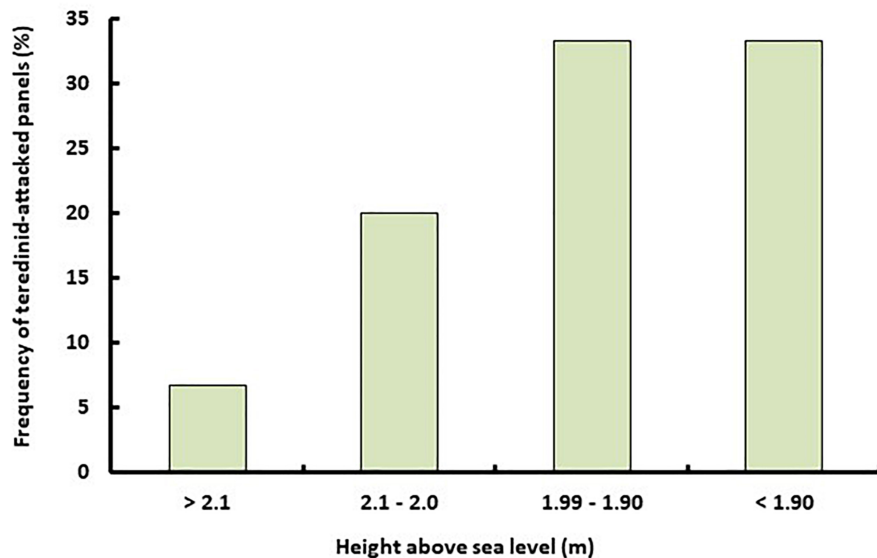
No basidiomycete fruiting bodies were observed in the Laulua mangrove forest. At this locality, in the high to mid intertidal elevations of the mangrove forest the area was dry. Large areas had been harvested, leaving open areas without dense canopy, reducing the humidity. These conditions were not favorable for basidiomycetes (Ananda and Sridhar, 2004).



**FIGURE 5 |** Elevation and wood-degrading guilds within Galua and Laulua mangrove forests. Scale bars for wood degrading guilds represent total abundance of LWD items found containing that guild. Red arrows represent biodegradation boundary between terrestrial and marine wood-degraders. LWD items with no visible degradation were not plotted, and the area between 220 and 600 m in Galua was not sampled. Strandline on x-axis = 0 m. Note, x-axes differ.

Temperature effects are likely to be patchy and affected by canopy gaps, and stochastic by nature. Many LWD items in this area were broken down by beetle larvae and termites. It is likely that beetle larvae and termites are unable to tolerate long periods of inundation. Beetle larvae and termites are therefore important for the structure and function of mangrove forests in those areas that are not environmentally suitable to basidiomycetes. Beetle larvae are responsible for canopy-gap formation, increased litter-fall due to localized tissue death, and nutrient enrichment due to their mechanical boring within

trees (Feller, 2002). Mangrove longhorn beetles, *Aeolesthes holosericeus*, were found in many standing and fallen *Excoecaria agallocha* trees in the high intertidal regions of the Sombano forest. Further toward the mid intertidal regions in the Langira, Galua, and Laulua mangrove forests false blister beetle larvae from the family Oedermeridae were abundant in LWD. These larvae are commonly found world-wide in marine driftwood and pilings (Arnett, 1951). Oedermerids can tolerate tidal immersion (Arnett, 1951), which explains their occurrence in LWD in this study.



**FIGURE 6 |** Teredinid biodegrading activity increases as emersion time decreases. The frequency of teredinid-attacked experimental wooden panels ( $n = 15$ ) placed in areas of different elevations in Galua forest, Kaledupa Island, Indonesia.

Also common were the termite genera, *Cryptotermes* and *Microcerotermes*, found in LWD in the high intertidal areas of the mangrove forests. *Coptotermes*, *Prorhinotermes*, and *Nasutitermes* were found in arboreal nests in the forest canopy of the Sombano, Galua, and Laulua forests. Arboreal termites live in well-defined nests that are separated from their foraging grounds (Korb, 2008). Tree-dwelling termites are long distance foragers (Roisin et al., 2006) that can travel in relative safety because their runways are covered (Merritt and Starr, 2010). Wood-dwelling mangrove termites, however, adopt a different strategy; they live in LWD within the galleries they create (Adams et al., 2007). Wood-dwelling termites, *Cryptotermes* were found frequently within decaying wood in the high intertidal regions of the Galua and Laulua forests. Their galleries extended above and below the high tide mark. *Cryptotermes* colonies remain within a single item of decaying wood where it serves as both food and shelter (Korb and Lenz, 2004). *Cryptotermes* can detect substrate-borne vibrations (Adams et al., 2007; Evans et al., 2007) that may help these termites avoid potential inundation by retreating from areas within their galleries that will be submerged. The well-developed nests also provide environmental buffering from unfavorable outside temperatures (Korb, 2008). Evaporative cooling takes place within LWD, protecting animals within the wood from increasing temperatures outside (Hendy et al., 2013, 2014). A large proportion of the LWD items in the high intertidal areas of the mangrove forests studied were consumed by termites, with estimates of LWD containing up to 70% by volume termite galleries (IWH personal observations).

Other wood-degrading organisms were found in LWD in the mid to low intertidal areas of the mangrove forests, such as limnoriid isopods and teredinids. Limnoriids and teredinids are predominantly aquatic and these animals are only able to tolerate limited tidal emersion in the mangrove forests. However,

limnoriids do not tolerate low salinities, though they were found, albeit infrequently, in the Langira site (Cookson et al., 2012). Limnoriids are more common within Caribbean mangrove forests where salinities remain at suitable levels (Cragg, 2007).

In the mid to low intertidal regions in the Langira, Galua, and Laulua mangrove forests (Figure 5), the breakdown of LWD was dominated by teredinids. Adult teredinids are sessile and disperse by means of planktonic larvae. The availability of LWD is highly variable in mangrove forests (Duke et al., 1981; Kauffman et al., 2020), which influences recruitment of different teredinid species. For example, the teredinid genera dominant in the LWD in this study, *Lyrodus* and *Teredo*, both brood larvae, have reduced free-swimming larval periods and hence more limited dispersal potential, whereas the much less abundant broadcast spawning genus *Spathoteredo* has a greater dispersal potential due to an extended free-swimming larval period. Although *Spathoteredo* has a broad equatorial distribution (Turner, 1966), *Lyrodus* and *Teredo* may have a competitive edge in areas with large volumes and frequent occurrences of LWD. Species with a short free-swimming larval phase tend to be retained in the natal environment (MacIntosh et al., 2012).

Within the low intertidal zones of the forests, LWD was smaller and more fragmented than the LWD within the high intertidal. Teredinid tunneling activity and wood consumption rates increased with decreasing emersion time. Increased teredinid activity along with higher wave energy in the lower intertidal, will increase fragmentation and breakdown of wood.

The rate of deposition and export of particulate and dissolved matter from LWD breakdown will vary with elevation in the mangrove forests. This can be seen in the differences in substrate between the high intertidal regions and the low to mid intertidal regions. The high intertidal regions have less wave energy with greater leaf litter cover and volume of LWD contributing toward



a deeper mud substratum. In contrast, the mid and low intertidal regions have calcareous-sand substrate where wave energy and tidal flow is stronger. Charles et al. (2020) found similar results with a 256% increase in the total below-ground carbon biomass in areas of the mangrove higher up the intertidal regions compared to the areas in mid to low intertidal regions. This has also been noted for temperate and Arctic coastal wetlands, where organic matter content has been found to be higher in the higher elevation areas than lower, particularly where accretion rates have historically exceeded sea level rise (Ward et al., 2014; Ward, 2020).

## Shifting Biodegradation Boundaries in Relation to Sea Level Rise

Wood breakdown and wood-degradation within mangrove forests may be affected by sea level rise, depending on the geomorphology of the mangrove. In Indo-Pacific, some mangrove forests may be able to keep pace with rising sea levels, however, mangroves with a low tidal range may become submerged (Lovelock et al., 2015). Moreover, Saintilan et al. (2020) report that the rate of global warming is increasing. Accelerating rising sea levels may exceed the rate of landward mangrove accretion within 30 years (Saintilan et al., 2020).

This study undertakes the first baseline survey of LWD biodegradation in a mangrove ecosystem, providing a framework for future predictions of biodegradation boundary changes in relation to climate change and sea level rise. In this study, the transition from terrestrial and marine wood-degrading organisms occurs over a distance of 1–2 m on the forest floor, therefore a rise in sea level would shift the distribution of wood degraders from terrestrial to marine guilds where teredinids would become dominant. Further, increasing access of LWD to teredinids will increase rates of LWD biodegradation, resulting in a faster turnover of carbon within mangrove forests. The shift in activity of wood degrading organisms will have significant consequences for mangrove forests where landward accretion is not possible, e.g., due to coral-rock plateaus, as found in the Sombano and Langira sites from this study, or where topographical or infrastructure barriers occur (Ward and Lacerda, 2021). This alteration in the spatial distribution of carbon in mangroves is likely to be regionally variable as a result of species differences in LWD degraders (Cragg, 2007), as well as linked to variations in the impacts of sea level rise globally and linked to tidal variation (Ward et al., 2016; Lacerda et al., 2022). Additionally, as LWD biodegrading guilds play a key role in the structure and function of mangrove forests and adjacent ecosystems, the altered proportion of wood degrading organisms will result in different LWD biodegradation processes, which will affect the sediment deposition and export of POM and DOM from mangrove forests (Robertson and Daniel, 1989).

## CONCLUSION

Tidal height has a large influence on the rate of wood degradation due to the environmental boundary set by inundation frequency and duration, between marine and terrestrial wood degrading consortia, which have markedly different degradation rates. The

high intertidal areas of the forests contained large volumes of soft LWD, slowly broken down by fungi. In the mid to low intertidal areas of the mangrove forests, the volumes of LWD were lower and the wood was less decayed, but was more rapidly broken down due to tunneling by teredinids. This leads to a rapid turnover of LWD in the mid to low intertidal regions, maintaining the important flux of carbon from within intertidal mangrove forests to adjacent ecosystems (Abrantes et al., 2015). Further studies from estuarine sites would enable the effect of salinity to be separated from the effect of inundation regime.

This baseline survey of LWD biodegradation in a tropical mangrove ecosystem, provides a framework for future predictions of biodegradation boundary changes in relation to climate change and sea level rise. The movement of the terrestrial/marine biodegradation boundary due to increasing sea levels will result in a shift in the wood degrading guilds, and in turn affect their carbon processing ability. These changes will impact the function of the mangrove ecosystem and carbon cycles. A full understanding of the factors affecting the biodegradation processes in mangrove forests is critical to accurately assess mangrove carbon stores and the fate of mangrove derived carbon.

## DATA AVAILABILITY STATEMENT

The original contributions presented in the study are included in the article/supplementary material, further inquiries can be directed to the corresponding author.

## AUTHOR CONTRIBUTIONS

IH: conceptualization, methodology, formal analysis, investigation, writing – original draft, writing – review and editing, visualization, and project administration. JRS: writing – formal analysis, review and editing, and project administration. MT: formal analysis, review and editing, and project administration. AE and RW: formal analysis, review and editing. SC: conceptualization, methodology, formal analysis, investigation, writing – original draft, writing – review and editing, visualization, and supervision. All authors: contributed to the article and approved the submitted version.

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