

2023-11-07

Evaluating the Relative Pollen Productivity Estimates using the REVEALS model: a case study from the cultural landscape in Shandong, China

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<https://pearl.plymouth.ac.uk/handle/10026.1/21596>

10.3389/fpls.2023.1240485

Frontiers in Plant Science

Frontiers Media

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1 **Validation of relative pollen productivities in temperate China for reliable pollen-based**
2 **quantitative reconstructions of Holocene plant cover**

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21 *Manuscript to the Journal Frontiers in Plant Science*

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23

24 **Abstract**

25 The Landscape Reconstruction Algorithm (LRA) is regarded as the most rigorous
26 approach for quantifying taxon-specific plant cover from pollen data. The reliability of relative
27 pollen productivity (RPP) estimates is fundamental in the accuracy of quantitative vegetation
28 reconstruction using the LRA approach. Inconsistent RPP estimates produced by different
29 studies can cast doubt on the reliability and applicability of quantitative vegetation
30 reconstruction. Therefore, it is crucial that the RPP estimates are evaluated before being applied
31 for quantitative vegetation reconstruction. We have tested two alternative approaches, namely
32 a leave-one-out cross-validation (LOO) method, and a splitting-by-subregion strategy, using
33 surface pollen assemblages and the REVEALS model – the first step of the LRA – to evaluate
34 the reliability of RPPs estimates of ten target taxa obtained in the cultural landscape of
35 Shandong. We compared the REVEALS estimates (RVs) with observations of regional
36 vegetation abundance (OBVs) and pollen proportions (PPs). The RVs of all taxa are generally
37 closer to OBVs than PPs and the degree of similarity depends strongly on the abundance of
38 individual taxa in plant and pollen, taxa dominant in the region show the highest similarity
39 between RVs and OBVs, such as *Artemisia*, Poaceae, and *Humulus*. The RVs of all herb taxa
40 except *Humulus* and Asteraceae SF *Cichorium*, are slightly over-represented, and the RVs of
41 all tree taxa are under-represented except for *Castanea*. The comparison of RVs with OBVs
42 collected from different spatial extents shows that the RVs of all herb taxa are more similar to
43 OBVs collected from shorter distances (100 km and 75 km for the entire region and the
44 subregion, respectively), while the RVs of all tree taxa are more similar to OBVs collected from
45 longer distances (150 km and 100 km for the entire region and the subregion, respectively).

46 Furthermore, our findings point to the importance of collecting adequate temporal and spatial
47 resolution vegetation survey data from various distances in evaluating the RVs, and that the
48 spatial representation of the REVEALS model is strongly related to the characteristic taxa and
49 size of the deposition basin. We consider the LOO strategy is the better approach for evaluating
50 the RPP estimates from surface moss polsters. We confirm the reliability of the obtained RPP
51 estimates for their potential applications in the quantitative reconstruction of vegetation
52 abundance in temperate China.

53 **Keywords:**

54 China, Relative pollen productivity, Validation, Observation of regional vegetation, REVEALS
55 model

56

57 1. Introduction

58 As an important part of the earth system, land cover plays an important role in the exchange of
59 mass and energy of the climate system. It is therefore of great importance to incorporate land-
60 cover information in climate modeling, and this has been one of the major goals of the PAGES
61 Working Group LandCover6k initiative. Pollen records have been used to infer long-term
62 vegetation changes for a century since von Post (1916) presented the potential of pollen
63 assemblages for representing the surrounding vegetation. However, quantitative
64 reconstructions of past vegetation across space and over time require a better understanding of
65 factors that affect the pollen representation of surrounding vegetation. The two most notable
66 factors that bias the representation of vegetation in pollen assemblages are the differences in
67 the pollen production of different plant species, and the dispersal ability of different pollen
68 types. Taxa with high pollen production and efficient dispersal ability are usually over-
69 represented in pollen assemblages, while taxa with low pollen production and poor dispersal
70 ability are under-represented (e.g. Davis, 1963; Andersen, 1970; Parsons and Prentice, 1981).
71 The correction of the representation bias is approached by modeling the relationships between
72 pollen and surrounding vegetation (Davis, 1963; Andersen, 1970; Parsons and Prentice, 1981;
73 Prentice and Parsons, 1983), and it is fundamental to vegetation reconstruction approaches.
74 Among these approaches, the Extended R-value (ERV) models (Parsons and Prentice, 1981;
75 Prentice and Parsons, 1983; Sugita et al., 1993) are now commonly used to infer taxon-specific
76 pollen-vegetation relationships, and generate relative pollen productivity (RPP) estimates from
77 modern pollen and vegetation data. RPP estimates are necessary for running the Landscape
78 Reconstruction Algorithm (LRA; Sugita, 2007 a, b), which quantifies past plant abundance at

79 regional and local scales, and the Multiple Scenario Approach (MSA; Bunting and Middleton,
80 2009), an approach that can test conceptual spatial arrangements and levels of taxon-specific
81 plant cover in the past.

82 The LRA is a two-step modeling process to estimate past regional vegetation cover (using the
83 REVEALS model: Sugita 2007a) and local vegetation cover (using the LOVE model: Sugita
84 2007b). The LRA incorporates the mechanism and factors within the relationship between
85 modern pollen and vegetation into reconstructions of past vegetation. Furthermore, it
86 reconstructs vegetation from different spatial scales by introducing the size of sediment basins
87 into the model. The LRA has been evaluated using simulation approaches (Sugita 2007a, b)
88 and empirically in North America (Sugita et al., 2010) and many parts of Europe (Hellman et
89 al., 2008a; Cui et al., 2014; Hjelle et al., 2015; Trondman et al., 2016; Marquer et al., 2020).
90 The results show that the LRA works well in predicting the percentage cover of large vegetation
91 units (Hellman et al., 2008a, b; Cui et al., 2014). In Europe, REVEALS-based maps
92 (Pirzamanbein et al., 2014; Trondman et al., 2015; Githumbi et al., 2022) or time series (e.g.,
93 Soepboer et al., 2010; Nielsen et al., 2012) of vegetation composition and landscape openness
94 were shown to be appropriate for climate modeling (Strandberg et al., 2014, 2022) and in
95 evaluating the anthropogenic land-cover change scenarios (Gaillard et al., 2010) and Dynamic
96 vegetation models (Marquer et al., 2017; Dallmeyer et al., 2023). The LOVE model has been
97 used to address biogeographical questions (Cui et al. 2013) and understand local-scale
98 transformation of vegetation by people in archaeological research (Mehl and Hjelle 2016; Fyfe
99 et al., 2018).

100 Fundamental to the quantitative reconstruction of vegetation by using the REVEALS and
101 LOVE models are reliable RPP estimates. Over recent years, a number of studies have been
102 conducted to calculate RPP estimates using the ERV model worldwide, such as in North
103 America (Calcote 1995; Sugita et al., 1998), in Europe (Broström et al., 2004; Räsänen et al.,
104 2007; Mazier et al., 2008), and in China (Li YC et al., 2011; Li YY et al., 2015; Li FR et al.,
105 2017; Wang and Herzschuh, 2011; Wu et al., 2013; Xu et al., 2014; Ge et al., 2015; Zhang et
106 al., 2017, 2021; Jiang et al., 2021, Wan et al., 2022). This effort has resulted in the application
107 of the REVEALS model in studies in northern China (Wang and Herzschuh 2011; Xu et al.,
108 2014; Li FR et al., 2020, 2023). Synthesis of RPP estimates in China and Europe found that the
109 ranking of RPP estimates of major plant types is more or less consistent among studies.
110 However, discrepancies exist between the estimated values from different studies (Bröstrom et
111 al., 2008; Mazier et al., 2012; Bunting et al., 2013; Li FR et al., 2018; Wicczorek and Herzschuh,
112 2020; Serge et al., 2023). RPP estimates validation is not common in China, with only single
113 studies of key biogeographical zones: in temperate China (Xu et al., 2014), subtropical China
114 (Jiang et al., 2020), and tropical China (Wan et al., 2022). Evaluation of the reliability of the
115 available RPP estimates is crucial to have confidence in the accuracy of the REVEALS
116 estimates, as inconsistent estimates of RPP estimates by different studies can cast doubt on the
117 reliability and applicability of quantitative vegetation reconstruction (Hellman et al., 2008a, b;
118 Soepboer et al., 2010; Trondman et al., 2015). Therefore, it is necessary to evaluate the RPP
119 estimates that are being applied in the REVEAL model for quantitative vegetation
120 reconstruction.

121 The reliability of RPP estimates can be evaluated by comparison of REVEALS estimated
122 vegetation on pollen records from large lakes and observed vegetation around the lakes
123 (Hellman et al., 2008a). However, in regions that only have a sparse distribution of large lakes
124 (such as the mountain regions of China) the usage of this validation strategy is hampered. In
125 theory, multiple small sites can be utilized to estimate regional vegetation composition (Sugita,
126 2007a), and this has been further tested with empirical data in southern Sweden (Trondman et
127 al., 2016). The REVEALS estimates from multiple small bogs are comparable with that of
128 relevant large lakes, although standard errors (SEs) of estimates from multiple small sites will
129 be larger than the ones from larger lakes due to between-sites variation (Sugita 2007a;
130 Trondman et al., 2016). This approach has been used to validate the LRA in Denmark (Nielsen
131 et al., 2004) and North America (Sugita et al., 2010). A second approach to evaluating relative
132 pollen productivity estimates is via simulated pollen and vegetation data (Sugita, 2007a).

133 In this study, we aim to: 1) rigorously evaluate the relative pollen productivity estimates
134 originally obtained from the cultural landscape of Shandong (Li FR et al., 2017), with
135 geographically detailed vegetation data that covers a large enough spatial extent, taking
136 advantage of the strength of the ERV and the REVEALS model, and with the possibility of
137 quantifying uncertainty; and 2) compare the two alternative strategies for evaluating these RPP
138 estimates, and propose possible approaches for future validation of RPP estimates.

139

140 2. Material and methods

141 2.1 Geographical setting, site characteristics

142 The study region, located in the lower reach of the Yellow River drainage basin (Figure 1), is
143 one of the most important agricultural regions of China. The potential natural vegetation is
144 deciduous mixed forest according to Wu (1980). At present the region is characterized by
145 cultivated modern fields on the plains, and traditional land use on the terraces in the low
146 mountain areas. The study region encompasses 16 land cover categories grouped into five
147 broader vegetation communities: **i) cultivated crops dominated by** *Triticum* spp. (wheat), *Zea*
148 *mays* (maize), *Arachis hypogaea* (peanut), and *Ipomoea batatas* (sweet potato) **ii) cultivated**
149 **fruit and nut trees consisting of** *Castanea mollissima* (chestnut), *Crataegus pinnatifida*
150 (Chinese hawthorn), *Diospyros kaki* (kaki), *Juglans regia* (walnut), *Malus domestica* (apple),
151 *Pyrus* spp. (pear), *Prunus cerasus* (cherry), and *Zanthoxylum bungeanum* (Chinese pepper). **iii)**
152 **secondary woodland** including mainly *Pinus tabulaeformis* (Chinese red pine), *P. thunbergii*
153 (black pine), *Platycladus orientalis* (Chinese thuja), and *Robinia pseudoacacia* (black locust).
154 **iv) ruderal community characteristic of** abandoned or managed agricultural area boundaries
155 **and v) meadows dominated by** *Artemisia mongolica* (besser), *A. annua* (sagewort), Asteraceae,
156 Caryophyllaceae, *Humulus scandens* (hop), *Lespedeza bicolor* and *L. tomentosa* (bush clovers),
157 and *Vitex negundo* (Chinese chaste tree). More detailed information on the spatial distribution
158 of the vegetation in this region is present in Li FR et al. (2017).

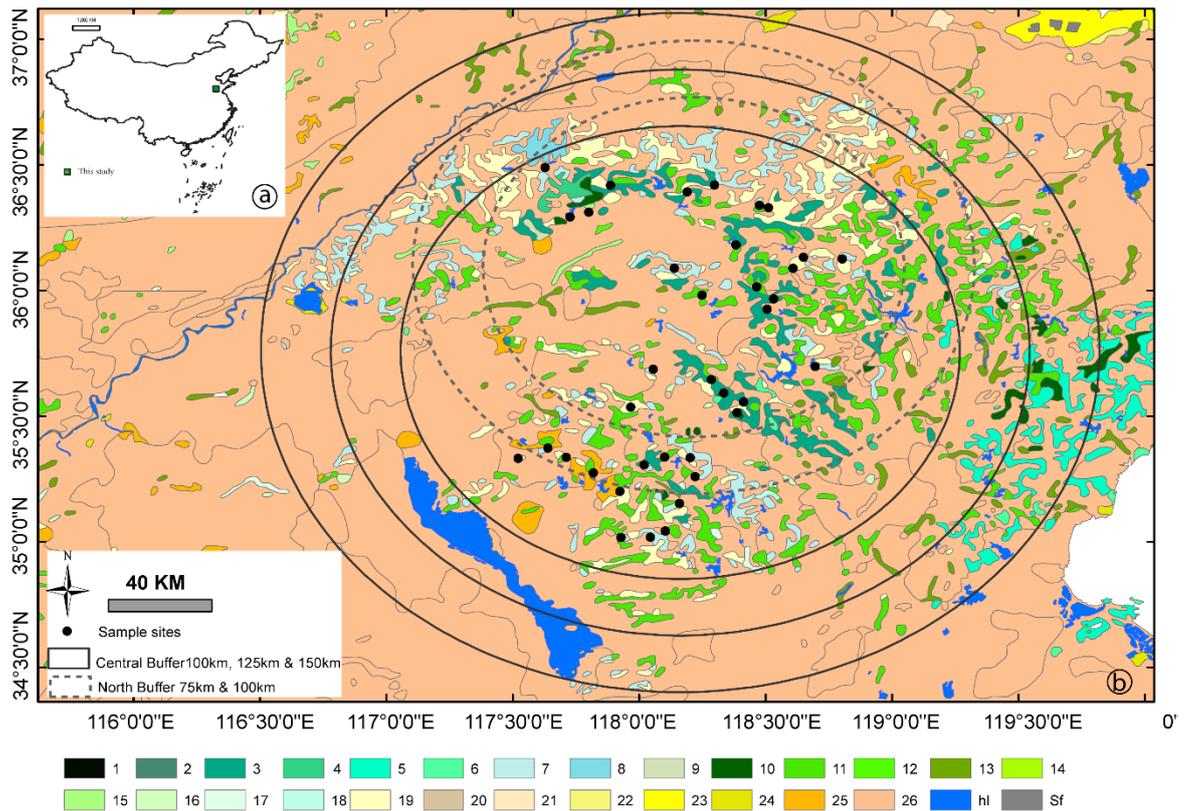


Figure 1. Location of the study area and distribution of the sampling sites with buffers from different distances, solid and dashed circles indicate the distances of vegetation data to be collected, where the dashed lines indicate distance of 75 km and 100 km in strategy I, and the solid lines indicate distance of 100 km, 125 km, and 150 km for strategy II, respectively; Vegetation communities: 1. *Picea jezoensis* forest; 2. *Juniperus komarovii* forest; 3. *Pinus tabuliformis* forest; 4. *Pinus tabuliformis* and *Robinia pseudoacacia* mixed forest; 5. *Pinus densiflora* forest; 6. *Pinus thunbergii* forest; 7. *Platycladus orientalis* forest; 8. *Platycladus orientalis* and *Bothriochloa ischaemum* steppe; 9. *Platycladus orientalis* and *Quercus variabilis* mixed forest; 10. *Quercus acutissima* forest; 11. *Robinia pseudoacacia* forest; 12. *Malus sieversii* forest; 13. *Salix matsudana* forest; 14. *Populus simonii* forest; 15. *Populus nigra* forest; 16. *Populus*, *Salix*, *Ulmus* mixed forest; 17. *Cotinus coggygria* shrub; 18. *Carex moorcroftii* steppe; 19. *Vitex negundo*, *Ziziphus jujuba*, *Bothriochloa ischaemum* shrub-steppe; 20. *Zoysia japonica* dominated meadow; 21. *Secale sylvestre* and *Leymus paboanus* dominated meadow with *Betula pendula*;

172 22. *Aeluropus pungens* dominated meadow; 23. *Suaeda glauca* dominated meadow; 24. *Phragmites*
173 *australis* Marsh; 25. Cultivated crops: *Triticum aestivum*, *Oryza sativa*, *Setaria italica*, *Ipomoea batatas*,
174 *Sorghum bicolor*, *Gossypium spp*, *Arachis hypogaea*, *Zea mays*; 26. Cultivated crops with cultivated
175 trees (*Juglans regia*, *Castanea mollissima*, *Crataegus pinnatifida*, *Malus domestica*, *Amygdalus persica*,
176 *Prunus armeniaca*, *Vitis vinifera*, *Ziziphus jujuba var. jujuba*); hl. Waterbody; Sf. Saltern.

177

178 2.2 Modern pollen data collection

179

180 Thirty-six surface moss polsters were collected randomly within an area of a 100 km radius in
181 May 2013 and 2014, among which, 17 and 19 sites are located within an area of about a 50 km
182 radius in the northeast, and in the southwest, respectively. The sample collection and pollen
183 extraction strategy was described by [Li FR et al. \(2017\)](#). Pollen identification was performed
184 under a light microscope at 400× magnification, and in special cases at 600× magnification, by
185 using the published Pollen flora of China ([Wanget al., 1995](#)) with a complimentary of European
186 pollen flora ([Beug, 2004](#)) and reference collections of physical specimens in the Paleoecology
187 lab of Linnaeus University. Pollen counts were limited to no less than 1000 grains produced by
188 terrestrial flowering plants for each sample.

189

190 2.3 Vegetation data surrounding the sampling sites

191 For evaluation of the RPP estimates obtained by [Li FR et al. \(2017\)](#), we collected the observed
192 local- and regional-scale plant cover for the region. We operationally define the local and

193 regional scale as within 1500 m from each site and up to 150 km from the geographical center
194 of all sites.

195 **2.3.1 Observed local vegetation survey within 1500 m of the sampling sites**

196 Vegetation surrounding the 36 moss polster sites was collected up to 1500 m for each site
197 following the standard protocol suggested by [Bunting et al. \(2013\)](#) during the same period as
198 pollen sample collecting. Where the vegetation collection strategy varies depending on the
199 distance from the sampling sites (Zone A, B, and C), and is briefly summarized as follows. 1)
200 Zone A (0–10 m): plant cover was estimated within 21 1 m² quadrats, one quadrat was located
201 at the center of the sampling site, and 20 other quadrats were placed at specific distances of N,
202 E, S, W, NE, SE, SW and NW directions. 2) Zone B (at 10–100 m): vegetation communities,
203 defined based on dominant taxa, were mapped using a compass and a hand-held GPS out to
204 100 m, vegetation was surveyed in each community within several randomly distributed 1 m²
205 quadrats for the open community while 6 m radius point surveys and 1 m² quadrats for semi-
206 open and forest community. 3) Zone C (at 100–1500 m): Google Earth image within 1500 m
207 from the center of each moss sample was used to identify land cover categories, where both
208 winter and summer images were adopted to separate deciduous from evergreen trees; plant
209 composition for each land-cover category was obtained from the field surveys. For a more
210 detailed vegetation survey strategy, the readers are referred to [Li FR et al. \(2017\)](#).

211 **2.3.2 Observed regional plant-cover**

212 To compare the REVEALS estimates with observed vegetation from different distances, we
213 obtained the plant cover within a radius of 100 km, 125 km, and 150 km respectively from the

214 geographical center of the 36 sampling sites. Meanwhile, the mean plant cover for the two
215 subregions was obtained within a circle of 75 km and 100 km radius from the geographical
216 center of the 17 sites and 19 sites respectively (figure 1). The distances were selected based on
217 previous empirical and theoretical studies, which show that REVEALS estimates represent
218 vegetation within an area of ca. 10^4 – 10^5 km² (e.g. [Sugita, 2007a](#); [Helman et al., 2008a](#);
219 [Soepboer et al., 2007](#); [Wan et al., 2022](#)).

220 The extracted regional plant abundance was compiled from three data sources. 1) Land cover
221 types were primarily extracted from a 1:1 million vegetation map of China ([Hou, 2019](#)), this
222 data source provides 542 Chinese land cover categories, of which 23 were included within the
223 designed areal extent of this study. Plant composition for each categorized land cover type was
224 obtained from 2) a field survey as in [Li et al. \(2017\)](#), and 3) literature ([Wang and Zhou,](#)
225 [2000](#)). The field survey results were generalized by calculating the mean species abundance of
226 all representative quadrats of the same categorized land cover type. For land cover categories
227 whose plant composition is not available from the field survey, the information was obtained
228 from the literature ([Wang and Zhou, 2000](#)). When the species composition in the literature is
229 described as Braun-Blanquet cover-abundance scale ([Braun-Blanquet, 1964](#)), the abundances
230 were translated to cover estimates within each vegetation type as follows: present-0.01%; I-5%;
231 II-10%; III-25%; IV-50%; V-75%.

232 The plant species recorded in the field survey and collected from the literature were harmonized
233 with pollen morphological taxonomy and nomenclature to overcome the variation in taxonomic
234 resolution of pollen identification, following two publications ([Wang et al., 1995](#); [Beug, 2004](#)).
235 The most common ten pollen morphological types that were present in at least 60% of the sites

236 and with a maximum percentage value larger than 1% were selected for this study. The plant
 237 composition (aggregated to pollen morphological taxonomy) of each categorized community
 238 is presented in table 1.

239 Table 1. List of vegetation categories (the same as in Figure 1) and the plant composition (aggregated to
 240 pollen morphological taxonomy) of each categorized community that are included within the buffer
 241 distance of 150 km. Plant composition for vegetation categories is collected from a field survey (Li FR
 242 et al., 2017; marked with¹) and published literature (Wang and Zhou, 2000; denoted with²). Plant species
 243 harmonized to pollen-types full names: Art *Artemisia*, Caryo Caryophyllaceae, Cas *Castanea*, Che
 244 Amaranth./Chenop., Cich Asteraceae SF. Cichorioideae, Hum Cannabis/Humulus, Pin *Pinus*, Poa
 245 Poaceae, Que *Quercus*, Ulm *Ulmus*.

Vege	land cover type	Art	Car	Cas	Che	Cic	Hu	Pin	Poa	Que	Ulm
3	<i>Pinus tabulaeformis</i>	4.82	1.69	0	0	1.74	0	74.5	20.1	2.2	6.25
5	<i>Pinus densiflora</i> forest ¹	4.82	1.69	0	0	1.74	0	74.5	20.1	2.2	6.25
6	<i>Pinus thunbergii</i> forest ¹	4.82	1.69	0	0	1.74	0	74.5	20.1	2.2	6.25
7	<i>Platycladus orientalis</i>	0.89	0	0	1.78	0.63	2.08	0	11.5	0	0.76
8	<i>Quercus acutissima</i>	3	3.01	0	0.04	0.1	0.4	6.72	16	73.9	0
9	<i>Robinia pseudoacacia</i>	1.74	0.8	0	2.48	1.3	0.3	1.82	26.3	0	0.08
10	<i>Malus sieversii</i> forest ²	0	0	0	0	0	0	0	21	0	0
11	<i>Salix matsudana</i> forest ²	0	0	0	0	0.1	0	0	103	0	0
14	<i>Populus, Salix, Ulmus</i>	0	0	0	0	0.1	0	0	103	0	36
15	<i>Cotinus coggygria</i> ²	2	0	0	0	0	0	0	4	0	0
16	<i>Carex moorcroftii</i> ²	2.35	4.71	0	0	0	0	0	7.1	0	0

17	<i>Vitex negundo</i> var.	1.8	4.41	0.69	0.17	0.63	0.15	0.04	20.1	0.07	1.85
22	<i>Phragmites communis</i>	0	0	0	0	0	0	0	92	0	0
23	Cultivated field ¹	0.35	0.05	0.97	0.6	1.86	0.15	0	1.05	0	0.01
24	Cultivated tree ¹	0.83	0.43	9.70	1.90	2.49	0.63	0.00	9.64	0	0.01

246

247 **2.4 RPP estimates validation using the REVEALS model**

248 To evaluate the relative pollen productivity (RPP) estimates from Shandong (Li FR et al., 2017),
249 we use pollen counts from 36 moss polsters, observed local vegetation, and regional vegetation.
250 When sub-fossil pollen data are available only from small-sized sites, such as within this region,
251 REVEALS results have been shown to be robust when pollen data from several small-sized
252 lakes and bogs are used, both theoretically (Sugita, 2007a) and empirically (Sugita et al., 2010;
253 Fyfe et al., 2013; Trondman et al., 2016). In attempting to maximize the effectiveness of our
254 validation, we have used two independent strategies to avoid circularity in the validation
255 process for obtaining RPP estimates (we rerun the ERV model based on 10 target taxa of this
256 study instead of the exact values based on 18 taxa published in Li FR et al. (2017)) and the
257 REVEALS estimates. Figure 2 summarizes the testing schemes.

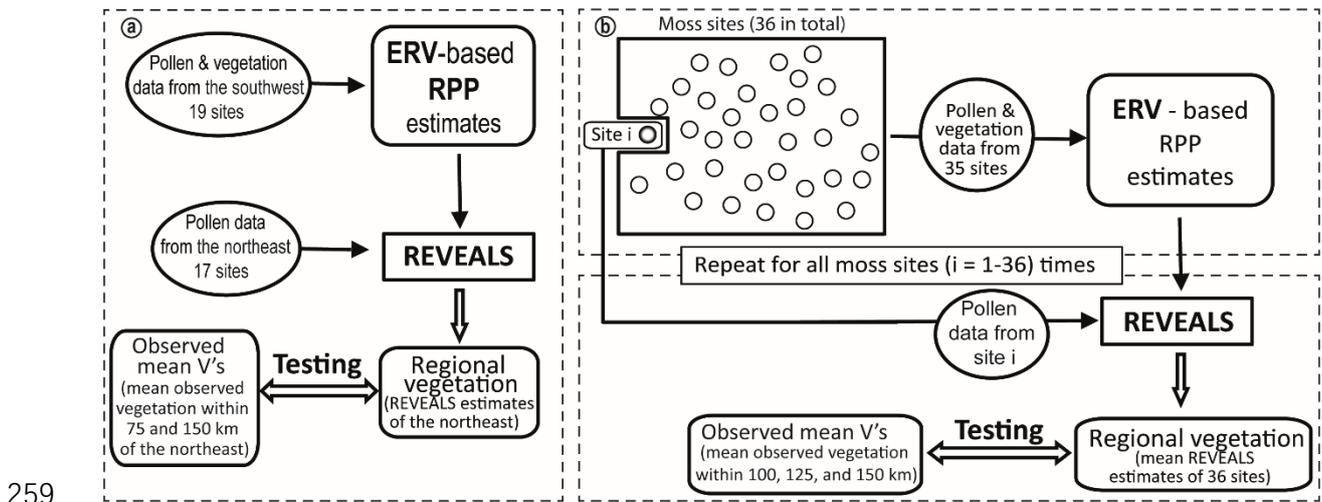


Figure 2. Flowchart of strategies used to validate relative pollen productivity (RPP) using the

REVEALS model: the splitting-by-subregion strategy (a) and the leave-one-out (LOO) strategy (b).

In strategy I, we handle the pollen and vegetation data from the northeast and the southwest sub-groups separately. We run the ERV model with pollen counts and vegetation data within 1.5 km of the 19 sites located in the southwest and the 17 sites located in the northeast of the study region to generate a new set of RPP estimates. The standard deviations (SDs) for all RPP estimates are lower than RPP values from the southwest, while the SDs of several taxa are larger than the RPP values from the northeast (results not shown), which motivated us to apply the REVEALS model on pollen counts of the 17 sites from the northeast and the RPP estimates obtained from the southwest. The 17 northeast moss polster samples are grouped within an area of about 50 km radius, we therefore compared the REVEALS estimates from the northeast with the observed regional vegetation from two different distances (75 km and 100 km) from the geographical center of the northeast to test the spatial scale of the REVEALS model. We refer to original pollen proportions as PPs, the REVEALS estimates of regional vegetation as

274 RVs_NE, and the observed vegetation as OBVs_NE75, and OBVs_NE100, for distances of 75
275 km and 100 km respectively.

276 In strategy II, we use the “leave-one-out” (LOO) cross-validation scheme (Efron and Tibshirani,
277 1998; Sugita et al., 2010; Mazier et al., 2015). We remove one site from the full dataset of 36
278 locations and apply the ERV model using pollen and vegetation data from the remaining 35
279 sites to obtain RPP estimates. We then apply the REVEALS model using the new RPP values
280 and pollen counts of the selected site that has been removed from the dataset (Figure 2). The
281 process is repeated 36 times, and mean REVEALS estimates and related SEs from the 36 runs
282 are estimated, where the SEs of the mean REVEALS estimates were calculated based on the
283 delta method (Stuart and Ord, 1994). We refer to the mean REVEALS estimates of the regional
284 vegetation obtained LOO as RVs_LOO, while the observed regional vegetation as
285 OBVs_LOO100, OBVs_LOO125, and OBVs_LOO150 for distances of 100 km, 125 km and
286 150 km respectively.

287 For all ERV model runs Poaceae is chosen as the reference taxon for which the RPP is set to 1,
288 and RPP of all other taxa is expressed relative to Poaceae. The fall speed of pollen (FSP),
289 derived from Li FR et al. (2017), was originally calculated from Stoke’s law (Gregory, 1973)
290 with measurements of the diameter on the short- and long-axes of the pollen grains. The taxon-
291 specific approach of Prentice (Prentice, 1985) was selected to calculate distance-weighted plant
292 abundance at a distance from the edge of the depositional basin (0.5 m radius), based on
293 Sutton’s atmospheric diffusion model (Gaussian plume diffusion model, GPM) for the
294 deposition of small particles in the air (Sutton 1953). ERV (Extended R-value) sub-model 3
295 (Sugita, 1994) was applied for the estimation of the RPP of each taxon with a maximum

296 likelihood method (Parsons & Prentice, 1981; Prentice and Parsons, 1983; Sugita, 1994); the
297 moving-window linear regression method (Gaillard et al., 2008) was used for estimating the
298 relevant source area of pollen (RSAP) simultaneously. For a detailed rationale for parameter
299 setting in the model run, the readers are referred to Li FR et al. (2017).

300 Except for the two strategies shown and mentioned above, we attempted randomly selecting
301 half number of sites to run the ERV model, and the other half to run the REVEALS model, and
302 only two among ten trials showed promising results (results not shown). Others either have
303 large SDs of RPP estimates from the ERV model or have too large SEs of the REVEALS
304 estimates, hence they are not shown here. Moreover, we have attempted the strategies adopted
305 by Sugita et al. (2010), that is using randomly selected half-number of sites to run the ERV
306 model to get RPP estimates and applying LRA with the LOO cross-validation strategy on the
307 other half of sites. However, the SEs of the REVEALS estimates or LOVE estimates are too
308 larger (results not shown).

309

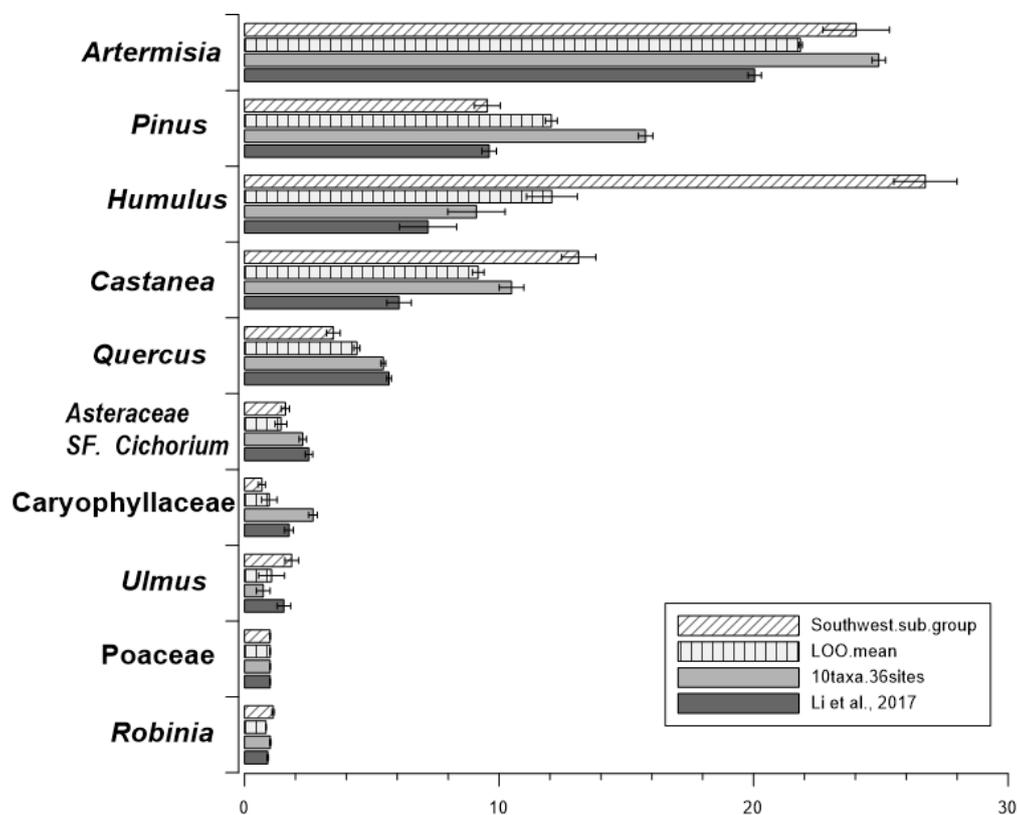
310 **3. Results**

311 These results describe the differences in the RPP estimates and the validation efficiency from
312 the two strategies separately in comparison with that of the entire region..

313 **3.1 Results from the strategy I**

314 **3.1.1 Relevant source area of pollen and relative pollen productivity estimates**

315 Log-likelihood against distance curves (ESM Figure S1) in ERV models indicate the change in
 316 goodness-of-fit of the data to the model, and log-likelihood (LLD) increase with distance. The
 317 point at which the curve reaches an asymptote is defined as the relevant source area of pollen
 318 (RSAP), measured in meters (Sugita, 1994). The curve of LLD approached the asymptote at a
 319 distance of 154 m when the southwest sites were used for running the ERV model, while the
 320 RSAP is 145 m both in the entire region and when all sites were used for calculating RPP
 321 estimates of 18 taxa (Li FR et al., 2017).



322

323 Figure 3. Bar-plot of the relative pollen productivity (RPP) estimates and their standard deviations
 324 (SDs) of the ten target taxa from the southwest, mean RPP estimates and related standard error of the
 325 LOO strategy, and the RPP estimates and SDs recalculated for the ten taxa for the entire region and the
 326 RPP estimates from Li FR et al. (2017).

327 The RPP estimates from strategy I and the mean RPP values from strategy II are presented in
328 Table 2 and Figure 3, alongside recalculated RPP based on ten target taxa from the entire region
329 and earlier published RPP values (Li FR et al., 2017). The SDs for RPP estimates from the
330 southwest and the entire region and the SE of mean RPPs from LOO are lower than RPP values.
331 The ranking of the RPP estimates from the southwest is *Humulus* > *Artemisia* > *Castanea* >
332 *Pinus* > *Quercus* > *Ulmus* > Asteraceae SF. *Cichorium* > *Robina* > Poaceae > Caryophyllaceae.
333 While the ranking is *Artemisia* > *Pinus* > *Humulus* > *Castanea* > *Quercus* > Asteraceae SF.
334 *Cichorium* > Caryophyllaceae > *Ulmus* > Poaceae > *Robina* for the entire region.

335 Table 2. The relative pollen productivity (RPP) estimates from strategy I (the southwest subregion), the mean RPP
336 estimates of strategy II (LOO) strategy, and recalculated RPP based on ten target taxa from the entire region and
337 earlier published RPP values (Li FR et al., 2017).

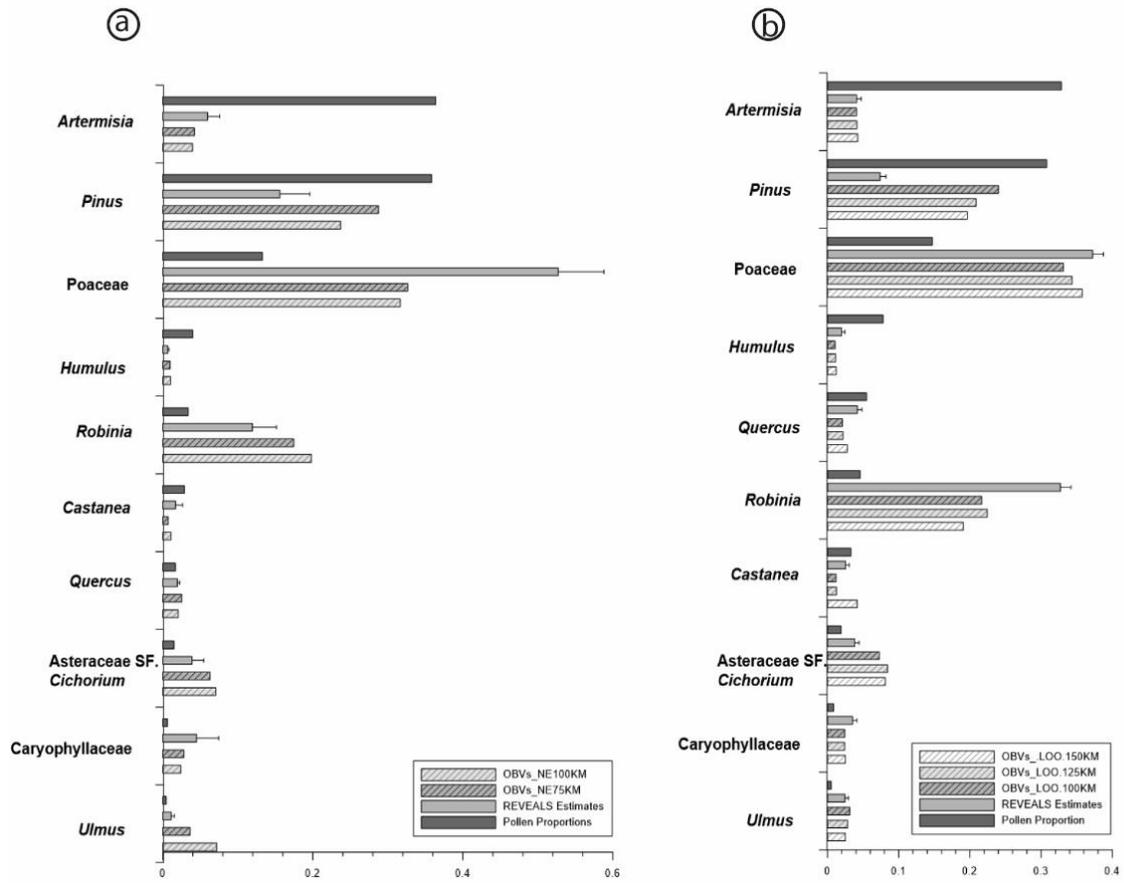
Taxa	Strategy I		Strategy II		Entire region		Li FR et al., 2017	
	(sub-region)	SD	(LOO)	SE	(10taxa, 36sites)	SD	(18taxa, 36 sites)	SD
<i>Artemisia</i>	24.03	0.07	21.84	1.31	24.92	0.26	20.04	0.15
<i>Pinus</i>	9.54	0.24	12.05	0.51	15.76	0.29	9.61	0.23
<i>Humulus</i>	26.75	1.00	12.08	1.24	9.12	1.12	7.21	0.90
<i>Castanea</i>	13.13	0.23	9.18	0.67	10.50	0.48	6.07	0.34
<i>Quercus</i>	3.49	0.12	4.41	0.27	5.45	0.09	5.67	0.17

Asteraceae SF.	1.60	0.23			2.28	0.15	2.53	0.11
<i>Cichorium</i>			1.43	0.16				
Caryophyllaceae	0.68	0.31	0.97	0.14	2.69	0.17	1.74	0.16
<i>Ulmus</i>	1.85	0.51	1.06	0.27	0.73	0.27	1.55	0.37
Poaceae	1	0	1	0	1	0	1	0
<i>Robinia</i>	1.12	0.01	0.84	0.05	1.01	0.03	0.91	0.03

338

339 3.1.2 REVEALS estimates from the strategy I

340 To visualize the accuracy of the estimated vegetation using the REVEALS model, the RV_NE
341 and PPs are shown in comparison with OBVs_NE75 and OBVs_NE100 for each of the
342 individual taxa (ten target taxa) from the strategy I (Figure 4a). Poaceae, *Robinia*, Asteraceae
343 SF. *Cichorium*, Caryophyllaceae, and *Ulmus* are under-represented in PPs in comparison to
344 OBVs. *Artemisia*, *Pinus*, *Humulus*, and *Castanea* are overestimated by PPs. The SEs for all
345 RVs are lower than RV values, and the RVs of all taxa are generally closer to OBVs than PPs
346 alone. RVs of all herb taxa, except *Humulus* and Asteraceae SF *Cichorium*, are slightly over-
347 represented, and RVs of all tree taxa are under-represented. The comparison of RVs with OBVs
348 collected at different spatial scales shows that the RVs of all herb taxa are closer to
349 OBVs_NE75, while the RVs of all tree taxa (*Pinus*, *Robinia*, *Quercus*, *Castanea*, and *Ulmus*)
350 are more similar to OBVs_NE100.



351

352 Figure 4: Bar Charts of REVEALS-based vegetation abundance (RVs) in comparison with that of
 353 pollen proportions (PP), and observed vegetation (OBVs) within different distances from strategy I (a)
 354 and strategy II (b).

355

356 **3.2 Results from strategy II**

357 **3.2.1 RPP estimates from LOO**

358 Log-likelihood values increase as the distance increased, and the curves approached the
 359 asymptote at the distance between 150 and 171 m for all 36 ERV runs in the LOO strategy

360 (ESM Figure S1), while the RSAP is 145 m based on the ERV model of the entire region and
361 Li FR et al. (2017).

362 The SDs for all RPP estimates from the 36 runs in strategy II are lower than RPP values, and
363 the rank order of mean RPP estimates of the 36 ERV runs in the LOO strategy and the entire
364 region is similar, which is *Artemisia*>*Humulus*>*Pinus*>*Castanea*>*Quercus*> Asteraceae SF.
365 *Cichorium* > *Ulmus* > Poaceae> Caryophyllaceae> *Robinia*, except that *Pinus* has a higher RPP
366 than *Humulus*, and Caryophyllaceae has a higher RPP than Poaceae and *Robinia* in the ERV
367 runs of the entire region.

368 **3.2.2 REVEALS estimates from strategy II**

369 To visualize the accuracy of the estimated vegetation using the REVEALS model, Figure 4b
370 depicts the RVs and PPs, in comparison with OBVs from 100 km, 125 km, and 150 km for
371 individual taxa from strategy II. Similar trends are displayed to those in strategy I: Poaceae,
372 *Robinia*, Asteraceae SF. *Cichorium*, and Caryophyllaceae are under-represented in PP, and
373 *Artemisia*, *Pinus*, and *Quercus* are overestimated. SEs of RVs for all taxa are smaller than the
374 RV values, and the REVEALS model improves the accuracy of estimated vegetation than using
375 PP alone for all taxa except for *Castanea*. RV underestimates cover of *Pinus*, *Robinia*, and
376 Asteraceae SF. *Cichorium*, and slightly overestimates cover of *Artemisia*, Poaceae,
377 Caryophyllaceae, and *Ulmus*. A comparison of the RVs with OBVs from different spatial scale
378 show similar trends to strategy I. RVs are more similar to OBVs_LOO100 for *Artemisia* and
379 Asteraceae SF. *Cichorium*, while RVs are more similar to OBVs_LOO125 for *Ulmus*, and
380 closer to OBVs_LOO150 for *Pinus*, Poaceae, *Quercus*, and *Robinia*. There is no discernible

381 difference between OBVs_LOO100, OBVs_LOO125, and OBVs_LOO150 km for *Humulus*
382 and Caryophyllaceae.

383

384 4. Discussion

385 4.1 Relative Pollen Productivities

386 The stability/reliability of RPP estimates of the ERV runs is firstly evaluated by the shape of
387 log-likelihood curve against distance. In all tests the curve increased and reached an asymptote,
388 which ensures the goodness of fit of the data to the model. However, the shape of the curve
389 from strategy I (dividing the datasets by region) shows a slight departure from the theoretical
390 behaviour, which should be stable after having reached the asymptote. The ERV results from
391 the LOO strategy and the whole data set are similar in terms of the changes in both the log-
392 likelihood and the likelihood function score against distance (ESM Figure S1).

393 The stability/reliability of RPP estimates is highly correlated with their standard deviations
394 (SDs) calculated in the ERV model. In this study, SDs of all RPP estimates from both strategies
395 are lower than the RPP values. The ranking of the RPP estimates does not show significant
396 differences between the southwest (strategy I), the LOO strategy (strategy II), and the entire
397 region. The variation in RPP estimates between the three datasets is within the range of 30%
398 for most of the taxa, except *Humulus*, Caryophyllaceae, and *Ulmus*, where the variation is up
399 to 50% between estimates. Greater differences in RPP estimates are found between the
400 southwest and the entire region than between the LOO strategy and the entire region.

401 Several factors may account for the dissimilarities between the RPP estimates. The noticeable
402 disparity can be seen in the vegetation openness and heterogeneity caused by differences in
403 human disturbance intensity from the two subregions. There are more sites from the southwest
404 sub-region that are either located within cultivated fields or include cultivated fields within 100
405 m compared to the northeast. Nevertheless, there is little direct evidence to believe that
406 differences in RPP estimates from the southwest and the entire region are due to human activity,
407 although previous studies have found that pollen concentration of non-cultivated plants is much
408 lower in cultivated fields than in the natural and abandoned field communities (Li MY et al.,
409 2012). Therefore, it would be worth testing the impact of human activity on RPP estimates in
410 the future.

411

412 **4.2. Validation**

413 **4.2.1 Evaluation of the validation strategy**

414 There are various different validation strategies for RPP estimates (Soepboer et al., 2007; Sugita
415 et al., 2010; Jiang et al., 2020; Wan et al., 2022). In this work, we have evaluated RPPs in terms
416 of the stability of RPP estimates, and similarities between observed and estimated regional
417 vegetation based on the three sets of RPP estimates generated here.

418 We first compare the RPP estimates from the two strategies (LOO and regional splitting) with
419 those based on ten taxa of the entire region. The RPP estimates of all taxa from the LOO strategy,
420 except for *Artemisia* and *Robinia*, are more similar to those of the entire region than those from

421 the regional splitting approach (Table 2). The LOO strategy is therefore better than the
422 subregion strategy if we measure the efficiency in terms of the similarity of the RPP estimates.

423 The validation efficiency, quality, and reliability of the REVEALS estimate are shown by the
424 SEs of the RVs. There are two sources of uncertainty in the RVs: the SDs of RPP values, and
425 the between multiple sites variation in pollen assemblages. Both are incorporated into the SEs
426 of the RVs. Our results show that the SEs of RVs for all taxa are lower than RV values. The
427 coefficients of variation between observed and estimated vegetation abundance (expressed as
428 standard variation divided by the OBV values) vary among taxa. The largest variation detected
429 between RVs and OBVs is less than 1.4-fold, whereas it is as large as 8.19-fold between PPs
430 and OBVs in strategy I. Similar trends are found in strategy II, where the greatest difference
431 between RVs and OBVs is less than 1.1-fold, whereas it is 7.1-fold between PPs and OBVs. In
432 general, the RV of all taxa, except *Pinus* in both strategies, show better results than PP alone,
433 which confirms that the application of RPP estimates in the REVEALS model provides
434 consistently better and more reliable estimates of vegetation cover. Overall, we confirm the
435 reliability of the RPP estimates, and we consider the LOO strategy to be a better choice for
436 evaluating the RPP estimates using the REVEALS model.

437 This study has shown that herb taxa, and in particular *Artemisia*, Caryophyllaceae, and Poaceae,
438 are overrepresented by the RVs. Some herbs match well (Asteraceae S.F. *Cichorium* and
439 *Humulus*). Tree taxa (*Pinus*, *Quercus*, *Robinia*, and *Ulmus*), although not *Castanea*, are
440 underrepresented by the RVs. This is because pollen records from small-sized sites capture
441 more pollen grains from plants in close proximity than those from farther away ([Prentice, 1985](#);
442 [Sugita, 1994](#)), and additional trees growing at a longer distance are underestimated in the mean

443 regional estimated vegetation by the REVEALS model. Similar trends are detected in the
444 mountain region of the northern Pyrenees (Marquer et al., 2020). In addition, the
445 overrepresentation of herbs and underrepresentation of trees by the REVEALS model shows
446 larger dissimilarity for aggregated taxa (herbs vs trees) than individual taxon (e.g. Marquer et
447 al., 2014).

448 Pollen assemblages represent the surrounding plant abundance well in terms of individual taxon
449 although the degree of similarity between observed and simulated vegetation abundance varies
450 among taxa. The highest similarity between OBVs and RVs is found in taxa with high
451 abundances in both pollen and vegetation, e.g. Poaceae, *Artemisia*, and Caryophyllaceae. The
452 discrepancies between RVs and OBVs are much higher for taxa with a low abundance of pollen
453 or vegetation, such as *Humulus*. In general, the RVs improve the accuracy of quantitative
454 reconstruction of vegetation around the sampling sites in comparison to PPs, which supports
455 earlier validation studies in China (e.g. Wan et al., 2022) and Europe (e.g. Hellman et al., 2008a,
456 b), although further tests using pollen records from larger lakes are necessary.

457 We attempted a randomized splitting of the sites to validate our RPPs, but obtained high SDs
458 in our RPP estimates and large SEs in the REVEALS results, most likely owing to the
459 heterogeneity in the study region of local vegetation. We therefore only used dataset splitting
460 by region.

461

462 **4.2.2 Spatial representation of the REVEALS model**

463 Our study revisits one of the most interlinked issues in the quantitative reconstruction of
464 regional vegetation, the spatial representation of the REVEALS model. A number of theoretical
465 (Soepboer, 2007; Hellman et al., 2008b) or empirical studies have been conducted in Europe
466 (Hjelle et al., 2015; Marquer et al., 2020) and China (Wan et al., 2022) to address the spatial
467 representation of the REVEALS model to vegetation, which varies between 50–200 km distance
468 from the sampling basin in prior studies. The RVs based on pollen records from small to middle-
469 sized lakes in the Swiss plateau are comparable with vegetation collected around 200 km
470 (Soepboer, 2007). The spatial extent of the RVs of pollen records from large lakes is about
471 100*100 km² (Ca. 100 km radius) in simulation tests in southern Sweden (Hellman et al.,
472 2008b). While RVs from small water reservoirs in tropical China show the highest similarity
473 with vegetation collected within 80 km, although there are no significant differences between
474 observed vegetation collected within 50 km and 80 km (Wan et al., 2022).

475 The spatial size of the vegetation dataset required for validation of REVEALS estimates of
476 plant cover and/or RPPs is an issue that was first evaluated and discussed by Hellman et al.
477 (2008a) in a study performed in southern Sweden. An area of 100 km x 100 km was shown to
478 provide the best fit between the vegetation dataset (actual vegetation/plant cover) and the plant
479 cover predicted by REVEALS. No comparable study has been performed since in other regions.
480 It has been assumed that the spatial scale of REVEALS reconstructions is of similar magnitude
481 independantly of the characteristics of the regional vegetation (e.g. Trondman et al., 2015;
482 Githumbi et al., 2022). In theory and practice, however, the strict definition of the pollen source
483 area is difficult for REVEALS application. Sugita (2007a) defined it as the area within which
484 most of the pollen comes from. Simulations and previous empirical studies (e.g. Sugita, 2007a,

485 b; Hellman et al., 2008a; Sugita et al., 2010; Mazier et al., 2012) have indicated that, when the
486 radius of the source area defined varies from 50 km to 400 km, the REVEALS results of
487 regional vegetation reconstruction do not change significantly. The usage of multiple small
488 moss polster sites for estimating regional vegetation is the most notable difference between this
489 study and previous work. The high homogeneity of vegetation collected from the vegetation
490 map in the sampling region induced high similarity in the mean observed regional vegetation
491 between different distances although the actual landscape is much patchy. Nevertheless, our
492 results demonstrate that herb covers estimated by REVEALS are more similar to the mean
493 observed vegetation within 75 km, whereas RVs of tree taxa, except *Robinia*, are closer to the
494 mean vegetation within 100 km in strategy I. Notably, the similarity between RVs and OBVs
495 collected from different distances varies from individual taxa in strategy II, RVs of *Artemisia*
496 and Asteraceae SF. *Cichorium* are much more similar to OBVs within 100 km, while RVs of
497 *Humulus* and *Ulmus* are more similar to OBVs within 125 km, and RVs of *Pinus*, Poaceae,
498 *Quercus*, and *Robinia* are closer to OBVs within 150 km. Overall, the RVs of herb taxa are
499 more similar to OBVs collected within shorter distances, while the RVs of trees are more
500 similar to OBVs collected within longer distances in both strategies. Our findings point to the
501 importance of collecting vegetation from different distances in evaluating the REVEALS
502 estimated regional vegetation, and that the spatial extent of the REVEALS model is related to
503 the character of the taxa and the size of the “deposition basin”.

504

505 **4.2.3 Discrepancies between RVs and OBVs**

506 The REVEALS model results in much better estimates of vegetation cover than unconverted
507 pollen percentages; however, discrepancies still exist between estimated and observed
508 vegetation. This is most clear for taxa that are dominant in some sites but absent in the others,
509 such as *Robinia* and *Castanea*, whose RPP values obtained in the testing process may not
510 sufficiently present the average/true values from this region. The RVs slightly overestimate
511 most of the herb taxa and under-estimate all tree taxa in strategy I, and RVs underestimate
512 *Pinus*, *Robinia*, and Asteraceae SF. *Cichorium*, while overestimate Poaceae, Caryophyllaceae,
513 and *Ulmus* in strategy II.

514 Sugita (2007a) suggests that the number and properties of sites that are necessary for
515 REVEALS to reconstruct representative regional vegetation depend on the regional spatial
516 complexity, gradients, species composition and diversity, and basin size. A variety of factors
517 could account for the discrepancies between observed and estimated vegetation from the
518 REVEALS model, and we consider the most important ones to be: (1) pollen assemblages; (2)
519 between-sites variation in the vegetation of multiple sites; (3) other factors related to
520 methodology.

521 *Pollen assemblages*

522 The temporal differences in the pollen sample and the collected vegetation might be responsible
523 for some of the discrepancies between the estimated and observed vegetation. Moss polsters
524 represent generally 1–2 (3) years of pollen deposition according to studies in Europe (e.g.,
525 Cundill, 1991; Räsänen et al., 2004). Given the small green fresh moss samples we collected in
526 Shandong, we believe the moss polsters represent vegetation from no more than two years.

527 However, the time span of the obtained vegetation abundance varies, where the local vegetation
528 applied in the ERV model is based on one year, whereas the regional vegetation is collected
529 from the available vegetation inventories conducted in the 1980s (Hou et al., 2019).
530 Nevertheless, there is little evidence to believe that the discrepancies between observed and
531 estimated vegetation are due to this factor, and the application of regional vegetation from the
532 available vegetation mapping does not necessarily affect the validation.

533 The second aspect is the difference in the gradients of the target taxa in pollen and vegetation.
534 The validation is applied on a few taxa with considerable amounts both in pollen and vegetation
535 (accounting for 67%-99% of the total pollen assemblages). However, the dominant taxa vary
536 in plant abundance among our sites, and several taxa (e.g. *Vitex*, *Ziziphus*, and *Zanthoxylum*
537 *bungeanum*), have significant cover in the vegetation survey but are rarely present in pollen
538 assemblages. These taxa are therefore not selected for validation. The variation between sites
539 with larger proportions of unselected taxa enhances between-sites variation, and in turn, results
540 in larger SEs of the REVEALS estimates. However, all RVs obtained are with relatively low
541 SEs, and none of the SEs are larger than the RVs in both strategies. Therefore, it is difficult to
542 test if the absence of unselected taxa is responsible for the discrepancies between observed and
543 simulated vegetation abundance.

544 *Vegetation data*

545 Another important and probably more crucial issue is the precision of the estimates of plant
546 abundance, which vary depending on the source and temporal resolution of the vegetation data.
547 One of the caveats of this study is the use of a low spatial resolution vegetation map. To reduce

548 the SE of the RVs, multiple small sites (in strategy I) and the mean RVs of the 36 REVEALS
549 runs based on a single small site (strategy II) were used to estimate regional vegetation. Mean
550 OBV collected from vegetation maps was used to compare with RVs. However, empirical
551 vegetation survey data in anthropogenic landscapes might do not always fulfill the requirements
552 of the ERV and the REVEALS model, such as the important condition of stationarity of the
553 spatial structure of vegetation and plant communities of study areas locally and regionally
554 within the study region at regional spatial scale, although the coarse resolution of the regional
555 vegetation cannot be effectively attributed to spatial variation in vegetation composition at finer
556 scales.

557 Our results show that the RV has a very good fit with OBV for a few taxa, such as *Pinus*,
558 *Artemisia*, *Poaceae*, *Robinia*, *Asteraceae* SF *Cichorium*; however, the similarity between
559 estimated and observed vegetation is less good for some other taxa, and cultivation may
560 responsible for some of the discrepancies. The cover of categorized communities, extracted
561 from available vegetation inventories conducted in the 1980s (Hou, 2019), may have changed,
562 and this is particularly the case for *Castanea*, because *Castanea* preferably grows at the foot of
563 the mountains at low altitudes. The abundance of *Castanea* cultivation has probably increased
564 over the last 40 years due to the extension of farmland through exploring new territory at the
565 foot of the mountains in the study region. Therefore, the validation of RPP estimates needs to
566 be applied to an adequate temporal resolution and areal extent vegetation survey in the future.

567 *Methodological issues*

568 The Landscape Reconstruction Algorithm, which includes REVEALS as the first step, is
569 developed in ideal circumstances for the sake of simplicity. Highly simplified assumptions
570 include: (1) the sedimentary basin for pollen deposition and source plants are placed on a flat
571 plane, (2) there are no spatial gradients in plant distribution and plant community structure in a
572 region, even though vegetation is patchy and heterogeneous, (3) wind blows all directions
573 evenly, (4) pollen grains deposited in the sedimentary basin are all airborne, and (5) source
574 plants for pollen do not grow in the sedimentary basin (Sugita, 2007a,b). In addition,
575 simulations demonstrate that large basins > 200-500 m in radius are ideal for the REVEALS
576 application for regional vegetation reconstruction in a 50-100 km radius from the sampling sites
577 (Sugita 2007a). Our study region is characterized by low mountains, where complex
578 topography limited the availability of large sites, particularly lakes. The application of the
579 REVEALS model on multiple small sites might explain some of the uncertainties of the
580 estimated vegetation abundance in the mountain region. Our results show tree taxa are
581 underrepresented while herbs are overrepresented by the REVEALS model due to the fact that
582 the source area of pollen assemblages in small-sized sites is usually limited in local scale, trees
583 growing at a longer distance were underestimated in the mean regional plant abundance
584 (Prentice, 1985; Sugita, 1994). This has been suggested by an earlier empirical study from the
585 mountain region of the northern Pyrenees (Marquer et al., 2020).

586 There are strengths and weaknesses in both strategies that have been used here. One of the
587 potential drawbacks of strategy I of reducing the number of sites for the ERV model is that it
588 can change the occurrence frequency of target taxa, resulting in shorter gradients. It is not
589 possible to produce reliable RPP estimates for target taxa if there are poor correlations between

590 pollen and vegetation, although there may be strong gradients within all sites in the region to
591 build up a strong pollen-vegetation relationship.

592 Understanding and quantifying the representation of pollen to vegetation is the focus of the
593 ERV model, and also one of the major goals of pollen analysis. Despite the issues discussed
594 above, the RPP values for nine of the ten target taxa are validated as reliable approximations in
595 terms of pollen representation to vegetation.

596 **4.3. Comparison with other RPP validations in China**

597 Three studies have compared and evaluated RPP estimates in China ([Xu et al., 2014](#); [Jiang et](#)
598 [al., 2020](#); [Wan et al., 2022](#)). Regardless of RPP selections, all studies found that vegetation
599 cover estimates using either REVEALS alone or both the REVEALS and LOVE model provide
600 a better representation of vegetation than pollen proportions.

601 Of the three studies, the one conducted in the steppe region in Inner Mongolia found that
602 predicted and observed vegetation abundances match well for most of the pollen taxa evaluated
603 (twelve among 18 taxa), and that the RPP estimates for those taxa are regarded as reasonable
604 ([Xu et al., 2014](#)), although the major caveat of their validation was the circular use of the
605 datasets, without considering spatial autocorrelation. The high similarity between the RPP
606 estimates in our study and those of Xu et al. ([2014](#)) raises the possibility for future validation
607 using the pollen records and observed regional vegetation from one study in combination with
608 RPP estimates obtained from a second study. This is an interesting idea and we could try it in
609 the future, and this would only be possible where RPP estimates are similar, and has not yet
610 been attempted.

611 The RPP estimates validation study in subtropical China of [Jiang et al. \(2020\)](#) found that both
612 local and regional vegetation are closer to observed vegetation than the pollen proportions alone.
613 They estimated regional mean vegetation abundance by the application of the REVEALS model
614 on pollen records from multiple small sites, and the local plant abundances by the application
615 of the LRA with the leave-one-out cross-validation strategy ([Jiang et al., 2020](#)). Similar
616 conclusions were made in tropical China, where the RPP estimates obtained from surface soil
617 were validated by the application of the REVEALS model on extra-surface pollen records from
618 six reservoirs with a radius between 400 to 750 m from the same region to estimate regional
619 vegetation abundance ([Wan et al., 2022](#)). This study was methodologically sound, as it used an
620 independent dataset for validation ([Wan et al., 2022](#)). The strategies that we used within our
621 study (both using LOO, and sub-dividing samples by region) fulfill the necessary validation
622 requirements of independent datasets validation.

623 Previous LRA validation studies have shown that the REVEALS model performs better on
624 groups of taxa (i.e. land-cover types) than individual taxa ([Hellman et al., 2008a](#); [Marquer et](#)
625 [al., 2020](#)), because the discrepancies between simulated and observed plant abundance for
626 individual taxa within one group are counteracted by other taxa within that same group. Our
627 study provides a high level of performance for groups of taxa (e.g. ‘trees’ and ‘herbs’), but
628 grouping taxa does not provide a better fit between RVs and OBVs than for individual taxa.
629 This is due to the fact that almost all tree taxa are underrepresented in the RVs of this study and
630 herbs are overrepresented, hence grouping by tree or herb taxa aggregates discrepancies rather
631 than obscuring them, unlike previous studies.

632

633 **5 Conclusions**

634 This study evaluated the relative pollen productivity (RPP) generated from an earlier study of
635 the cultural landscape of Shandong (Li FR et al., 2017). The results show that the RVs of all
636 taxa are generally closer to OBVs than PPs, and the degree of similarity between RVs and
637 OBVs depends strongly on their abundance in the vegetation and pollen assemblages, it is
638 highest for taxa dominant in the region, such as *Artemisia*, Poaceae, and *Humulus*. The RVs of
639 all herb taxa except *Humulus* and Asteraceae SF *Cichorium*, are slightly over-represented, and
640 the RVs of all tree taxa are under-represented except for *Castanea*. This study confirms the
641 reliability of obtained RPP estimates for nine of ten taxa, *Artemisia*, *Humulus*, *Castanea*,
642 *Quercus*, Asteraceae SF. *Cichorium*, Caryophyllaceae, *Ulmus*, Poaceae, and *Robinia* for their
643 potential applications in the quantitative reconstruction of vegetation abundance in temperate
644 China.

645 The comparison between the RVs and OBVs show that the RVs of all herb taxa are more similar
646 to OBVs collected from shorter distances, while the RVs of all tree taxa are more similar to
647 OBVs collected from longer distances. This study points to the importance of collecting
648 adequate temporal and spatial resolution vegetation survey data from various distances in
649 evaluating the RVs, and that the spatial representation of the REVEALS model is strongly
650 related to the characteristics of taxa, and size of the deposition basin.

651 This study proposed two new alternatives for RPP estimates validation, which promotes the
652 possibility of evaluating the reliability of the RPP estimates using the REVEALS model. The
653 results show that the LOO strategy is the better approach for evaluating the RPP estimates from

654 surface moss polsters. Two better alternatives for evaluating the reliability of the RPP estimates
655 from this data set could be i) using pollen and vegetation from other sets to evaluate the RPP
656 estimates; ii) collecting more RPP values to compare with the obtained RPP estimates for the
657 taxa, such as *Castanea*, whose RPP estimates are not reliable from this or other regions.
658 Therefore, comparative analyses of RPP estimates and validation need to be carried out across
659 regions and vegetation types where the RPP estimates are obtained.

660

661 **Data Availability Statement**

662 The relative pollen productivity estimates and related vegetation composition data will be
663 published alongside the paper as tables and supplementary files.

664

665 **Author Contributions**

666 FL and MJG conceptualized and coordinated the study as a contribution to the PAGES working
667 group “LandCover6k” and collected the pollen samples and vegetation survey in the field. FL,
668 MJG, and SS designed the validation strategies, SS solved all specific issues related to the
669 application of the ERV and REVEALS model. FL prepared pollen samples, counted pollen,
670 and had the major responsibility of vegetation data collection and handling. FL prepared the
671 first draft of the manuscript and all figures and tables and finalized the manuscript for
672 submission. RF and SS edited the manuscript. All co-authors contributed comments about and
673 corrections to the manuscript.

674

675 **Acknowledgments**

676 This work is supported by the National Key R&D Program of China (PI Yan Zhao) [grant
677 number, 2022YFF0801501], the National Science Foundation of China (NSFC) (PI Furong Li)
678 [grant number, 42101143] and funds from the Swedish Strategic Research Area Modelling
679 the Regional and Global Ecosystem, MERGE (<http://www.merge.lu.se/>) (Furong Li (until 2019)
680 and Marie-José Gaillard). This study was undertaken as part of the Past Global Changes
681 (PAGES) project and its working group LandCover6k in turn received support from the Swiss
682 National Science Foundation, the Swiss Academy of Sciences, the US National Science
683 Foundation, and the Chinese Academy of Sciences.

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