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# Wildfire and degradation accelerate northern peatland carbon release

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## Abstract:

The northern peatland carbon sink plays a vital role in climate regulation; however, the future of this carbon sink is uncertain, in part, due to the changing interactions of peatlands and wildfire. Here, we provide the first estimates of carbon emissions associated with boreal and temperate non-permafrost peatlands that specifically include peatland degradation status, wildfire combustion and post-fire dynamics. Wildfire processes reduced the magnitude of carbon uptake in pristine peatlands by 35 % and further enhanced emissions from degraded peatlands by 10 %. The system's current small net sink is vulnerable to the interactions of peatland degraded area, burn rate, and peat burn severity. Modelled climate change impacts accelerated carbon losses and weakened the carbon sink function (burn severity; 38 % reduction and burn rate; 65 % reduction, by 2100), however, we also demonstrate the potential for active peatland restoration to buffer these impacts.

1 Peatlands store approximately one-third of the global soil carbon stock in 3 % of the land  
2 area, making them the most carbon dense ecosystem on Earth<sup>1</sup>. Northern peatlands, in  
3 boreal and temperate regions, account for ~90 % of global peatland area<sup>2</sup> and have  
4 sequestered ~500 Gt C since the last glacial maximum<sup>1,3</sup>, regulating the global climate  
5 throughout the Holocene<sup>4</sup>. Yet, the future of this peatland carbon stock is uncertain<sup>5-7</sup>, in  
6 part, due to the changing interactions of peatlands and wildfire<sup>8-10</sup>. Despite the critical role  
7 of peatlands in the global carbon cycle, recent reports and literature that may influence  
8 policy do not explicitly account for the impacts of fire on peatland emissions estimates  
9 (e.g.,<sup>11</sup>). While estimates of the contribution of peatland drainage to global GHG  
10 emissions have been made<sup>12,13</sup>, no such evaluation has been conducted for, or includes,  
11 the interacting effects of peatland degradation and wildfire. The absence of this  
12 assessment results in additional uncertainty regarding the impact of climate change on  
13 the peatland carbon sink<sup>12</sup>.

14  
15 Carbon emissions from pristine peatland wildfires can vary considerably, however, they  
16 typically average 1–5 kg C m<sup>-2</sup> <sup>10,14,15</sup>. These relatively small peat carbon losses from  
17 combustion can be re-accumulated within 10 to 30 years post-fire<sup>16</sup>, enabling peatlands  
18 to remain a net carbon sink over typical fire-free intervals<sup>17,18</sup>. Conversely, peatland  
19 degradation, such as peatland drainage, not only increases ignition potential<sup>19</sup> but can  
20 also inflate carbon emissions from peatland wildfires by one or more orders of magnitude,  
21 to 10–25 kg C m<sup>-2</sup> equating to 500 to >1000 years of carbon sequestration<sup>10,15,19,20</sup>. Given  
22 that >25 Mha (7 %) of boreal and temperate peatlands have been drained for  
23 anthropogenic use<sup>21</sup>, with some regional or national estimates of ~50 %<sup>11</sup>, these  
24 degraded peatlands represent high risk areas where wildfire could lead to large carbon  
25 emissions.

26  
27 The difference in net carbon fluxes between pristine and drained peatland wildfires are  
28 exacerbated when examining post-fire dynamics. Alterations to CO<sub>2</sub> and methane (CH<sub>4</sub>)  
29 fluxes immediately after fire affect the short-term carbon balance<sup>22-24</sup> while post-fire  
30 vegetation recovery controls the long-term carbon balance<sup>8,16</sup>. While most pristine  
31 peatlands return to a net carbon sink post-fire, evidence suggests that the greater burn  
32 severity in degraded peatlands increases the potential for ecosystem regime shifts<sup>8</sup>, a  
33 change from a carbon accumulating peatland to a carbon releasing ecosystem with non-  
34 peatland vegetation, further increasing the impact of peatland wildfires on long-term  
35 carbon balance. As such, the inclusion of peatland drainage and post-fire net carbon  
36 fluxes is paramount for the accurate evaluation of peatland wildfire carbon emissions.

37  
38 Rapid changes to regional wildfire regimes are compounding the impacts of drainage on  
39 peatland wildfire. In the boreal zone, annual area burned<sup>25</sup> and the frequency of extreme  
40 fire weather conditions<sup>26</sup> are increasing as enhanced evapotranspiration is leading to drier

41 wildfire fuels, particularly in peatland ecosystems<sup>27</sup>. Similarly, in the temperate zone  
42 increased wildfire activity has been associated with severe droughts<sup>28</sup>, and long-term  
43 drying has been observed in peatlands<sup>29</sup>. Increased lightning occurrence, reduced  
44 snowpacks and multi-year droughts are predicted to further increase annual area  
45 burned<sup>30</sup>. Such combinations of climate change-mediated stressors in northern  
46 peatlands, along with pervasive peatland degradation, are likely to increase peatland burn  
47 rate (percent of peatland area burned per year), peat burn severity, and associated  
48 carbon losses<sup>9,10</sup>. Despite evidence that individual northern peat fires can produce  
49 teragrams of carbon emissions<sup>20,31</sup>, the fire return interval (FRI) in northern peatlands is  
50 often only assessed on a regional or per-site basis (e.g.,<sup>32</sup>). The lack of consistent  
51 methodology for assessing peatland burn rate across northern regions has hindered the  
52 evaluation of the current and future contribution of northern peatland fires to global carbon  
53 emissions. Hence, here we provide the first estimates of spatially explicit northern  
54 peatland burn rates and the contribution of peatland wildfire and post-fire dynamics to  
55 global carbon emissions. We then illustrate the impact of peatland degradation and  
56 climate change on the future of the northern peatland carbon sink.

57

### 58 **Empirical modelling of peatland net ecosystem exchange and methane emissions**

59 To address this challenge, we undertook a synthesis of empirical datasets from natural,  
60 degraded (currently drained or previously drained and unrestored), and restored  
61 peatlands in non-permafrost boreal and temperate regions. We then used these data to  
62 model the net ecosystem exchange (NEE; CO<sub>2</sub>) and CH<sub>4</sub> fluxes of peatlands over time,  
63 integrating post-fire dynamics (recovery rate and final NEE) and averaging over a  
64 distribution of FRIs (Table ED1, Methods). The inclusion of peat carbon loss from  
65 combustion and post-fire net carbon fluxes reduced the mean (sd) NEE and CH<sub>4</sub> flux sink  
66 strength from -50.7 (61.8) g C m<sup>-2</sup> yr<sup>-1</sup> (No Burn - natural) to -32.9 (63.2) g C m<sup>-2</sup> yr<sup>-1</sup> in  
67 natural (pristine) peatlands experiencing fire. The moderate (~35 %) reduction in the sink  
68 strength evidences the impact of fire on peatland carbon balance but also the resilience  
69 of the natural peatland carbon sink function under a typical wildfire regime (Fig. 1).

70

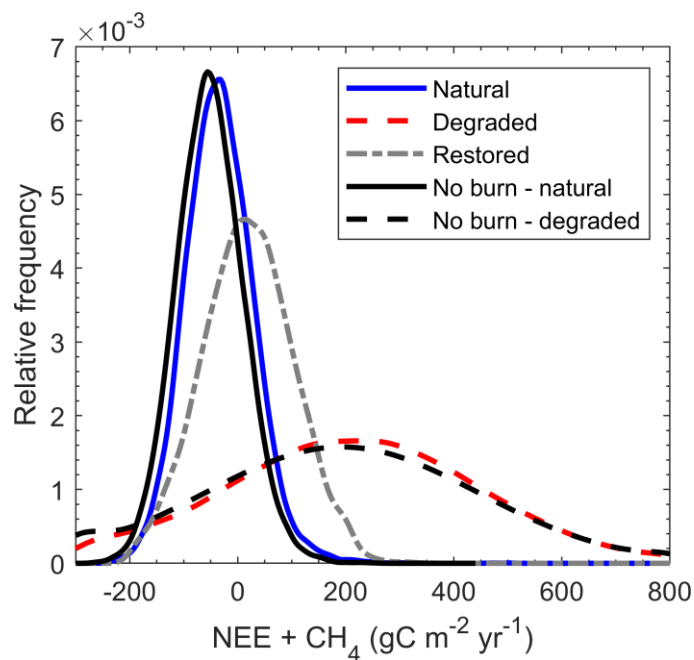
71 Across the variability in burn rate and the impacts of the fire (i.e., severity, recovery rate)  
72 the NEE + CH<sub>4</sub> of degraded peatlands remained a consistent source of carbon with an  
73 average flux of 213 (229) g C m<sup>-2</sup> yr<sup>-1</sup> to the atmosphere, a 10 % increase compared to  
74 No burn – degraded (194 (242) g C m<sup>-2</sup> yr<sup>-1</sup>). Meanwhile, the restoration of peatlands prior  
75 to fire mitigated extensive carbon release (92 % reduction in emissions compared to  
76 Degraded), yet restored peatlands remained a small source of carbon with average NEE  
77 + CH<sub>4</sub> emission of 17.3 (85.5) g C m<sup>-2</sup> yr<sup>-1</sup> (Fig. 1). As such, our modelling indicates that  
78 excluding peatland wildfire from peatland NEE and CH<sub>4</sub> calculations results in a  
79 misrepresentation of peatland carbon balance and may impact estimated regional to

80 national emissions budgets, especially in fire-prone areas with a high proportion of  
81 degraded peatlands.

82

83 Our empirical approach includes uncertainty in the magnitude of peat carbon loss, burn  
84 rate, the rate of recovery, and the initial and final recovered NEE (Methods, Figure ED1).  
85 Our synthesis highlighted limited availability of post-fire carbon flux data, especially from  
86 degraded and restored sites, resulting in a wider distribution of modelled NEE and CH<sub>4</sub>  
87 flux in these scenarios. To further constrain peatland NEE and CH<sub>4</sub> distributions and  
88 accurately include peatlands in earth system models, plot- to ecosystem-scale carbon  
89 flux data at varying times post-fire, especially in degraded and restored ecosystems, is a  
90 critical research need.

91



92

93 Fig. 1. Distribution of net ecosystem exchange (NEE) and methane (CH<sub>4</sub>) fluxes derived from  
94 Monte Carlo simulation model outputs accounting for variation in the magnitude of peat carbon  
95 loss from combustion, burn rate, the rate of recovery, and the initial and final recovered NEE for  
96 peatlands. The same burn rate distribution is used for all peatland states across simulations  
97 (Methods). Categories of peatland states include Natural (pristine), Degraded, and Restored (prior  
98 to fire), and not accounting for wildfire; No burn – natural and No burn - degraded.

99

### 100 **Effect of climate-mediated drying and changes to burn rate on peat fire emissions**

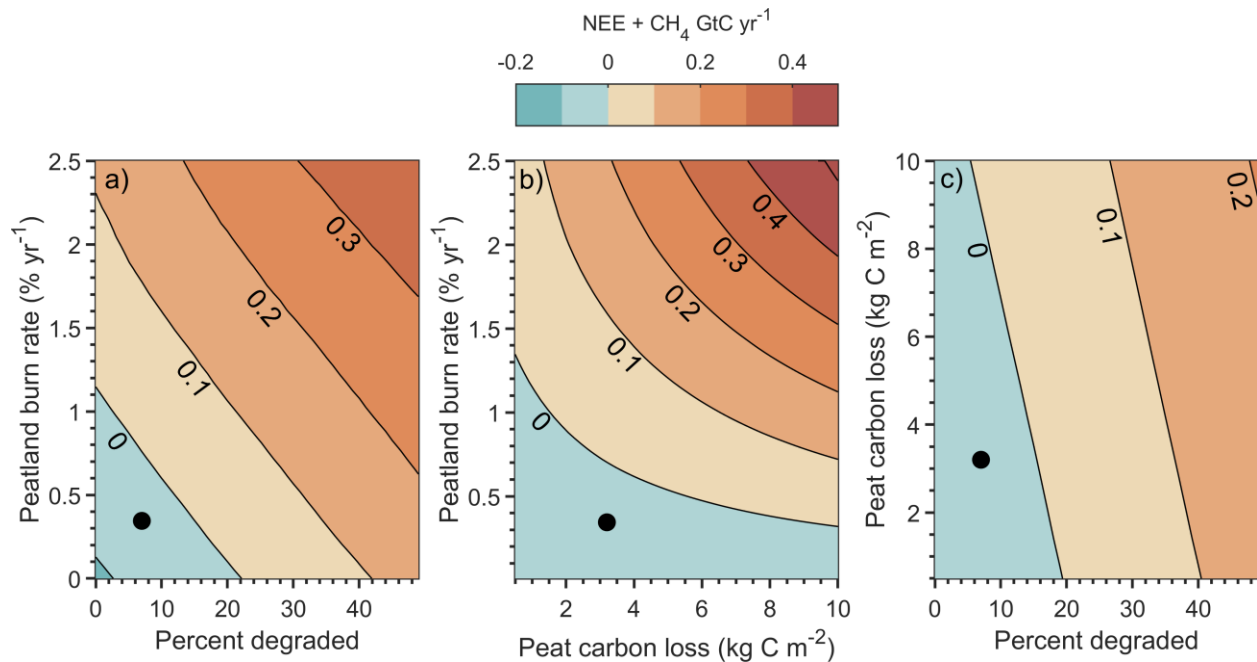
101 In addition to the impact of degradation, peatland NEE + CH<sub>4</sub> fluxes are also sensitive to  
102 the increasing pressures of climate-mediated drying and associated increases in peat  
103 carbon loss from combustion<sup>10</sup>. By aggregating a global dataset of fire perimeters from  
104 2001–2021 (FIRE<sup>44</sup>; Fig. S1), we calculated the average burn rate (percent of land area  
105 burned per year) for boreal and temperate non-permafrost regions over the last two  
106 decades (Methods). Average burn rate varied between 0.0001 and 1.48 % yr<sup>-1</sup> amongst

107 boreal and temperate ecoregions (Table S1), with a spatially weighted average of 0.35 %  
108 yr<sup>-1</sup>, equivalent to a FRI of 290 years. Assessment of the relationship between peatland  
109 (histosol) areal coverage<sup>3</sup> and burn rate found no significant trends (Fig. ED2, S2 and  
110 S3), suggesting that peatland cover does not exert a strong control over regional area  
111 burned. Compilation of national inventories found that degradation due to drainage for  
112 agriculture, horticulture, and forestry varies between <1 and 54 % of peatland area per  
113 country<sup>34</sup> and the proportion of total drained northern peatland area is ~7 % (26.1 Mha;  
114 <sup>21</sup>). These data were used to evaluate the current state of the boreal and temperate non-  
115 permafrost peatland system.

116  
117 At the broadest scale, without accounting for future climate change impacts to peatlands  
118 or wildfire regimes, we estimate that the total NEE + CH<sub>4</sub> flux for boreal and temperate  
119 non-permafrost peatlands is a small net carbon sink (filled dots Fig. 2), however, the  
120 system becomes a net carbon source given an annual average peatland burn rate of  
121 more than 0.77 % based on the current estimates of drained peatland area (Fig. 2a).  
122 Accordingly, and important for regional carbon balances, a greater percentage of  
123 degraded peatlands reduces the burn rate required to switch the system from a net carbon  
124 sink to a net carbon source by 0.05 % yr<sup>-1</sup> per additional 1 % degraded peatland area.

125  
126 Similarly, increased peat carbon loss from combustion reduces the carbon sink strength  
127 and may contribute to switching the system to a net source. Increasing the average peat  
128 carbon loss from combustion in pristine peatlands to represent a moderate degree of  
129 climate change drying (1.5 kg C m<sup>-2</sup> added to the original distribution; Methods) reduces  
130 the annual burn rate required to switch from a carbon sink to source to 0.55 % (Fig. 2b).  
131 This equates to a required lengthening of the average FRI by ~50 years to maintain active  
132 net carbon sequestration at a landscape level. Further, there is a strong interactive effect  
133 of percent degraded and peat carbon loss on NEE + CH<sub>4</sub> (Fig. 2c). Using the spatially  
134 weighted average burn rate of 0.35 % yr<sup>-1</sup>, NEE + CH<sub>4</sub> fluxes are sensitive to changes in  
135 percent degraded, where a relatively small reduction in percent degraded (e.g., by one-  
136 third from 15 to 10 %) via active restoration counteracts potential increases in average  
137 peat carbon loss from combustion caused by climate-mediated drying.

138



139  
140

141 Fig. 2. The interactive effect of fire regime changes and degraded peatland area, on NEE + CH<sub>4</sub>  
 142 (GtC yr<sup>-1</sup>). a) Peatland burn rate and percent degraded, where peat carbon loss is weighted based  
 143 on percent degraded. b) Peatland burn rate and peat carbon loss, where percent degraded is  
 144 held constant at 7 %. c) Peat carbon loss and percent degraded, with a 0.35 % (spatially-weighted  
 145 average) burn rate. Filled dots represent the current boreal and temperate non-permafrost  
 146 peatland system in the NEE and CH<sub>4</sub> flux phase space. Axes do not show zero where a zero  
 147 value results in a no-data point.

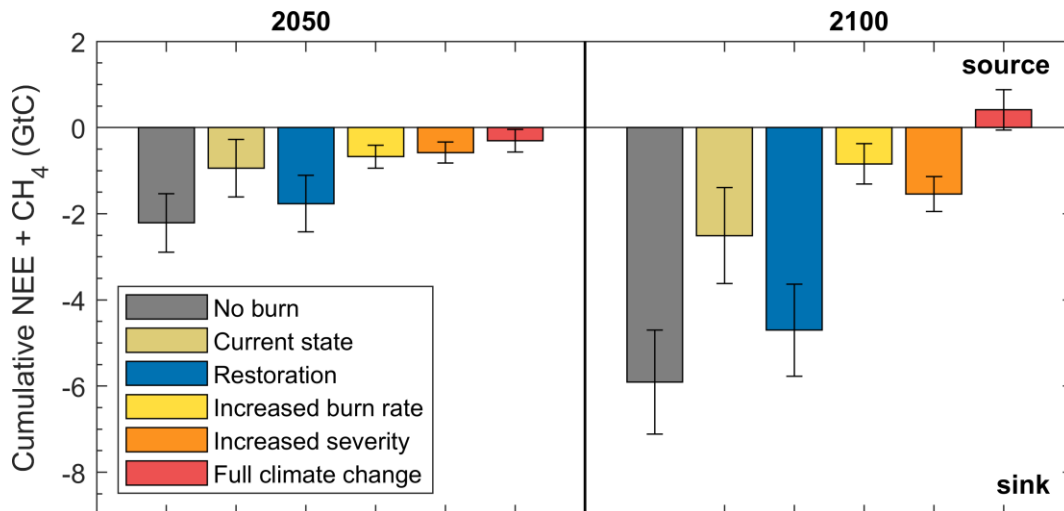
148

### 149 The future of the northern peatland carbon sink

150 To illustrate the impact of peatland degradation status and climate change on the  
 151 magnitude of the peatland carbon sink we evaluated cumulative annual net fluxes from  
 152 our NEE + CH<sub>4</sub> simulations. We developed scenarios that combine different peatland  
 153 degradation and climate change factors and assessed the impact on total carbon  
 154 sequestration (or emission) by 2050 and 2100 (Methods). Scenarios include i) No burn,  
 155 ii) Current state, iii) Restoration, iv) Increased burn rate, v) Increased (burn) severity, and  
 156 vi) Full climate change (increased burn rate and burn severity).

157





158  
159

160 Fig. 3. Cumulative NEE + CH<sub>4</sub> flux (GtC) for boreal and temperate non-permafrost peatlands in  
161 2050 and 2100. Negative values represent a carbon sink while positive values represent a carbon  
162 source. Scenarios include: Current state, No burn (Current state with 0 % burn rate), Restoration  
163 (100 % of degraded peatlands restored), Increased burn rate (annual rate doubling by 2100),  
164 Increased severity (additional 1.5 kg C m<sup>-2</sup> loss across all peatland states), and Full climate  
165 change (increased burn rate and severity). Error bars (±1 sd) represent the uncertainty estimated  
166 via Monte Carlo simulations using distributions of burn rate and NEE + CH<sub>4</sub>.

167

168 Accounting for peatland wildfire emissions reduces the magnitude of the estimated  
169 peatland carbon sink by 1.3 GtC, or 57 %, by 2050 when comparing the No burn scenario  
170 (-2.2 GtC) to our Current scenario (-0.94 GtC; Fig. 3). Meanwhile, the restoration of all  
171 degraded peatlands (Restoration scenario) results in a sink of 1.8 GtC by 2050, an  
172 additional 0.82 GtC sequestered compared to the Current scenario, evidencing the short-  
173 term gains to be made from peatland restoration. Restoration increases the carbon sink  
174 by almost 90 % in 2100, increasing it from a sink of 2.5 (Current scenario) to 4.7 GtC  
175 (Restoration scenario).

176

177 Increasing peatland burn rate (linear increase to 0.7 % by 2100) and increasing burn  
178 severity (+1.5 kg C m<sup>-2</sup> peat carbon loss) decrease the peatland carbon sink by similar  
179 amounts by 2050 with a 0.25 and 0.36 GtC reduction relative to the Current scenario,  
180 respectively. However, the modelled burn rate increase throughout the remainder of the  
181 century results in a large decrease in the carbon sink strength in the Increased burn rate  
182 scenario by 2100, reducing the total carbon sequestration to a sink of 0.88 GtC, a 65 %  
183 decrease compared to the Current scenario (-2.5 GtC). Concerningly, when the impacts  
184 of climate change are combined (Full climate change scenario) the system shows a  
185 potential switch from a carbon sink to a carbon source, with a mean estimated source of  
186 0.4 GtC to the atmosphere by 2100. The acceleration of carbon release from boreal and  
187 temperate non-permafrost peatlands and associated diminishment of the strength of the

188 carbon sink over the coming decades has critical implications for global climate change  
189 and emissions targets.

## 190 **Assessing the importance of peatland restoration for global climate change and** 191 **emissions targets**

192 This study highlights the resilience of pristine northern peatland ecosystems to wildfire,  
193 with natural peatlands returning to a net carbon sink in most of our simulations across the  
194 range of fire severity and post-fire dynamics. Conversely, we demonstrate unequivocally  
195 that degraded peatlands are responsible for the largest peatland carbon emissions<sup>19,20,31</sup>.  
196 We show that the restoration of degraded peatlands prior to fire greatly reduces long-term  
197 emissions<sup>33</sup>. Our results add to the growing literature base that suggests climate and land-  
198 use change increase the vulnerability of peatland ecosystems and their carbon stocks to  
199 fire, with significant and far-reaching ecological, hydrological, and societal  
200 consequences<sup>34,35</sup>.

201  
202 While future anthropogenic fossil fuel emissions can be curbed, the climatic changes  
203 already induced by rising atmospheric CO<sub>2</sub> concentrations will likely continue to increase  
204 peatland wildfire emissions over the coming century, reducing the strength of the peatland  
205 carbon sink. We show that although the peatland carbon sink is currently resilient,  
206 changes in degraded peatland area, average burn rate (FRI) and peat burn severity may  
207 lead to climate neutrality or net carbon release. Climate-mediated peatland drying across  
208 the spectrum of peatland condition<sup>27,29</sup> could contribute to increases in peatland burn  
209 rate<sup>36</sup> and peat carbon loss via enhanced burn severity<sup>10</sup> in line with the increasing  
210 availability of critically dry peatland fuels<sup>9</sup>. Forested peatlands (natural or managed) may  
211 be more prone to positive (amplifying) ecohydrological feedbacks that promote high  
212 severity smouldering fire<sup>37</sup>, when compared to arable peatlands in northern regions,  
213 however, the vulnerability of peatlands to wildfire under different management regimes is  
214 currently relatively unstudied.

215  
216 To maintain the northern peatland carbon sink function, decreases in the area of  
217 degraded peatland through active peatland restoration must occur to counteract potential  
218 increases in average peat carbon loss due to climate-mediated drying. Our restoration  
219 scenario (representing the restoration of all degraded peatlands) resulted in an estimated  
220 increase in the carbon sink by almost 90 % by 2100 compared to the current scenario.  
221 Despite the hypothetical nature of our restoration scenario, it serves to support research  
222 highlighting the important role peatlands can play in reducing global emissions if they are  
223 protected<sup>38</sup> and restored<sup>39</sup> appropriately. We also strongly advocate for better  
224 management of carbon-rich ecosystems alongside behavioural changes to stop  
225 accidental and unnecessary ignitions<sup>40</sup> especially areas with a high proportion of  
226 degraded peatlands (e.g., Europe)<sup>11</sup>.

227  
228 On a regional level we provide evidence of the importance of accurately measuring  
229 (degraded) peatland area, as well as burn rate, since these factors will affect the ability  
230 of countries/regions to account for emissions and potentially, to achieve emissions  
231 targets. The proportion of peatlands affected by land use change varies considerably  
232 between countries and regions but can be substantial (<1 to ~50 % degraded<sup>41</sup>). While  
233 there are likely differences in the ignition potential of different peatland land-uses<sup>19</sup> there  
234 is a scarcity of these data in the literature. Peatland type and landscape position have  
235 been found to impact burn rate<sup>32</sup> and fire severity<sup>42</sup>, yet peatlands are often misclassified  
236 in fire risk, spread, and emissions models<sup>43</sup>, highlighting the need to improve peatland  
237 mapping for use alongside remotely sensed fire products (e.g.,<sup>44</sup>). Appropriate accounting  
238 of carbon emissions from peatlands, accounting for wildfire, may guide national/regional  
239 restoration and conservation strategies (e.g., the UK<sup>45</sup>).

240  
241 Interdisciplinary collaborations will be crucial to accurately represent the northern  
242 peatland carbon balance in earth system models and ensure community- to international-  
243 level climate policies include important peatland processes, such as fire, in their strategies  
244 to maintain the impacts of climate change within liveable bounds. While remote-sensing  
245 applications, such as FIRED<sup>44</sup>, have enabled consistent burn rate mapping across large  
246 regions, the limited precision and consistency of peatland type and carbon stock maps  
247 creates challenges for further reducing the uncertainty surrounding estimates of the  
248 strength of the northern peatland carbon sink<sup>38</sup>. Further, our carbon sink estimates don't  
249 account for fluvial export of carbon, nor the anthropogenic additions/removals of biomass  
250 on agricultural peatlands. Data corresponding to methane emissions in different peatland  
251 types immediately post-fire e.g.,<sup>22,23</sup> would further constrain estimates of the peatland  
252 carbon sink.

253  
254 The direction and magnitude of the peatland-climate feedback will be driven by the  
255 combined effects of peatland degradation and restoration, and the global emissions  
256 pathway that will influence rates of climate-induced drying<sup>27</sup> and changes in burn rate<sup>26</sup>.  
257 Northern peatlands have regulated global climate over the Holocene but if the predicted  
258 increases in peat burn severity and fire activity outweigh carbon sequestration from  
259 peatland expansion in high-latitude regions<sup>47</sup>, the northern peatland system will become  
260 a shrinking carbon sink and a potential future carbon source, exacerbating the rapidly  
261 closing window of time to avoid the most severe impacts of global climate change. Our  
262 scenario results found that increasing burn rate and peat burn severity drastically reduced  
263 the amount of carbon sequestered in peatlands, but overall the system maintained a  
264 carbon sink status in 2050. However, the continued and compounding impacts of climate  
265 change resulted in an estimated small net carbon source to the atmosphere by boreal  
266 and temperate non-permafrost peatlands by 2100. Notably this estimate does not account

267 for increased peatland tree growth stimulated by drier conditions<sup>10</sup>, however, given the  
268 positive correlation between tree size and burn severity<sup>10</sup> and the dominance of long-term  
269 carbon storage in peat rather than above-ground vegetation<sup>48</sup>, it is unlikely that increases  
270 in above-ground biomass will translate into significant increases in carbon storage over  
271 long (>1 FRI) time periods. The likely reduction in peatland carbon sink strength will create  
272 further challenges to remaining below critical global climate targets.

273

274 Against the global backdrop of increases in burn rate and extreme wildfire weather<sup>26</sup>,  
275 integrated regional wildfire management solutions are urgently required to mitigate severe  
276 climatic and societal impacts of peatland wildfire<sup>36,37</sup>. In regions with higher proportions  
277 of peatland degradation we find that a strong trade-off with burn rate (i.e., large  
278 investments in direct fire suppression) is required to preserve the critical climate  
279 regulation function of peatlands. Where this balance is not maintained peatland wildfire  
280 emissions may represent an under-appreciated source component in carbon accounting  
281 that could be detrimental to achieving emissions targets. We demonstrate here that,  
282 despite notable impacts of peatland burn rate and burn severity, peatland restoration  
283 represents a large opportunity to minimize impacts to the boreal and temperate peatland  
284 carbon sink over the coming century when accounting for peatland wildfire emissions.  
285 Our results suggest an immediate need to start including active restoration of degraded  
286 peatlands as a cost-effective tool to support the mitigation of extensive carbon emissions  
287 and detrimental impacts on human health.

288

289

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298  
299 **Competing Interests**

300 The authors have no relevant financial or non-financial interests to disclose.

301  
302 **Author Contributions**

303 SW: Conceptualisation, data curation, formal analysis, methodology, visualization, writing  
304 – original draft, writing – review & editing. RA: Conceptualisation, methodology,  
305 visualization, writing – review & editing. PM: Data curation, formal analysis, methodology,  
306 writing – review & editing. SD: Data curation, writing – original draft, writing – review &  
307 editing. GG: Data curation, writing – original draft, writing – review & editing. JMW:  
308 Conceptualization, funding acquisition, methodology, supervision, writing – review &  
309 editing.

310  
311 **References**

- 312 1. Yu, Z., Loisel, J., Brosseau, D.P., Beilman, D.W., and Hunt, S.J. 2010. Global Peatland  
313 Dynamics since the Last Glacial Maximum. *Geophysical Research Letters* 37 (13):  
314 L13402. <https://doi.org/10.1029/2010GL043584>
- 315 2. Xu, J., Morris, P.J. Liu, J. and Holden, J. 2018. PEATMAP: Refining Estimates of Global  
316 Peatland Distribution Based on a Meta-Analysis. *Catena*, 160: 134–40.  
317 <https://doi.org/10.1016/j.catena.2017.09.010>
- 318 3. Hugelius, G., Loisel, J., Chadburn, S., Jackson, R.B., Jones, M., MacDonald, G.,  
319 Marushchak, M. et al. 2020. Large Stocks of Peatland Carbon and Nitrogen Are Vulnerable  
320 to Permafrost Thaw. *Proceedings of the National Academy of Sciences of the United*  
321 *States of America*, 117 (34): 20438–46. <https://doi.org/10.1073/pnas.1916387117>
- 322 4. Frolking, S. and Roulet, N.T. 2007. Holocene Radiative Forcing Impact of Northern Peatland  
323 Carbon Accumulation and Methane Emissions. *Global Change Biology*, 13 (5): 1079–88.  
324 <https://doi.org/10.1111/J.1365-2486.2007.01339.X>
- 325 5. Gallego-Sala, A.V., Charman, D.J., Brewer, S., Page, S.E., Prentice, I.C., Friedlingstein, P.,  
326 Moreton, S. et al. 2018. Latitudinal Limits to the Predicted Increase of the Peatland Carbon  
327 Sink with Warming. *Nature Climate Change*, 2018 8:10 8 (10): 907–13.  
328 <https://doi.org/10.1038/s41558-018-0271-1>
- 329 6. Ferretto, A., Brooker, R., Aitkenhead, M., Matthews, R. and Smith, P. 2019. Potential Carbon  
330 Loss from Scottish Peatlands under Climate Change. *Regional Environmental Change*, 19  
331 (7): 2101–11. <https://doi.org/10.1007/S10113-019-01550-3/TABLES/2>
- 332 7. Loisel, J., Gallego-Sala, A.V., Amesbury, M.J., Magnan, G., Anshari, G., Beilman, D.W.,  
333 Benavides, J.C., et al. 2020. Expert Assessment of Future Vulnerability of the Global  
334 Peatland Carbon Sink. *Nature Climate Change* 2020 11:1 11 (1): 70–77.

- 335 <https://doi.org/10.1038/s41558-020-00944-0>
- 336 8. Kettridge, N., Turetsky, M.R., Sherwood, J.H., Thompson, D.K., Miller, C.A., Benscoter, B.W.,  
337 Flannigan, M.D., Wotton, B.M. and Waddington, J.M. 2015. Moderate Drop in Water Table  
338 Increases Peatland Vulnerability to Post-Fire Regime Shift. *Scientific Reports*, 5: 8063.  
339 <https://doi.org/10.1038/srep08063>
- 340 9. Turetsky, M.R., Benscoter, B.W., Page, S.E., Rein, G., Van Der Werf, G.R., and Watts, A.  
341 2015. Global Vulnerability of Peatlands to Fire and Carbon Loss. *Nature Geoscience* 8:1 8  
342 (1): 11–14. <https://doi.org/10.1038/ngeo2325.2014b>
- 343 10. Wilkinson, S. L., Moore, P.A., Flannigan, M.D., Wotton, B.M. and Waddington, J.M. 2018.  
344 Did Enhanced Afforestation Cause High Severity Peat Burn in the Fort McMurray Horse  
345 River Wildfire? *Environmental Research Letters*, 13 (1): 014018.  
346 <https://doi.org/10.1088/1748-9326/AAA136>
- 347 11. UNEP. 2022. Peatland Emissions; Section 2.7.3 in “Global Peatlands Assessment – The  
348 State of the World’s Peatlands: Evidence for action toward the conservation, restoration,  
349 and sustainable management of peatlands”. Main Report. Global Peatlands Initiative.  
350 United Nations Environment Programme, Nairobi, Kenya.
- 351 12. Leifeld, J., Wüst-Galley, C. and Page, S. 2019. Intact and Managed Peatland Soils as a  
352 Source and Sink of GHGs from 1850 to 2100. *Nature Climate Change* 2019 9:12 9 (12):  
353 945–47. <https://doi.org/10.1038/s41558-019-0615-5>
- 354 13. Humpenöder, F., Karstens, K., Lotze-Campen, H., Leifeld, J., Menichetti, L., Barthelmes, A.  
355 and Popp, A. 2020. Peatland Protection and Restoration Are Key for Climate Change  
356 Mitigation. *Environmental Research Letters*, 15 (10): 104093. <https://doi.org/10.1088/1748-9326/ABAE2A>
- 357
- 358 14. Poulter, B., Christensen, N.L. and Halpin, P.H. 2006. Carbon Emissions from a Temperate  
359 Peat Fire and Its Relevance to Interannual Variability of Trace Atmospheric Greenhouse  
360 Gases. *Journal of Geophysical Research: Atmospheres*, 111 (D6): 6301,  
361 <https://doi.org/10.1029/2005JD006455>
- 362 15. Turetsky, M. R., Donahue, W.F., and Benscoter, B.W. 2011. Experimental Drying Intensifies  
363 Burning and Carbon Losses in a Northern Peatland. *Nature Communications*, 2 (1).  
364 <https://doi.org/10.1038/ncomms1523>
- 365 16. Wieder, R.K., Scott, K.D., Kamminga, K., Vile, M.A., Vitt, D.H., Bone, T., Xu, B., Benscoter,  
366 B.W. and Bhatti, J.S. 2009. Postfire Carbon Balance in Boreal Bogs of Alberta, Canada.  
367 *Glob. Change Biol.*, 15 (1): 63–81. <https://doi.org/10.1111/j.1365-2486.2008.01756.x>
- 368 17. Kuhry, P. 1994. The role of fire in the development of Sphagnum-dominated peatlands in  
369 western boreal Canada. *Journal of Ecology*, 899-910. <https://doi.org/10.2307/2261453>
- 370 18. Ingram, R. C., Moore, P. A., Wilkinson, S.L., Petrone, R.M. and Waddington, J.M. 2019.  
371 Postfire Soil Carbon Accumulation Does Not Recover Boreal Peatland Combustion Loss in  
372 Some Hydrogeological Settings. *Journal of Geophysical Research: Biogeosciences*, 124  
373 (4): 775–88. <https://doi.org/10.1029/2018JG004716>
- 374 19. McCarter, C. P.R., Wilkinson, S.L., Moore, P.A. and Waddington, J.M. 2021.  
375 Ecohydrological Trade-Offs from Multiple Peatland Disturbances: The Interactive Effects of  
376 Drainage, Harvesting, Restoration and Wildfire in a Southern Ontario Bog. *Journal of*  
377 *Hydrology* 601 (October). <https://doi.org/10.1016/j.jhydrol.2021.126793>
- 378 20. Davies, G.M., Gray, A., Rein, G. and Legg, C.J. 2013. Peat Consumption and Carbon Loss

- 379 Due to Smouldering Wildfire in a Temperate Peatland. *Forest Ecology and Management*,  
380 308: 169–77
- 381 21. Leifeld, J., and Menichetti, L. 2018. The Underappreciated Potential of Peatlands in Global  
382 Climate Change Mitigation Strategies. *Nat. Commun.*, 9 (1).  
383 <https://doi.org/10.1038/s41467-018-03406-6>
- 384 22. Davidson, S.J., Van Beest, C., Petrone, R. and Strack, M. 2019. Wildfire Overrides  
385 Hydrological Controls on Boreal Peatland Methane Emissions. *Biogeosciences*, 16 (13):  
386 2651–60. <https://doi.org/10.5194/BG-16-2651-2019>.
- 387 23. Gray, A., Davies, G.M., Domènech, R., Taylor, E. and Levy, P.E. 2020. “Peatland Wildfire  
388 Severity and Post-Fire Gaseous Carbon Fluxes. *Ecosystems*, 24 (3): 713–25.  
389 <https://doi.org/10.1007/s10021-020-00545-0>
- 390 24. Morison, M. Q., Petrone, R.M., Wilkinson, S.L., Green, A. and Waddington, J.M. 2020.  
391 Ecosystem Scale Evapotranspiration and CO<sub>2</sub> Exchange in Burned and Unburned  
392 Peatlands: Implications for the Ecohydrological Resilience of Carbon Stocks to Wildfire.  
393 *Ecohydrology*, 13 (2): e2189. <https://doi.org/10.1002/ECO.2189>
- 394 25. Hanes, C.C., Wang, X., Jain, P., Parisien, M.A., Little, J.M. and Flannigan, M.D. 2019. Fire-  
395 Regime Changes in Canada over the Last Half Century. *Canadian Journal of Forest*  
396 *Research*, 49 (3): 256–69. [https://doi.org/10.1139/CJFR-2018-0293/SUPPL\\_FILE/CJFR-  
397 2018-0293SUPPLA.DOCX](https://doi.org/10.1139/CJFR-2018-0293/SUPPL_FILE/CJFR-2018-0293SUPPLA.DOCX)
- 398 26. Jain, P., Castellanos-Acuna, D., Coogan, S.C.P., Abatzoglou, J.T. and Flannigan, M.D.  
399 2021. Observed Increases in Extreme Fire Weather Driven by Atmospheric Humidity and  
400 Temperature. *Nature Climate Change*, 2021 12:1 12 (1): 63–70.  
401 <https://doi.org/10.1038/s41558-021-01224-1>
- 402 27. Helbig, M., Waddington, J.M., Alekseychik, P., Amiro, B.D., Aurela, M., Barr, A.G., Black,  
403 T.A. et al. 2020. Increasing Contribution of Peatlands to Boreal Evapotranspiration in a  
404 Warming Climate. *Nature Climate Change*, 10 (6): 555–60. [https://doi.org/10.1038/s41558-  
405 020-0763-7](https://doi.org/10.1038/s41558-020-0763-7)
- 406 28. Hari, V., Rakovec, O., Markonis, Y., Hanel, M. and Kumar, R. 2020. Increased Future  
407 Occurrences of the Exceptional 2018–2019 Central European Drought under Global  
408 Warming. *Scientific Reports*, 10 (1): 1–10. <https://doi.org/10.1038/s41598-020-68872-9>
- 409 29. Swindles, G.T., Morris, P.J., Mullan, D.J., Payne, R.J., Roland, T.P., Amesbury, M.J.,  
410 Lamentowicz, M. et al. 2019. Widespread Drying of European Peatlands in Recent  
411 Centuries. *Nature Geoscience*, 2019 12:11 12 (11): 922–28.  
412 <https://doi.org/10.1038/s41561-019-0462-z>
- 413 30. Půčik, T., Groenemeijer, P., Rädler, A.T., Tijssen, L., Nikulin, G., Prein, A.F., van Meijgaard,  
414 E., Fealy, R., Jacob, D. and Teichmann, C. 2017. Future Changes in European Severe  
415 Convection Environments in a Regional Climate Model Ensemble. *Journal of Climate*, 30  
416 (17): 6771–94. <https://doi.org/10.1175/JCLI-D-16-0777.1>.
- 417 31. Mickler, R.A., Welch, D.P. and Bailey, A.D. 2017. Carbon Emissions during Wildland Fire on  
418 a North American Temperate Peatland. *Fire Ecology*, 13 (1): 34–57.  
419 <https://doi.org/10.4996/fireecology.1301034>
- 420 32. Turetsky, M. R., Amiro, B. D., Bosch, E., & Bhatti, J. S. 2004. Historical burn area in western  
421 Canadian peatlands and its relationship to fire weather indices. *Global Biogeochemical*  
422 *Cycles*, 18(4).

- 423 33. Granath, G., Moore, P.A. Lukenbach, M.C. and Waddington, J.W. 2016. Mitigating Wildfire  
424 Carbon Loss in Managed Northern Peatlands through Restoration. *Sci. Rep.* 6: 28498.  
425 <https://doi.org/10.1038/srep28498>
- 426 34. Crump, J. 2017. Smoke on Water: Countering Global Threats from Peatland Loss and  
427 Degradation | GRID-Arendal. <https://www.grida.no/publications/355>
- 428 35. UNEP. 2022. Spreading like Wildfire: The Rising Threat of Extraordinary Landscape Fires |  
429 UNEP - UN Environment Programme. [https://www.unep.org/resources/report/spreading-  
430 wildfire-rising-threat-extraordinary-landscape-fires](https://www.unep.org/resources/report/spreading-wildfire-rising-threat-extraordinary-landscape-fires)
- 431 36. Thompson, D. K., Simpson, B. N., Whitman, E., Barber, Q. E., & Parisien, M. A. (2019).  
432 Peatland hydrological dynamics as a driver of landscape connectivity and fire activity in the  
433 boreal plain of Canada. *Forests*, 10(7), 534.
- 434 37. Nelson, K., Thompson, D., Hopkinson, C., Petrone, R.M. and Chasmer, L. 2021. Peatland-  
435 Fire Interactions: A Review of Wildland Fire Feedbacks and Interactions in Canadian  
436 Boreal Peatlands. *Science of the Total Environment.*, 769: 145212  
437 <https://doi.org/10.1016/j.scitotenv.2021.145212>.
- 438 38. Harris, L.I., Richardson, K., Bona, K.A., Davidson, S.J., Finkelstein, S.A., Garneau, M.,  
439 McLaughlin, J. et al. 2021. The Essential Carbon Service Provided by Northern Peatlands.  
440 *Frontiers in Ecology and the Environment*. <https://doi.org/10.1002/FEE.2437>.
- 441 39. Strack, M., Davidson, S. J., Hirano, T., & Dunn, C. 2022. The Potential of Peatlands as  
442 Nature-Based Climate Solutions. *Current Climate Change Reports*, 1-12.
- 443 40. Ganteaume, A., Camia, A., Jappiot, M., San-Miguel-Ayanz, J., Long-Fournel, M. and  
444 Lampin, C. 2013. A Review of the Main Driving Factors of Forest Fire Ignition Over Europe.  
445 *Environmental Management*, 51: 651–662 <https://doi.org/10.1007/s00267-012-9961-z>
- 446 41. Strack, M. (Ed.). 2008. *Peatlands and climate change*. IPS, International Peat Society.
- 447 42. Wilkinson, S. L., Moore, P. A., & Waddington, J. M. 2019. Assessing drivers of cross-scale  
448 variability in peat smoldering combustion vulnerability in forested boreal  
449 peatlands. *Frontiers in Forests and Global Change*, 2, 84.  
450 <https://doi.org/10.3389/ffgc.2019.00084>
- 451 43. Kukavskaya, E.A., Soja, A.J., Petkov, A.P., Ponomarev, E.I., Ivanova, G.A., and Conard,  
452 S.G. 2013. Fire Emissions Estimates in Siberia: Evaluation of Uncertainties in Area  
453 Burned, Land Cover, and Fuel Consumption. *Canadian Journal of Forest Research*, 43 (5):  
454 493–506. [https://doi.org/10.1139/CJFR-2012-0367/ASSET/IMAGES/LARGE/CJFR-2012-  
455 0367F6.JPEG](https://doi.org/10.1139/CJFR-2012-0367/ASSET/IMAGES/LARGE/CJFR-2012-0367F6.JPEG)
- 456 44. Mahood, A. L., Lindrooth, E. J., Cook, M. C., & Balch, J. K. (2022). Country-level fire  
457 perimeter datasets (2001–2021). *Scientific data*, 9(1), 458. [https://doi.org/10.1038/s41597-  
458 022-01572-3](https://doi.org/10.1038/s41597-022-01572-3)
- 459 45. Evans, C., Artz, R., Moxley, J., Smyth, M.A., Taylor, E., Archer, E., Burden, A., Williamson,  
460 J., Donnelly, D., Thomson, A. and Buys, G. 2017. Implementation of an emissions  
461 inventory for UK peatlands (pp. 1-88). Centre for Ecology and Hydrology.
- 462 46. Mishra, S., Page, S.E. Cobb, A.R. Ser Huay Lee, J., Jovani-Sancho, A.J., Sjögersten, S.,  
463 Jaya, A., Aswandi, and Wardle, D.A. 2021. Degradation of Southeast Asian Tropical  
464 Peatlands and Integrated Strategies for Their Better Management and Restoration. *Journal  
465 of Applied Ecology*, 58 (7): 1370–87. <https://doi.org/10.1111/1365-2664.13905>
- 466 47. Magnan, G., Sanderson, N.K., Piilo, S., Pratte, S., Väiliranta, M., van Bellen, S., Zhang, H.



467 and Garneau, M. 2022. Widespread Recent Ecosystem State Shifts in High-Latitude  
 468 Peatlands of Northeastern Canada and Implications for Carbon Sequestration. *Global*  
 469 *Change Biology*, 28 (5): 1919–34. <https://doi.org/10.1111/GCB.16032>

470 48. Beaulne, J., Garneau, M., Magnan, G., & Boucher, É. 2021. Peat deposits store more  
 471 carbon than trees in forested peatlands of the boreal biome. *Scientific reports*, 11(1), 1-11.

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 473 **Extended Data**

474  
 475 Table ED1. Input parameters derived from data synthesis used in Monte Carlo simulations for  
 476 calculation of peatland net ecosystem carbon balance. The restored group here includes  
 477 rewetted sites. The fire return interval used in the model is taken from the burn rate as 100/(burn  
 478 rate).

Input	State	Distribution	Parameter 1	Parameter 2
NEE (g C m <sup>-2</sup> yr <sup>-1</sup> )	burned	Normal	Mean = 71.4	SE = 53.6
	degraded	Normal	Mean = 191.8	SE = 249.5
	restored	Normal	Mean = -5.7	SE = 84.7
	pristine	Normal	Mean = -62.6	SD = 57.8
Fire C-loss (kg C)	degraded	log-normal	Mean = 1.846	SD = 0.846
	pristine	log-normal	Mean = 0.587	SD = 0.907
Burn rate (% yr <sup>-1</sup> )	–	exponential	Mean = 0.345	N/A
$t_1$	–	uniform	Min. = 1	Max. = 10
$t_2$	–	uniform	Min. = 11	Max = 60

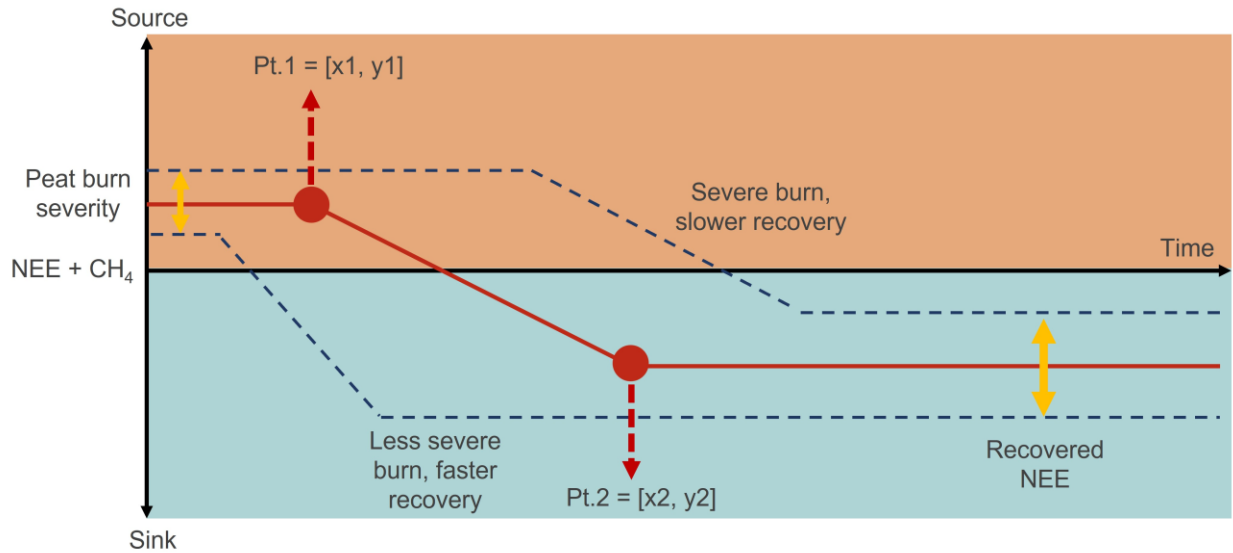
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 482 Table ED2. Data from FIRED (non-permafrost land area) with area burned over a 19.75 year  
 483 period from 2001 to 2021. Only ecoregions within each biome which contained peatlands  
 484 (histosols) were considered.

<b>Region</b>	<b>Biome</b>	<b>Total area (10<sup>6</sup> km<sup>2</sup>)</b>	<b>Burned area (10<sup>6</sup> km<sup>2</sup>)</b>	<b>Fire return interval (years)</b>	<b>Burn rate (% yr<sup>-1</sup>)</b>
Asia	Boreal	3.08	0.261	233	0.43
	Temperate	3.93	0.452	172	0.58
Europe	Boreal	2.32	0.022	2,060	0.05
	Temperate	4.55	0.320	281	0.36
North America	Boreal	3.37	0.238	279	0.36
	Temperate	3.08	0.089	682	0.15

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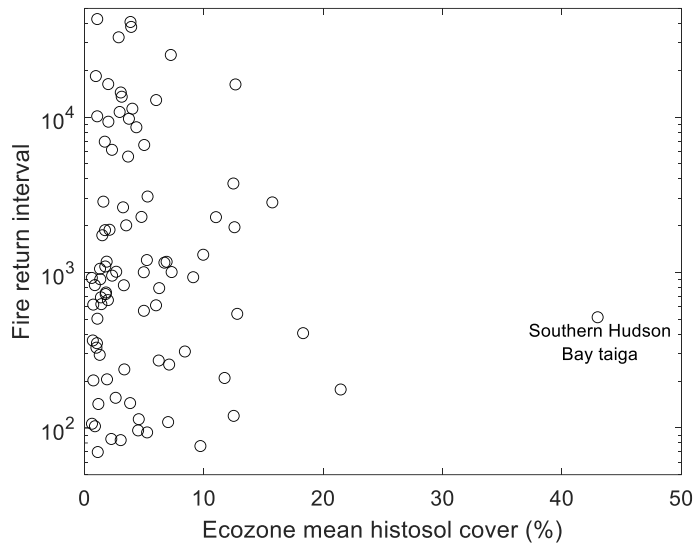
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Figure ED1. Conceptual diagram of the modelling design developed to incorporate peat carbon loss from wildfire (peat burn severity) and post-fire carbon dynamics (recovery rate and recovered net ecosystem exchange (NEE)) in peatland GHG emissions. Where  $y_1$  represents the  $NEE + CH_4$  of a burned peatland,  $x_1$  represents the time lag between wildfire and the initiation of post-fire recovery,  $x_2$  represents the time at which “recovered” NEE is achieved and  $y_2$  represents the magnitude of the recovered carbon sink. The variability in peat burn severity, time lag, recovery rate, and recovered NEE are depicted by the blue dashed lines and yellow arrows.

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503 Figure ED2. Fire return interval ( $100/(\text{burn rate})$ ) per ecozone, and mean ecozone histosol

504 cover. The Southern Hudson Bay taiga ecozone is highlighted as the region with the highest

505 histosol cover (~43%).