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# The role of kelp in the transport and fate of negatively buoyant marine plastic

Turner, Andrew

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1 **The role of kelp in the transport and fate of negatively buoyant marine**  
2 **plastic**

3

4 Andrew Turner\*<sup>1</sup> and Tracey Williams<sup>2</sup>

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6 <sup>1</sup>School of Geography, Earth and Environmental Sciences,

7 University of Plymouth

8 Drake Circus

9 Plymouth PL4 8AA, UK

10 aturner@plymouth.ac.uk

11

12 <sup>2</sup>The Lost At Sea Project

13 Old Bridge House

14 Porth Bean Road

15 Newquay TR7 3LU, UK

16

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## 22 **Abstract**

23 We report observations of negatively buoyant plastics washed up on the shores of a beach in  
24 southwest England in association with deposits of kelp. Items retrieved included polyethylene  
25 terephthalate (PET) bottle fragments, polyvinyl chloride toys, electrical casings and construction  
26 plastics, polycarbonate goggles, carcass tags, pot rubber, fishing float fragments and PET textiles,  
27 with an upper estimate of 36 tonnes of plastic deposited during a single kelp beaching event. It  
28 appears that after kelp becomes detached from the bed during strong winds or tides or through  
29 senescence, biological debris becomes associated with or acts to transport benthic plastic landwards  
30 through entrainment, entanglement and rafting and, sometimes, attachment to the holdfast.  
31 Observations provide evidence of significant local and regional ecological transportation of plastics  
32 in temperate coastal environments that is not driven directly or solely by buoyancy, and afford a  
33 convenient, non-invasive insight into the makeup of plastic that is encountered in the benthic  
34 environment.

35 **Keywords:** plastic litter; kelp; Laminariales; beaches; density; coastal zone

36

## 37 **1. Introduction**

38 Subtidal kelps are fast-growing seaweeds that dominate wave-exposed rocky and boulder shores in  
39 temperate and sub-polar regions (Steneck et al., 2002; Hannah and Cowie, 2009). Kelp beds and  
40 forests support high primary and secondary productivity, provide habitats for a diversity of  
41 organisms, including commercially important species, and influence light levels, water flow and  
42 sedimentation rates (Smale et al., 2013).

43 Detachment of kelp from the substrata can be caused by strong tides or intense wave activity  
44 (Krumhansl and Scheibling, 2012; López et al., 2019), with detritus subsequently depositing in local  
45 intertidal zones, accumulating in embayments or rafting farther afield (Krumhansl and Scheibling,  
46 2012; Waters and Craw, 2017). It has been estimated that about 2.5% of kelp biomass is cast on  
47 beaches where it is subject to fragmentation, dehydration, decomposition, consumption by  
48 detritivores and microbes and burial and, depending on tidal conditions, return offshore (Colombini  
49 et al., 2000; Colombini and Chelazzi, 2003).

50 Around the rocky reefs of the UK, subtidal kelps are dominated by brown macroalgae up to 3 m in  
51 length belonging to the Laminariales. These consist of a holdfast, stipe and divided or undivided  
52 blade, and include *Laminaria hyperborea*, *Laminaria digitata*, *Saccharina latissima*, *Sacchorhiza*  
53 *polyschides* and *Himanthalia elongate* (Hannah and Cowie, 2009; Burrows et al., 2014). Sublittoral

54 kelp beds and forests are encountered from the low water mark to depths of up to 35 m along more  
55 than 19,000 km of the coastline (Smale et al., 2013), with the precise makeup of communities  
56 dependent on a variety of factors, such as wave exposure, depth, turbidity, temperature and the  
57 abundance and type of grazers (Burrows et al., 2014).

58 In the UK, large accumulations of beach-cast kelp are sometimes considered as a nuisance or a  
59 health hazard because of the proliferation of kelp flies, the emission of hydrogen sulphide gas by  
60 anaerobic bacteria and the retention of faecal pathogens (Hannah and Cowie, 2009). An additional  
61 environmental concern with beach-cast kelp, however, is its association with litter. This has been  
62 mentioned in the literature (Colombini et al., 2003; Hannah and Cowie, 2009) but, thus far, has not  
63 undergone investigation or characterisation. The purpose of this study, therefore, was to examine  
64 the nature and type of plastic litter washed up with beach-cast biogenic detritus in order to  
65 understand the origin and transport of this material and whether its behaviour in the littoral zone  
66 can be linked directly to the life-cycle of kelp. For sampling and observations, we selected an  
67 accessible location in southwest England where significant kelp deposits containing visible items and  
68 fragments of plastic regularly appear.

69

## 70 **2. Methods**

### 71 *2.1. Study site*

72 Porth Beach (~ 200 m in length and ~ 500 m wide) is a west-facing sandy cove in the county of  
73 Cornwall, UK, that is backed by rocky cliffs to the north and south and an elevated road and  
74 embankment to the east (Figure 1). The hydrography and climatology of the region have been  
75 reviewed by Uncles and Stephens (2007). Briefly, climate is cool temperate and winds are  
76 predominantly westerly and south-westerly, with frequent gales and storms and a significant wave  
77 height exceeded for 10% of the year of between 2.5 and 3 m. Tidal currents are considerable (the  
78 maximum tidal range at Porth is 8 m) and are able to mix the water column, resuspend silt and  
79 mobilise sand and gravel, with net sand transport in a south-westerly direction.

80

### 81 *2.2. Observations and sampling*

82 New deposits of kelp at Porth Beach were visited from September 2017 to October 2019 on sixteen  
83 occasions, with the evolution and life-span of some deposits ascertained by regular, and often daily,  
84 re-visitations. The date, tidal range and the maximum wind speed and average wind direction

85 reported for the day on which each deposit was first observed are shown in Table 1. Also shown is  
86 the maximum daily wind speed and its average direction for the five days directly preceding each  
87 visit. Accumulations were accessed around low tide and plastic (including articles and fragments of  
88 synthetic rubber) visible to the naked eye at or just below the surface and at or near the full length  
89 of the seaward edge was retrieved and photographed. The type of polymer was noted where the  
90 resin identification code or any other relevant indicator was evident, and the buoyancy of selected  
91 washed-dried, whole objects or offcuts thereof in seawater at room temperature was noted in a  
92 clear container.

93

### 94 **3. Results**

95 Accumulations of kelp were variable in size and location in the intertidal zone but were most  
96 frequently encountered as mats of up to 400 m in length, several tens of m in width and up to 1 m in  
97 depth backed up along the northern, rocky boundary of the beach where incoming seawater meets  
98 freshwater issuing from a small stream (Figure 1). Detritus consisted of a heterogeneous assortment  
99 of the remains of whole individuals or components (blades, stipe or holdfast) of various *Laminaria*,  
100 with remnants of other intertidal and subtidal macroalgae, that were dominated by fucoids but with  
101 significant contributions from rhodophytes, and the soft coral, *Eunicella verrucosa*. The deposition of  
102 beach-cast kelp occurred in all seasons, over a range of tidal conditions and during or after winds  
103 originating from all directions and of variable maximum speeds (Table 1). Residence time on the  
104 beach ranged from a single tidal cycle for smaller accumulations to several days for larger deposits.

105 Figure 2 provides an indication of the relative abundance and variety of plastic that was typically  
106 visible at the surface of the kelp deposits. Overall, tens of thousands of pieces of plastic were  
107 retrieved from beach-cast kelp deposits for inspection, with over a thousand pieces sampled from  
108 single kelp accumulations that were visited on several occasions after successive tidal incursions had  
109 exposed or redistributed the litter. The size of plastic detected ranged from pot rubber strips up to 4  
110 m in length to clothing buttons of a few mm in diameter, although we note that textiles and drying,  
111 perishing rubber are likely to represent significant, indirect and localised sources of much smaller  
112 microplastics and microfibers. Most rigid materials were also usually smooth and rounded and some  
113 contained fouled calcareous casts of keel worms and skeletons of bryozoans.

114 Table 2 categorises plastics according to usage or appearance, along with specific examples of items  
115 that were commonly observed either as whole objects or distinctive fragments. Also shown are the  
116 most likely sources and types of polymer for each category. Sources identified are littering on land,

117 on beaches and at sea, spillages from ships transporting new products or material for recycling,  
118 abandonment of, for example, electronic and fishing gear at sea, accidental loss on beaches or at  
119 sea, and inputs of municipal waste to the marine environment via rivers or collapsing coastal landfill  
120 sites, and source attribution was based on the matrix scoring system given in Tudor and Williams  
121 (2004). Polymer type was ascertained from any relevant signage or was deduced from the materials  
122 most commonly employed in the manufacture of such items.

123 It was difficult to make quantitative assessments of plastic abundance or type because of the  
124 heterogeneity of litter, the tendency of smaller pieces to evade detection and settle through the  
125 deposits, entanglement of rope, line and rubber strips by kelp fragments, encapsulation of plastics  
126 below the surface by a thick matting of detritus, and specific interactions with kelp fragments (as  
127 described below). Nevertheless, based on material retrieved, the most common category of plastic  
128 on a number basis (and exceeding 50% of samples collected or visible in most cases) was food and  
129 drink, which was dominated by fragments of clear (or less frequently, coloured) polyethylene  
130 terephthalate (PET) bottles. The seaward edge of one large accumulation of beach-cast kelp that was  
131 re-visited on several occasions yielded over 1000 bottle pieces. Remains ranged from small pieces of  
132 the body or bottom to whole bottles with multiple punctures, and many fragments that included the  
133 shoulder and neck had the polyethylene cap and/or collar still attached and intact (Figure 3a). Also  
134 commonly encountered were plastics associated with fishing activities, and in particular fragments  
135 of colourful, impact-modified, moulded trawl, gill and net floats that appeared to be constructed of  
136 polystyrene or acrylonitrile butadiene styrene (ABS) (Figure 3b) and strips of rubber used on lobster  
137 pots (Figure 3c). Significant among remaining samples that were identifiable were swimming goggles  
138 and masks that are typically constructed of clear polycarbonate (Figure 3d), footwear, including  
139 beach shoes and remains of diving flippers made from polyester or rubber (Figure 3e), items or  
140 remains of polyester-based clothing (including beachwear; Figure 3f) and ABS- or polyvinyl chloride-  
141 (PVC-) based toys and sports equipment.

142 Distinctly lacking from our surveys, however, were low-density foamed plastics or polyolefin-based  
143 samples that are commonly found deposited around the high water strandline in the absence of  
144 beach-cast kelp. These include fragments of expanded and extruded polystyrene, and polyethylene  
145 and polypropylene-based pre-production pellets, pens, fishing beads, sand moulds, biobeads, rawl  
146 plugs, screw caps, cotton buds, straws, crate strapping, containers, cartridges, nozzles, bottle tops  
147 and packaging (Turner and Solman, 2016; Fok et al., 2017).

148

#### 149 **4. Discussion**

150 The plastic samples retrieved from beach-cast kelp are a heterogeneous assortment of fragments  
151 and articles whose most important sources appear to be related to offshore fishing and littering and  
152 accidental loss, both at sea and onshore, with a smaller contribution arising from municipal waste.  
153 Samples also display a range of ages, with online searches of branded sunglasses, toothbrushes and  
154 toys and documented, regional spillages of cargo revealing a manufacturing period spanning more  
155 than sixty years. Significantly, however, from a physical perspective and as ascertained from visible  
156 indicators (including resin codes and calcareous fouling), typical polymer applications and empirical  
157 observations, the majority of samples appear to be constructed of plastics (and synthetic rubbers)  
158 that are denser than seawater. Exceptions here included polyolefin-based fishing ropes and beads  
159 visibly entangled with the biomass and low density plastics combined with denser materials as  
160 composites (e.g. polyethylene-aluminium laminate drink pouches).

161 In the marine environment, plastics whose inherent density,  $\rho$ , is lower than seawater ( $\rho \sim 1.025 \text{ g}$   
162  $\text{cm}^{-3}$  for local coastal water) are predicted to exist in suspension or become beached along the  
163 strandline of the littoral zone, and include polyethylene and polypropylene ( $\rho = 0.9\text{-}1.0 \text{ g cm}^{-3}$ ) and  
164 foamed polystyrene ( $\rho < 0.1 \text{ g cm}^{-3}$ ). In contrast, plastics whose density is greater than seawater,  
165 including polyester-PET ( $\rho \sim 1.4 \text{ g cm}^{-3}$ ), polycarbonate ( $\rho \sim 1.2 \text{ g cm}^{-3}$ ), PVC ( $\rho \sim 1.4 \text{ g cm}^{-3}$ ) and  
166 polyamide ( $\rho \sim 1.2 \text{ g cm}^{-3}$ ) are predicted to sink and settle on the seafloor. This simple concept of  
167 fractionation is generally borne out by observations of the presence and distribution of different  
168 types of plastics in marine settings (Munari et al., 2017; Erni-Cassola, 2019; Schonlau et al., 2020)  
169 and is a key component of plastic modelling in the coastal zone (Critchell and Lambrechts, 2016;  
170 Collins and Hermes, 2019). Contrary to these studies, however, our observations suggest that  
171 negatively buoyant plastics have been transported from the sub-littoral zone throughout a depth  
172 range supporting kelp beds and forests (up to 35 m; Smale et al., 2016) to the intertidal zone.

173 An upper-estimate of the mass of negatively buoyant plastic beached by a single casting of kelp at  
174 the location under study can be gained for a 400 m long, 15 m wide and 0.5 m deep deposit, similar  
175 to that shown in Figure 2. Thus, the volume of this deposit of  $3000 \text{ m}^3$  is equivalent to a mass of  
176 720,000 kg of material if a porosity (or air space) of 80% and a density of  $1.2 \text{ g cm}^{-3}$  are assumed. A  
177 relative abundance of plastic in the casting of 5% therefore yields a mass of plastic deposited of  
178 about 36,000 kg. This is comparable with remotely sensed estimates of low density (e.g. polystyrene  
179 foam and polyethylene) plastic debris on southeast Pacific beaches of a similar size to Porth that are  
180 known to act as accumulators of more buoyant oceanic debris (up to about 20,000 kg; Acuña-Ruz et  
181 al., 2018). Although our estimate is subject to uncertainties and temporal variation, it nevertheless  
182 illustrates that significant quantities of relatively dense plastic can be deposited on beaches in  
183 association with biogenic debris.

184 What is unclear is how such plastics become beach-cast with kelp, and whether kelp itself plays a  
185 direct role in the onshore transportation of plastic. Thus, it is possible that kelp and plastic are  
186 associated because both components are carried onshore with winds or currents of a suitably  
187 favourable speed or direction, with a pre-requisite of preceding conditions that are sufficiently  
188 turbulent to defoliate kelp and detach macroalgae from the substrate. In the current context,  
189 however, it is difficult to define the precise hydrographic or meteorological conditions from the data  
190 provided in Table 1. The transport of plastic could be facilitated by kelp that is either positively  
191 buoyant or negatively buoyant if it becomes entangled or entrained amongst the fragments, fronds  
192 and stipes of detached individuals. In some large species of kelp, the blades have greater buoyancy  
193 than the holdfasts because of the presence of gas bubbles or gas-filled pneumatocysts in the former  
194 (Thiel and Gutow, 2005) and, once detached, individuals could also act to increase drag on the  
195 seafloor and 'sweep' plastic items as bedload in the direction of travel. This process is akin to the  
196 physical abrasion of the understory of communities of *L. digitata* that is engendered by the long  
197 flexible stipe and finger-like blades of the alga (Burrows et al., 2014).

198 By analogy, a recent study found that fibres of the seagrass, *Posidonia oceanica*, intertwined into  
199 balls (aegagropilae) and recovered from Mediterranean beaches had up to about 1500 (micro-)  
200 plastic items per kg of plant material and that the polymers identified were mainly negatively  
201 buoyant (including PET, polyamide and PVC; Sanchez-Vidal et al., 2021). It was suggested that  
202 shallow, coastal seagrass meadows are able to trap sinking plastics and subsequently package them  
203 into fibrous aggregates during the erosion of leaf sheafs, with resulting aegagropilae transported  
204 onshore during stormy conditions. As with kelp, seagrass appears to provide an ecosystem service by  
205 buffering, trapping and redistributing high density plastics in the coastal environment.

206 Our observations also reveal that kelp and plastic may interact with each other in situ more actively  
207 and persistently than physical intertwining and entrainment. For example, Figure 4 shows the  
208 threaded neck of a green-dyed PET bottle around the stipe of *L. digitata* and a golf ball (consisting of  
209 a thermoplastic cover over a solid rubber core) attached to the rhizoids of the holdfast of *L.*  
210 *hyperborea*. Experiments aimed at studying hatchery-reared macroalgal juveniles have  
211 demonstrated that certain plastics are suitable substrates for bioadhesion (Kerrison et al., 2019) but  
212 the latter of our observations appear to be the first to document thigmotactic attachment in the  
213 environment. The implications of this association are that kelp may be more prone to become  
214 detached or dislodged than when bound to a hard, rocky substrate, or, conversely, kelp may act to  
215 stabilise loosely deposited plastic on the seafloor.

216 Beach-cast kelp will eventually be removed from the intertidal zone through a variety of processes.  
217 For example, biomass will gradually disappear through degradation, consumption and, if buried by  
218 sand, anaerobic decomposition. Accumulations in some cases may also be disposed of by  
219 intervention (e.g. beach cleaning), but if deposited on or around a neap tide, spring tides are likely to  
220 resuspend unprocessed material in the water column or at the bed and return it offshore (Hannah  
221 and Cowie, 2009; Krumhansl and Scheibling, 2012). The rather short residence times of kelp on Porth  
222 beach (~ days) and lack of systematic intervention suggest that hydrodynamic processes are likely to  
223 be the most significant removal mechanism in the current study.

224 The subsequent transport and fate of plastic washed up with kelp may or may not be associated with  
225 that of the biomass. Thus, some smaller plastics could undergo burial with fragments of kelp while  
226 items becoming detached from assemblages may be dispersed elsewhere. For relatively dense kelps  
227 or components thereof (e.g. stipes), encapsulated plastic may retreat offshore with the biomass,  
228 while for relatively low density kelps or components thereof (e.g. blades), kelp and associated plastic  
229 may be resuspended by tidal action and rafted offshore. The latter mechanism affords a means by  
230 which plastic may be biologically dispersed and deposited far further from its point of origin than  
231 would be predicted from physical (i.e. density) considerations alone. By analogy, rafts of bull kelp,  
232 *Durvillaea antarctica*, retrieved from beaches of south-eastern New Zealand revealed geogenic  
233 debris on many holdfasts that had travelled more than 100 km in the Southern Ocean (Waters and  
234 Craw, 2017). Since deposition of kelp on beaches represents less than half of the amount that is  
235 eroded and detached at sea (Colombini and Chelazzi, 2003), rafting without beaching (for instance,  
236 during offshore winds) could represent a more significant, general dispersal mechanism of dense  
237 plastic.

238 In summary, negatively buoyant plastics, like bottle fragments, textiles, fishing gear and electronic  
239 housings, exhibit a physical association with beach-cast kelp in the coastal zone of rocky, temperate  
240 environments, highlighting an additional means by which coastal plants are able to interact with  
241 marine plastics (Poeta et al., 2017; Battisti et al., 2020; Sanchez-Vidal et al., 2021). Clearly, modelling  
242 of plastic behaviour in the coastal setting should not rely on density considerations alone but factor  
243 in onshore ecologically-assisted transport and dispersal of bedload. Beach-cast deposits of kelp also  
244 provide a convenient and accessible medium in which to study this type of plastic litter without  
245 incurring the costs and difficulties associated with in situ benthic surveys.

246

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250

251

## 252 **References**

253 Acuña-Ruz, T., Uribe, D., Taylor, R., Amezcuita, L., Guzman, M.C., Merrill, J., Martinez, P., Voisin, L.,  
254 Mattar, C.M., 2018. Anthropogenic marine debris over beaches: Spectral characterization for remote  
255 sensing applications. *Remote Sensing of Environment* 217, 309-322.

256 Battisti, C., Fanelli, G., Filpa, A., Cerfolli, F., 2020. Giant reed (*Arundo donax*) wrack as sink for plastic  
257 beach litter: First evidence and implication. *Marine Pollution Bulletin* 155, 111179.

258 Burrows, M.T., Smale, D., O'Connor, N., Van Rein, H., Moore, P., 2014. Marine Strategy Framework  
259 Directive Indicators for UK Kelp Habitats. Part 1: Developing proposals for potential indicators. Joint  
260 Nature Conservancy Committee, Peterborough, 80pp.

261 Collins, C., Hermes, J.C., 2019. Modelling the accumulation and transport of floating marine micro-  
262 plastics around South Africa. *Marine Pollution Bulletin* 139, 46-58.

263 Colombini, I. Chelazzi, L., 2003. Influence of marine allochthonous input on sandy beach  
264 communities. *Oceanography and Marine Biology: An Annual review* 41, 115-159.

265 Colombini, I., Aloia, A., Fallaci, M., Pezzoli, G., Chelazzi, L. 2000. Temporal and spatial use of stranded  
266 wrack by the macrofauna of a tropical sandy beach. *Marine Biology* 136, 531–541.

267 Critchell, K., Lambrechts, J., 2016. Modelling accumulation of marine plastics in the coastal zone;  
268 what are the dominant physical processes? *Estuarine, Coastal and Shelf Science* 171, 111-122.

269 Erni-Cassola, G., Zadjelovic, V., Gibson, M.I., Christie-Oleza, J.A., 2019. Distribution of plastic polymer  
270 types in the marine environment; A meta-analysis. *Journal of Hazardous Materials* 369, 691-698.

271 Fok, L., Cheung, P.K., Tang, G.D., Li, W.C., 2017. Size distribution of stranded small plastic debris on  
272 the coast of Guangdong, South China. *Environmental Pollution* 220, 407-412.

273 Hannah, F., Cowie, P.R., 2009. The potential risks to human health posed by living, attached  
274 seaweeds and dead, beachcast material associated with sandy beaches: A preliminary report.  
275 Environment Agency, Bristol, 54pp.

276 Kerrison, P.D., Stanley, M.S., De Smet, D., Buyle, G., Hughes, A.D., 2019. Holding (not so) fast: Surface  
277 chemistry constrains kelp bioadhesion. *European Journal of Phycology* 54, 291-299.

278 Krumhansl, K.A., Scheibling, R.E., 2012. Production and fate of kelp detritus. *Marine Ecology Progress*  
279 *Series* 467, 281-302.

280 López, B.A., Macaya, E.C., Jeidres, R., Valdivia, N., Bonta, C.C., Tala, F., Thiel, M., 2019. Spatio-  
281 temporal variability of strandings of the southern bull kelp *Durvillaea antarctica* (Fucales,  
282 Phaeophyceae) on beaches along the coast of Chile—linked to local storms. *Journal of Applied*  
283 *Phycology* 31, 2159-2173.

284 Munari, C., Scoponi, M., Mistri, M., 2017. Plastic debris in the Mediterranean Sea: Types, occurrence  
285 and distribution along Adriatic shorelines. *Waste Management* 67, 385-391.

286 Poeta, G., Fanelli, G., Pietrelli, L., Acosta, A.T.R., Battisti, C., 2017. Plasticsphere in action: evidence for  
287 an interaction between expanded polystyrene and dunal plants. *Environmental Science and*  
288 *Pollution Research* 24, 11856-11859.

289 Ryan, P.G., 2015. Does size and buoyancy affect the long-distance transport of floating debris?  
290 *Environmental Research Letters* 10, 084019, DOI: 10.1088/1748-9326/10/8/084019

291 Sanchez-Vidal, A., Canals, M., de Haan, W.P., Romero, J., Veny, M., 2021. Seagrasses provide a novel  
292 ecosystem service by trapping marine plastics. *Scientific Reports* 11, 254.

293 Schonlau, C., Karlsson, T.M., Rotander, A., Nilsson, H., Engwall, M., van Bavel, B., Karrman, A., 2020.  
294 Microplastics in sea-surface waters surrounding Sweden sampled by manta trawl and in-situ pump.  
295 *Marine Pollution Bulletin* 153, 111019.

296 Schwarz, A.E., Lighthart, T.N., Boukris, E., van Harmelen, T., 2019. Sources, transport, and  
297 accumulation of different types of plastic litter in aquatic environments: A review study. *Marine*  
298 *Pollution Bulletin* 143, 92-100.

299 Smale, D.A., Burrows, M.T., Moore, P., O'Connor, N., Hawkins, S.J., 2013. Threats and knowledge  
300 gaps for ecosystem services provided by kelp forests: a northeast Atlantic perspective. *Ecology and*  
301 *Evolution* 3, 4016-4038.

302 Smale, D.A., Burrows, M.T., Evans, A.J., King, N., Sayer, M.D.J., Yunnice, A.L.E., Moore, P.J., 2016.  
303 Modelling accumulation of marine plastics in the coastal zone; what are the dominant physical  
304 processes? *Marine Ecology Progress Series* 542, 79-95.

305 Steneck, R.S., Graham, M.H., Bourque, B.J., Corbett, D., Erlandson, J.M., Estes, J.A., Tegner, M.J.,  
306 2002 Kelp forest ecosystems: Biodiversity, stability, resilience and future. *Environmental*  
307 *Conservation* 29, 436-45.

308 Thiel, M., Gutow, L., 2005. The ecology of rafting in the marine environment. I. The floating  
309 substrata. *Oceanography and Marine Biology: An Annual Review* 42, 181-264.

310 Tudor, D.T., Williams, A.T., 2004. Development of a 'Matrix Scoring Technique' to determine litter  
311 sources at a Bristol Channel beach. *Journal of Coastal Conservation* 9, 119-127.

312 Turner, A., Solman, K.R., 2016. Analysis of the elemental composition of marine litter by field-  
313 portable-XRF. *Talanta* 159, 262-271.

314 Uncles, R.J., Stephens, J.A., 2007. SEA 8 Technical Report - Hydrography. PML Applications,  
315 Plymouth, UK, 105pp.

316 Waters, J.M., Craw, D., 2017. Large kelp-rafted rocks as potential dropstones in the Southern Ocean.  
317 *Marine Geology* 391, 13-19.

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319

Table 1: Dates on which beach-cast kelp deposits were first observed, along with the tidal range (from the first low water to first high water), maximum wind speed ( $v_{\max}$ ) and average wind direction. Also shown are the maximum daily wind speed reported for the five days previous to the date shown ( $v_{\max}$  5-day) along with the average direction of this wind.

accumulation	date	tidal range, m	$v_{\max}$ , mph	direction	$v_{\max}$ 5-day, mph	direction
1	28/09/2017	2.6	16	139	18.4	116
2	10/10/2017	5.8	16	232	19.6	296
3	27/02/2018	4.9	13	60	18	99
4	11/03/2018	2.1	17	101	22	261
5	27/03/2018	3.9	22	272	16	205
6	08/04/2018	2.5	9	19	21	306
7	16/09/2018	3.7	13	191	15	358
8	12/12/2018	4.5	16	124	38	268
9	26/12/2018	6.4	8	118	23	255
10	25/01/2019	6.6	15	292	26	328
11	10/02/2019	4.9	30	291	33	208
12	17/04/2019	6.1	16	102	29	117
13	30/05/2019	3.8	14	246	21	286
14	31/07/2019	5.9	15	270	28	306
15	26/08/2019	3.1	10	333	16	123
16	22/10/2019	2.9	8	126	21	245

Table 2: Categorisation, sources, examples and polymeric construction of plastics retrieved from beach-cast kelp deposits at Porth, Cornwall. Sources: Ab = abandoned; Li = litter; Lo = lost; Mu = municipal; Sp = spillage. Polymers: ABS = acrylonitrile butadiene styrene; CA = cellulose acetate; PA = polyamide; PC = polycarbonate; PE = polyethylene; PET = polyethylene terephthalate (polyester); PS = polystyrene; PMMA = Poly(methyl methacrylate); PVC = polyvinyl chloride; R = synthetic rubber.

Category	Sources	Examples	Polymers
Plumbing and construction	Mu,Sp	pipes, adaptors, hosing, seals, gaskets, nuts, bolts, screws, cable ties, hooks	PA, PC, PVC, R
Textiles	Ab,Lo,Sp	clothing, swimwear, carpet offcuts, parasols, bodyboard covers	PA, PET
Toys	Lo,Mu,Sp	figures, whistles, party toys, inflatables, Lego, dog toys	ABS, PVC
Maritime sports	Lo	surf board screws, diving fins, goggles, masks	PC, R
Electronic	Sp,Lo	cable-wire insulation, plug sockets, radio and computer casings, keyboard keys	ABS, PS
Food and drink	Li,Lo	water bottles, drink pouches, single-use cutlery, snack spreaders, mugs and plates, jar seals, ice cream spoons	PE, PET, PS
Cosmetic and health	Lo,Mu	combs, dentures, toothbrush heads, razor handles, spectacles, sunglasses, medical lancets	CA, PA, PMM, PS
Fishing	Ab,Li,Lo	carcass tags, rope, fishing line, fishing float fragments, pot rubber, gloves, laminated card	ABS, PA, PE, PET, PS, R
Eyewear, footwear, accessories	Lo	beach shoes, spectacles, costume jewellery, sunglasses, zips,	ABS, PC, PET, PMMA
Unidentifiable objects and fragments	Ab,Li,Lo,Mu,Sp	buttons, hair-bands and ties, buckles	various

Figure 1: Porth Beach, showing local bathymetry and approximate locations of beach-cast kelp.

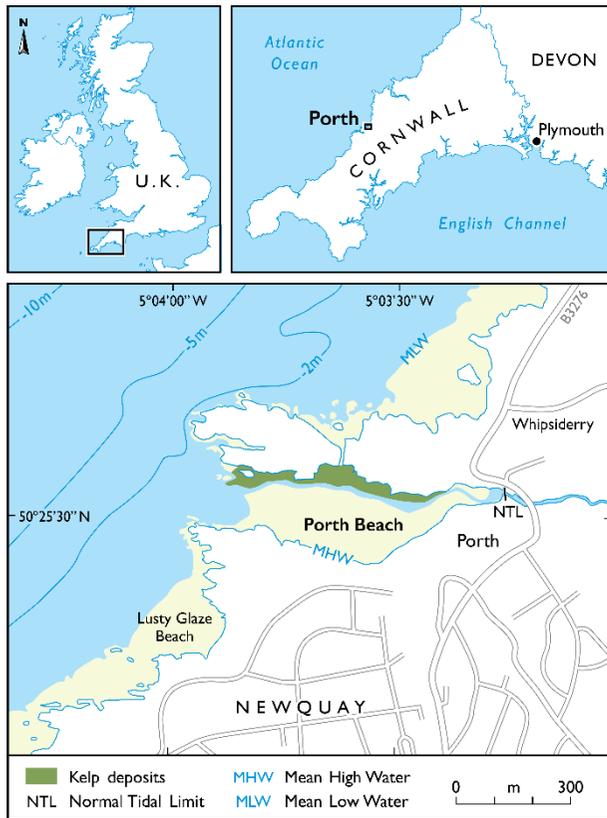


Figure 2: Beach-cast kelp on the northern edge of Porth Beach.



Figure 3: A selection of plastic articles and fragments retrieved from beach-cast kelp on Porth Beach. (a) PET bottle fragments, (b) fishing float fragments, some of which have visible encrustations of keel worm casts; (c) pot rubber strips; (d) goggles and masks; (e) footwear; (f) clothing.



Figure 4: The neck of a PET bottle around the lower stipe of *L. hyperborea* and a golf ball attached to the holdfast of *L. digitata*.

