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# Microplastics in different tissues of fish and prawn from the Musa Estuary, Persian Gulf.

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## 14 **Highlights**

- 15 • Microplastics (MPs) have been determined in tissues of fish and crustaceans from the Musa  
16 estuary and Persian Gulf
- 17 • 828 MPs of mainly a fibrous nature were detected in all tissues and species examined
- 18 • Mean abundance ranged from 7.8 in tiger prawn to 21.8 in bartail flathead
- 19 • The means by which MPs enter non-digestive tissues is unclear
- 20 • The occurrence of MPs in seafood for human consumption is cause for concern

21

## 22 **Abstract**

23 Commercially-important species of fish and a crustacean from four sites in the Musa estuary and  
24 a site in the Persian Gulf have been analysed for the presence and location of microplastics (MPs).  
25 A total of 828 MPs were detected in the guts (gastrointestinal tracts), skin, muscle, gills and liver  
26 of demersal and pelagic fish (*Platycephalus indicus*, *Saurida tumbil*, *Sillago sihama*, *Cynoglossus*  
27 *abbreviatus*) from all five sites and in the exoskeleton and muscle of the tiger prawn, *Penaeus*  
28 *semisulcatus*, from three sites. On an individual basis, MPs were most abundant in *P. indicus*  
29 (mean = 21.8) and least frequently encountered in *P. semisulcatus* (mean = 7.8), but when  
30 normalized on a mass basis, MPs ranged from 0.16 g<sup>-1</sup> for *C. abbreviatus* to 1.5 g<sup>-1</sup> for *P.*  
31 *semisulcatus*. Microscopic analyses (polarized light, fluorescence, SEM/EDS) revealed that MPs  
32 were mainly fibrous fragments (with a few angular fragments) of various colour and size (< 100  
33 µm to > 1000 µm) and with strong C and O signatures. Additional particles detected that were  
34 distinctly different in colour, morphology, brittleness and elemental composition (part-metallic,  
35 and containing Cu) were suspected of being fragments of antifouling paint. The means of entry of

36 MPs into tissues not involved in digestion are unclear but could be related to translocation or  
37 adherence. Regardless of the mode of accumulation, the presence of MPs in heavily fished species  
38 of fish and crustacean raises concerns about the potential transfer of synthetic materials into  
39 humans.

40

41 **Keywords:** Microplastics; fish; prawns; accumulation; microscopy; Persian Gulf

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43

## 44 **1. Introduction**

45 While the effects of chemical pollutants on marine ecosystems have been studied for many  
46 decades, the pervasiveness and impacts of litter on marine life have been recognized more recently  
47 (Auta et al., 2017; do Sul and Costa, 2014). Plastics, as identifiable primary objects or secondary  
48 fragmented pieces, comprise the largest pool of litter on both a mass and number basis and enter  
49 the oceans via rivers, sewage discharge, land run-off, and spillages and discharges from ships at  
50 sea (Moore, 2008; Andrady, 2011; Barnes et al., 2009; Gregory and Andrady, 2003). Given the  
51 expected future demand and discharges of plastic, coupled with the resistance of synthetic  
52 polymers to environmental degradation, it is clear that the marine plastic inventory will continue  
53 to increase beyond at least the next decade (Jambeck et al., 2015).

54

55 Of particular concern with respect to both the direct impacts on marine life and transfer through  
56 the foodchain are readily ingestible microplastics (MPs), or synthetic particles ranging from a few  
57 micrometers to five millimeters in any dimension (Alomar et al., 2016; Turner, 2017; Abbasi et

58 al., 2017). Primary MPs include abrasive micro-beads in face scrubber cosmetics and toothpaste,  
59 synthetic fibres and pre-production resin pellets, while secondary MPs are generated in situ by the  
60 mechanical and oxidative breakdown of larger plastics (Hidago-Ruz et al., 2012). As well as the  
61 inherent composition of the polymer and the presence of any additives, the chemistry of MPs may  
62 be modified by the adsorption of toxic substances from ambient sea water to the hydrophobic  
63 plastic surface (e.g. organic pollutants; Teuten et al. 2007; Ziccardi et al. 2016) or to more  
64 hydrophilic, hydrogenous or biogenic phases coating the surface (e.g. heavy metals; Ashton et al.,  
65 2010; Holmes et al., 2014).

66

67 A wide range of marine organisms, including bivalves, zooplankton, fish, invertebrates, birds and  
68 cetaceans, incidentally take up MPs from sediment or the water column because they mistake them  
69 for food (Cole et al. 2013; Lusher et al. 2015; Ferreira et al. 2016). Ingesting MPs of no nutritional  
70 value may induce physical and chemical toxicity, block or damage the digestive tract, or decrease  
71 individual fitness, ultimately resulting in death (Wright et al., 2013; Luís et al. 2015; De Sá and  
72 Guilhermino 2015). Moreover, MPs of tens of micrometers in dimension have the propensity to  
73 translocate from the gut to the circulatory system in many organisms where they may reside for  
74 relatively long periods of time (Browne et al., 2008; van Cauwenberghe et al., 2015; Collard et al.,  
75 2017). While the effects of translocated MPs on chronic animal health are unknown, their presence  
76 is of particular concern because consumption of contaminated food, including fish and shellfish,  
77 may act as a vehicle for the ingestion and translocation of MPs in humans (Li et al., 2015; Rist et  
78 al., 2018).

79

80 The Musa is one of the biggest estuaries in the northern Persian Gulf and is the most important  
81 fishery resource for people in the cities of Mashahr (population 150,000), Sarbandar (75,000) and  
82 Hendijan (50,000). While the coast of the estuary is flanked by agricultural land, there are also  
83 various industrial plants and extensive docks that support the petrochemical and shipping  
84 industries and municipal and industrial sewage from the catchment is poorly treated (Hosseini et  
85 al., 2013; Rastegari Mehr et al., 2016). Given these conflicting uses of the estuary, the aim of the  
86 present study was to determine whether MPs are accumulating in different organs of five abundant  
87 and commercially valuable species of fish and crustacean that are heavily consumed by local  
88 people. Specifically, we target the skin, gastrointestinal tract, liver, muscle and gills of two  
89 demersal fish, the bartail flathead (*Platycephalus indicus*) and greater lizardfish (*Saurida tumbil*),  
90 one pelagic fish, the northern whiting (*Sillago sihama*), and one mesopelagic species fish, the  
91 tongue sole (*Cynoglossus abbreviatus*), and the skin and muscle of the tiger prawn, *Penaeus*  
92 *semisulcatus*.

93

## 94 **2. Materials and methods**

### 95 **2.1. Sampling and sample preparation**

96 Fish and prawn samples were caught from along the coastal waters of the Persian Gulf during June  
97 2015 by a trawl net from five locations (see Figure 1), one of which served as a control site (S5;  
98 the fishery port of Hendijan located outside the estuary and 70 km from any petrochemical  
99 facilities). At each station, up to five samples of each species were collected, with a total catch of  
100 56 specimens among all species. Samples were transported in a cooler to the laboratory where they  
101 were stored at -20 °C pending processing and analysis.

102

103 Given the ubiquity of MPs in the indoor environment (Gasperi et al., 2018), suitable measures  
104 were undertaken to prevent plastic and fibre contamination in the laboratory. Thus, all chemical  
105 reagents were filtered (8- $\mu$ m, Whatman No. 540) before use and white cotton laboratory coats,  
106 single-use latex gloves and face masks were used throughout sample manipulation and processing.  
107 Working surfaces were thoroughly cleaned with ethanol and all glassware, tools and fish and  
108 prawn skin surfaces were washed successively with a commercial dishwashing liquid, HPLC-  
109 grade distilled water and ethanol before being dried in an oven at 105 °C (glassware and tools) or  
110 at room temperature in a metal cabinet (skin surfaces). Analysis of two procedural blanks (without  
111 tissues) and distilled water contained in two wide dishes that had been left exposed during the  
112 duration of sample processing revealed no contamination from MPs under the working conditions  
113 in the laboratory.

114

## 115 **2.2. Extraction of MPs**

116 As required, specimens were thawed and the fork length from the mouth to the central point of the  
117 caudal fin and body weight were recorded. Each fish was gutted and dissected in a metal tray using  
118 a scalpel, forceps and scissors and the muscle, skin, gills, liver and gut (gastrointestinal tract)  
119 retrieved. The pooled livers, guts and gills from each species and at each site were transferred  
120 directly to covered petri dishes while pooled muscles and skin, after separation, were homogenized  
121 using an Electric Meat Grinder (KENWOOD MG510, UK) before about 15 g of each was retained  
122 and stored in a petri dish. For the (smaller) prawns, tissue retained for analysis was restricted to  
123 the muscle and skin (exoskeleton) that was pooled from individuals and homogenized as above.

124

125 Tissues of fish and prawn were subject to digestion to remove organic matter and leave behind  
126 silica/aluminosilicates and any plastic (Karami et al., 2017). Thus, tissues were emptied into a  
127 series of 500 mL glass beakers to which approximately 30 mL of 35% H<sub>2</sub>O<sub>2</sub> (Arman Sina) and 30  
128 mL of 4% KOH (Merck) were added. The contents were digested for 72 h at 60 °C in an oven to  
129 dissolve the soft organic components of the tissues, before a 10:40 ml mixture of 68% HClO<sub>4</sub> and  
130 65% HNO<sub>3</sub> (both Merck) was added to completely digest more resistant material like the gills and  
131 skin-exoskeleton. After a few minutes of acid extraction, digests were diluted with warm distilled  
132 water to preserve the integrity of MPs. Plastics were separated from all tissues with the exception  
133 of the gut by shaking digests at 350 rpm for 5 min and subsequently centrifuging triplicate aliquots  
134 for 5 min at 4000 rpm. Supernatants were directly filtered under vacuum through S & S grade  
135 589/3 filters which were subsequently stored and dried (at room temperature) in individual petri  
136 dishes pending analysis.

137  
138 For MPs embedded in the gastrointestinal tract of fish, remaining digests were agitated at 350 rpm  
139 for 5 min in a solution of concentrated sodium iodide (NaI, Merck; density = 1.6-1.8 g cm<sup>-3</sup>) to  
140 separate plastics from additional material that had been ingested with subsequent filtration and  
141 storage undertaken as above.

### 142 143 **2.3. Observation and validation of MPs**

144 A visual assessment of material retained on the filters, and including any arising from the  
145 procedural control, was made according to colour, size and morphology (elongated fibre versus  
146 angular fragment) and at up to 200 x magnification using a Carl-Zeiss binocular microscope. The  
147 presence of plastic was verified by the colours returned by polarized light microscopy using an

148 Olympus BX41TF microscope and by fluorescence microscopy using an Olympus CX31  
149 microscope. Images from all microscopic techniques were captured using an Olympus Pen EPL 1  
150 digital camera.

151  
152 Based on the optical microscopy results, the topography and elemental composition of selected  
153 MPs were determined through high vacuum SEM/EDS. We used a Tescan VEGA 3 electron  
154 microscope (with a resolution of 2 nm at 20 kV) and an Oxford Instruments X-Max 50 silicon drift  
155 detector with AZtec and INCA software after samples that had been carefully brushed from the  
156 filters were mounted on double-sided adhesive carbon tabs on aluminium SEM stubs.

157

### 158 **3. Results**

#### 159 **3.1. Size and weight of fish and prawns**

160 Table 1 summarises the catch from each sampling site (note that the number of species caught at  
161 each site varied and that some species were absent from sites 1, 2 and 3). Also shown are the mean,  
162 minimum and maximum lengths and weights of each of the five species, serving to illustrate  
163 differences in size among species and between sites and, for a given species, differences in age  
164 and, therefore, propensity to have accumulated MPs.

165

#### 166 **3.2. MPs in fish and prawns**

167 Table 2 shows the number of MPs in the tissues of the five species at each site, with data pooled  
168 for the number of individuals indicated in Table 1. Note that MPs were detected visually (Figure  
169 2), with the synthetic nature of samples confirmed by fluorescence and polarized light

170 microscopies for characteristic response to visible and ultraviolet light (Woodall et al., 2015; Wang  
171 et al., 2016; Figure 3) and, for selected samples, by SEM/EDS for surface morphology and  
172 elemental composition (mainly carbon). By comparison, no particles of this nature were observed  
173 on the two filters arising from the procedural controls.

174  
175 Among the catch, 828 pieces of MP were detected, being encountered across all tissues from each  
176 species. In only isolated cases (e.g. the liver of *P. indicus* from sites 1, 2 and 5 and the gut of *P.*  
177 *indicus* at site 5) were MPs absent, with numbers exceeding 25 in the skin of *S. sihama* at site 2,  
178 the gills of *P. indicus* at site 4 and the skin of *P. indicus* at site 5. On this basis, there were no clear  
179 differences in the total number of MPs accumulated by each species or between sites (and  
180 including the control site), but numbers tended to be higher in the skin, muscle and gills than the  
181 gut and liver of *S. sihama* and *P. indicus* and were always greater in the skin than in muscle from  
182 *P. semisulcatus*. When considered on an individual basis, or after total numbers for each species  
183 had been normalized for the number of samples analysed, MPs are most abundant in *P. indicus*  
184 (mean = 21.8) and least frequently encountered in *P. semisulcatus* (mean = 7.8); when normalized  
185 on a mass basis, however, the mean abundance of MPs ranged from 0.16 g<sup>-1</sup> for *C. abbreviatus* to  
186 1.5 g<sup>-1</sup> for *P. semisulcatus*. By comparison, a recent study by Akhbarizadeh et al. (2018) in the  
187 northeast of the Persian Gulf reports an average abundance of MPs in muscle of the fish, *P. indicus*,  
188 *Sphyraena jello* and *Epinephelus voioides*, and the shrimp, *Alepes djedaba*, of 1.85 ± 0.46, 0.57 ±  
189 0.17, 0.78 ± 0.22 and 0.80 ± 0.12 g<sup>-1</sup>, respectively.

190  
191 Nearly all MPs encountered were filamentous fragments (consisting of single fibres) of different  
192 size and colour and as illustrated in Figure 2. In only five cases were non-fibrous plastics found

193 among the different species of fish: specifically, two white fragments in the muscle of *C.*  
194 *abbreviatus* from site 5, one yellow fragment in the gills of *S. sihama* from site 4 and one blue  
195 fragment in the gastrointestinal tract of both *S. sihama* at site 4 and *S. tumbil* at site 1.

196  
197 The size distributions of MPs are shown in Table 3 for individual tissues and in Figure 4 for whole  
198 organisms. Thus, there is a wide range of lengths of (mainly) filamentous material across all  
199 species, with the most abundant sizes between either 100 and 250  $\mu\text{m}$  (*S. sihama*, *P. indicus*, *P.*  
200 *semisulcatus*) or 250 to 500  $\mu\text{m}$  (*C. abbreviatus*, *S. tumbil*). With respect to the different tissue  
201 types, the digestive organs appear to contain a high proportion of relatively large MPs, while  
202 particles above 250  $\mu\text{m}$  are absent from the liver.

203  
204 The colour distribution of the MPs that had visibly accumulated is shown in Figure 5. Thus,  
205 overall, 71% of MPs consisted of black or grey filamentous fragments, with blue and green  
206 fragments comprising about 12% of the MP pool. White-transparent and red-pink fragments  
207 contributed about 7 and 8%, respectively, with yellow-orange material lowest in overall abundance  
208 at about 1.3%. There were no clear differences in colours accumulated by different species or in  
209 different organs. However, there were notable differences in the distribution of certain colours  
210 between the different sites; for instance, only one white-transparent fragment and no yellow-  
211 orange fragments were recorded at site 1 while six yellow-orange and 20 white-transparent  
212 fragments were observed at sites 4 and 5, respectively.

213  
214 In addition to the MPs described above and quantified in Tables 2 and 3, a number of fragmented  
215 particles of between a few tens of nm to a few hundred  $\mu\text{m}$  in diameter were observed in the guts

216 and gills of (mainly) pelagic fish that were distinctly different. Thus, EDS revealed the presence  
217 of metals, and mainly Cu, in addition to C and O, while manipulation during analysis and SEM  
218 imagery showed that the material was highly brittle (Figure 6). It is possible that these particles  
219 were of metal construction, at least in part. However, given the detection of both organic material  
220 and Cu, we suspect that these particles are small flakes of paint impregnated with Cu. Most  
221 contemporary antifouling paint formulations employ Cu as a biocide and are generated abundantly  
222 at boat maintenance and repair facilities and are also shed from boat hulls and other painted  
223 maritime structures while in use (Turner, 2010).

224

#### 225 **4. Discussion**

226 This study is one of an emerging number demonstrating the accumulation of MPs by marine  
227 organisms. Of the MPs detected, and consistent with previous environmental studies, they are  
228 mainly fibrous (Lusher et al., 2013; Rochman et al., 2015; Pazos et al., 2017), with sizes ranging  
229 from  $< 100 \mu\text{m}$  to  $> 1000 \mu\text{m}$ . MPs are generally larger in the gills and gastrointestinal tract than  
230 in other organs because larger material can readily enter the digestive environment with relatively  
231 little obstruction; the abundance of MPs in the digestive environment is also rather variable,  
232 reflecting variations in the amount and type of consumed food both between individuals of the  
233 same species and among different species.

234

235 Despite some planktivorous fish seeming to select MPs that are visually similar to their diet (i.e.  
236 blue fragments) (Ory et al., 2017), without information on the colour distribution of MPs in the  
237 water column or sediments of the Musa estuary and Persian Gulf there is no evidence in the present  
238 study for the preferential ingestion or accumulation of MPs according to appearance. We also do

239 not have specific information on the type of plastics found in the organisms sampled, although  
240 MPs retrieved from littoral sediments of the Persian Gulf indicate a predominance of polyethylene,  
241 nylon and polyethylene terephthalate (Naji et al., 2017).

242

243 On an individual basis, MP abundance ranges from about 8 for the prawn, *P. semisulcatus*, to over  
244 20 for the demersal fish, *P. indicus*, that forages in the sediment and where most of the denser MPs  
245 reside. These values are higher than those reported for fish in previous studies; for example, up to  
246 7.2 items per individual were observed in coastal and freshwater fish from China (Jabeen et al.,  
247 2017), up to about 4 per individual were detected in the semi-pelagic Mediterranean fish, *Boops*  
248 *boops* (L.). (Nadal et al., 2016), and an average of 1.6 items per fish were recorded in various  
249 demersal fish in Spanish coastal waters (Bellas et al., 2016). However, it is important to appreciate  
250 that these studies focused on the retrieval of MPs from the digestive tract only. When our data are  
251 restricted to the gut, the average number of MPs per individual ranges from about 1.5 in *S. sihama*  
252 to 3 in *C. abbreviatus* (see Table 2).

253

254 The discrepancies referred to above arise from the general assumption that accumulation of plastics  
255 by fish and other organisms proceeds mainly through ingestion and is, therefore, dependent on  
256 factors like feeding strategy and gut structure as well as the extent of local plastic pollution (Romeo  
257 et al., 2015; Jabeen et al., 2017). Thus, MPs may be accumulated directly and incidentally or  
258 deliberately while feeding from the water column or sifting through contaminated sediment, or  
259 indirectly through the consumption of contaminated prey (Cannon et al., 2016; Jovanović, 2017).  
260 The detection of MPs in the present study in organs not directly involved with ingestion-digestion

261 suggests that other factors may be significant for the accumulation and, potentially, translocation  
262 of MPs in fish.

263  
264 Results of laboratory experiments have reported the occurrence of MPs in the circulatory system  
265 or non-digestive organs of marine invertebrates (Browne et al., 2008; von Moos et al., 2012) and  
266 in the liver of zebrafish (Lu et al., 2016). However, particles employed in these studies were on  
267 the order of tens of micrometers in diameter or less, thereby facilitating passage across the gill or  
268 gut epithelium through cell internalization and subsequent translocation. Collard et al. (2017)  
269 suggest that detection of larger MPs (and of dimensions comparable to those observed here) in the  
270 livers of European anchovies (*Engraulis encrasicolus*) may result from two processes: the  
271 agglomeration of smaller particles and/or passage through the gut barrier by some form of  
272 intracellular or paracellular endocytosis. The former mechanism is unlikely in the present study  
273 because SEM images revealed distinct and relatively smooth fibrous fragments, and without  
274 knowledge of the locations of MPs in (homogenized) tissue the latter mechanism cannot be fully  
275 explained.

276  
277 Alternatively, it has recently been suggested that adherence affords an additional means by which  
278 fibrous MPs may associate with organs independent of the digestive system, in a manner by which  
279 seaweeds accumulate plastics (Gutow et al., 2016). Thus, under laboratory conditions, about 50%  
280 of microfibrils exceeding 100  $\mu\text{m}$  in marine mussels could be accounted for through adherence,  
281 with surface area and “stickiness” two important controls in this respect (Kolandhasamy et al.,  
282 2018). Regardless of the mechanisms by which MPs enter or associate with non-digestive tissues,  
283 their occurrence has a number of implications for evaluating the inventory, location and toxicity

284 of MPs in marine animals, as well as for human health through seafood consumption. Specifically,  
285 if the gut is considered as the sole receptacle, where MPs may either be in transit or entrapped, the  
286 total number of MPs accumulated by an individual may be considerably underestimated. With  
287 respect to toxicity, accumulation outside the digestive tract may induce histological changes and  
288 oxidative stress (Lu et al., 2016) or release contaminants associated with or adsorbed to MPs  
289 (Ashton et al., 2010). The potential for MPs to be transferred to humans should not be  
290 underestimated given that the soft tissue of the species considered are important to the regional  
291 fishing industry. According to the Institute of Standards and Industrial Research of Iran in 2010,  
292 daily average fish muscle consumption is about 7 g/person/day, meaning that about 5 MPs could  
293 be consumed on a daily basis. While there is currently no regulatory framework concerning the  
294 presence of MPs in sea food (European Food Safety Authority 2016), this does not exclude the  
295 possibility that MPs are able to interact with human cells and tissues and facilitate the delivery of  
296 harmful contaminants to the bloodstream (Santillo et al., 2017).

297

## 298 **5. Conclusions**

299 This study has demonstrated the presence of MPs of mainly a fibrous nature and of length < 100  
300  $\mu\text{m}$  to > 1000  $\mu\text{m}$  in various commercially important species of fish and a crustacean collected  
301 from the Musa estuary and the Persian Gulf. Average quantities of MPs ranged from 0.16  $\text{g}^{-1}$  for  
302 the mesopelagic fish, *C. abbreviatus*, to 1.5  $\text{g}^{-1}$  for the prawn, *P. semisulcatus*, with particles  
303 encountered in various tissues from both digestive and non-digestive organs across all species. The  
304 occurrence of MPs outside the digestive system suggests that material can be translocated  
305 following ingestion or that additional, non-ingestive mechanisms (e.g. adherence) are significant.

306 The presence of MPs in non-digestive organs has the potential to induce toxic effects on  
307 individuals and affords an exposure route to humans who consume contaminated fish.

308

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312

### 313 **References**

- 314 • Abbasi, Sajjad., Keshavarzi, Behnam., Moore, Farid., Delshab, Hossein., Soltani,  
315 Naghmeh., Sorooshian, Armin., 2017. Investigation of microrubbers, microplastics and  
316 heavy metals in street dust: A study in Bushehr city, Iran. *Environmental Earth Sciences*.  
317 DOI: 10.1007/s12665-017-7137-0.
- 318 • Akhbarizadeh, R., Moore, F., Keshavarzi, B., 2018. Investigating a probable relationship  
319 between microplastics and potentially toxic elements in fish muscles from northeast of  
320 Persian Gulf. *Environmental Pollution* 232, 154–163.
- 321 • Alomar, C., Estarellas, F., Deudero, S., 2016. Microplastics in the Mediterranean Sea:  
322 Deposition in coastal shallow sediments, spatial variation and preferential grain size.  
323 *Marine Environmental Research* 115, 1-10.
- 324 • Andrady, A.L., 2011. Microplastics in the Marine Environment. *Marine Pollution Bulletin*  
325 62 1596–1605.
- 326 • Ashton, K., Holmes, L., Turner, A., 2010. Association of metals with plastic production  
327 pellets in the marine environment. *Marine Pollution Bulletin* 60, 2050-2055.

- 328 • Auta, H.S., Emenike, C., Fauziah, S., 2017. Distribution and importance of microplastics  
329 in the marine environment: A review of the sources, fate, effects, and potential solutions.  
330 Environment International 102, 165–176.
- 331 • Barnes, D.K.A., Galgani, F., Thompson, R.C., Barlaz, M., 2009. Accumulation and  
332 fragmentation of plastic debris in global environments. Philosophical Transactions of the  
333 Royal Society of London B: Biological Sciences 364, 1985–1998.
- 334 • Bellas, J., Martinez-Armental, J., Martinez-Camara, A., Beseda, V., Martinez-Gómez, C.,  
335 2016. Ingestion of microplastics by demersal fish from the Spanish Atlantic and  
336 Mediterranean coasts. Marine Pollution Bulletin 109, 55-60.
- 337 • Browne, M.A., Dissanayake, A., Galloway, T.S., Lowe, D.M., Thompson, R.C., 2008.  
338 Ingested microscopic plastic translocates to the circulatory system of the mussel, *Mytilus*  
339 *edulis* (L.). Environmental Science and Technology 42, 5026-5031.
- 340 • Cannon, S.M.E., Lavers, J.L., Figueiredo, B., 2016. Plastic ingestion by fish in the  
341 Southern Hemisphere: A baseline study and review of methods. Marine Pollution Bulletin  
342 107, 286-291.
- 343 • Cole, M., Lindeque, P., Fileman, E., Halsband, C., Goodhead, R., Moger, J., Galloway,  
344 T.S., 2013. Microplastic ingestion by zooplankton. Environmental Science and  
345 Technology 47, 6646–6655.
- 346 • Collard, F., Gilbert, B., Compère, P., Eppe, G., Das, K., Jauniaux, T., Parmentier, E., 2017.  
347 Microplastics in livers of European anchovies (*Engraulis encrasicolus*, L.). Environmental  
348 Pollution 229, 1000–1005.
- 349 • De Sá, L.C., Luís, G.L., Guilhermino, L., 2015. Effects of microplastics on juveniles of the  
350 common goby (*Pomatoschistus microps*): confusion with prey, reduction of the predatory

- 351 performance and efficiency, and possible influence of developmental conditions.  
352 Environmental Pollution 196, 359–362.
- 353 • do Sul, J.A.I., Costa, M.F., 2014. The present and future of microplastic pollution in the  
354 marine environment. Environmental Pollution 185, 352–364.
- 355 • European Food Safety Authority, 2016. Presence of microplastics and nanoplastics in food,  
356 with particular focus on seafood. EFSA Journal 14, 4501 (30 pp).
- 357 • Ferreira, P., Fonte, E., Soares, M.E., Carvalho, F., Guilhermino, L., 2016. Effects of multi-  
358 stressors on juveniles of the marine fish *Pomatoschistus microps*: gold nanoparticles,  
359 microplastics and temperature. Aquatic Toxicology 170, 89–103.
- 360 • Gasperi, J., Wright, S.L., Dris, R., Collard, F., Mandin, C., Guerrouache, M., Langlois, V.,  
361 Kelly, F.J., Tassin, B., 2018. Microplastics in air: Are we breathing it in? Current Opinion  
362 in Environmental Science & Health 1, 1–5.
- 363 • Gutow, L., Eckerlebe, A., Gimenez, L., Saborowski, R., 2016. Experimental evaluation of  
364 seaweeds as a vector for microplastics into marine food webs. Environmental Science and  
365 Technology 50, 915–923.
- 366 • Hidalgo-Ruz, V., Gutow, L., Thompson, R.C., Thiel, M., 2012. Microplastics in the marine  
367 environment: A review of the methods used for identification and quantification.  
368 Environmental Science and Technology 46, 3060–3075.
- 369 • Holmes, L.A., Turner, A., Thompson, R.C., 2014. Interactions between trace metals and  
370 plastic production pellets under estuarine conditions. Marine Chemistry 167, 25–32.
- 371 • Hosseini, M., Nabavi, S.M.B., Mansoori, A., Saadatmant, M., 2013. Mercury Levels in  
372 Selected Tissues of Blue Crab *Thalamita prymna* (Portunidae) from Musa Estuary of the  
373 Persian Gulf. Journal of the Persian Gulf 4, 33–38.

- 374 • Jabeen, K., Su, L., Li, J.N., Yang, D.Q., Tong, C.F., Mu, J.L., Shi, H.H., 2017.  
375 Microplastics and mesoplastics in fish from coastal and fresh waters of China.  
376 Environmental Pollution 221, 141-149.
- 377 • Jambeck, J.R., Geyer, R., Wilcox, C., Siegler, T.R., Perryman, M., Andrady, A., Narayan,  
378 R., Law, K.L., 2015. Plastic waste inputs from land into the ocean. Science 347, 768–771.
- 379 • Jovanović, B., Ingestion of microplastics by fish and its potential consequences from a  
380 physical perspective. Integrated Environmental Assessment and Management 13, 510-515.
- 381 • Kolandhasamy, P., Su, L., Li, J., Qu, X., Jabeen, K., Shi, H., 2018. Adherence of  
382 microplastics to soft tissue of mussels: A novel way to uptake microplastics beyond  
383 ingestion. Science of the Total Environment 610-611, 635-640.
- 384 • Li, J., Yang, D., Li, L., Jabeen, K., Shi, H., 2015. Microplastics in commercial bivalves  
385 from China. Environmental Pollution 207, 190–195.
- 386 • Lu, Y., Zhang, Y., Deng, Y., Jiang, W., Zhao, Y., Geng, J., Ding, L., Ren, H., 2016. Uptake  
387 and Accumulation of Polystyrene Microplastics in Zebrafish (*Danio rerio*) and Toxic  
388 Effects in Liver. Environmental Science & Technology 50, 4054–4060.
- 389 • Luís, L.G., Ferreira, P., Fonte, E., Oliveira, M., Guilhermino, L., 2015. Does the presence  
390 of microplastics influence the acute toxicity of chromium (VI) to early juveniles of the  
391 common goby (*Pomatoschistus microps*)? A study with juveniles from two wild estuarine  
392 populations. Aquatic Toxicology 164, 163–74.
- 393 • Lusher, A., McHugh, M., Thompson, R.C., 2013. Occurrence of microplastics in the  
394 gastrointestinal tract of pelagic and demersal fish from the English Channel. Marine  
395 Pollution Bulletin 67, 94-99.

- 396 • Lusher, A.L, Tirelli, V., O'Connor, I., Officer, R., 2015. Microplastics in Arctic polar  
397 waters: The first reported values of particles in surface and sub-surface samples. Scientific  
398 Reports 5, 14947
- 399 • Moore, C.J., 2008. Synthetic polymers in the marine environment: a rapidly increasing,  
400 long-term threat. Environmental Research 108, 131–139.
- 401 • Nadal, M.A., Alomar, C., Deudero, S., 2016. High levels of microplastic ingestion by the  
402 semipelagic fish bogue *Boops boops* (L.) around the Balearic Islands. Environmental  
403 Pollution 214, 517–23.
- 404 • Naji, A., Esmaili, Z., Khan, F.R., 2017. Plastic debris and microplastics along the beaches  
405 of the Strait of Hormuz, Persian Gulf. Marine Pollution Bulletin 114, 1057-1062.
- 406 • Ory, N.C., Sobral, P., Ferreira, J.L., Thiel, M., 2017. Amberstripe scad *Decapterus*  
407 *muroadsi* (Carangidae) fish ingest blue microplastics resembling their copepod prey along  
408 the coast of Rapa Nui (Easter Island) in the South Pacific subtropical gyre. Science of the  
409 Total Environment 586, 430-437.
- 410 • Pazos, R.S., Maiztegui, T., Colautti, D.C., Paracampo, A.H., Gómez, N., 2017.  
411 Microplastics in gut contents of coastal freshwater fish from Rio de la Plata estuary. Marine  
412 Pollution Bulletin 122, 85-90.
- 413 • Rastegari Mehr, M., Keshavarzi, B., Moore, F., Sacchi, E., Lahijanzadeh, A.R., Eydivand,  
414 S., Jaafarzadeh, N., Naserian, S., Setti, M., Rostami, S., 2016. Contamination level and  
415 human health hazard assessment of heavy metals and polycyclic aromatic hydrocarbons  
416 (PAHs) in street dust deposited in Mahshahr, southwest of Iran. Human and Ecological  
417 Risk Assessment: An International Journal 22, 1726–1748.

- 418 • Rist, S., Almroth, B.C., Hartmann, N.B., Karlsson, T.M., 2018. A critical perspective on  
419 early communications concerning human health aspects of microplastics. *Science of The*  
420 *Total Environment* 626, 720–726.
- 421 • Rochman, C.M., Tahir, A., Williams, S.L., Baxa, D.V., Lam, R., Miller, J.T., Teh, F.C.,  
422 Werorilangi, S., Teh, S.J., 2015. Anthropogenic debris in seafood: plastic debris and fibers  
423 from textiles in fish and bivalves sold for human consumption. *Scientific Reports*. 5, 14340.
- 424 • Romeo, T., Pietro, B., Peda, C., Consoli, P., Andaloro, F., Fossi, M.C., 2015. First evidence  
425 of presence of plastic debris in stomach of large pelagic fish in the Mediterranean Sea.  
426 *Marine Pollution Bulletin* 95, 358-361.
- 427 • Santillo, D., Miller, K., Johnston, P., 2017. Microplastics as contaminants in commercially  
428 important seafood species. *Integrated Environmental Assessment and Management* 13,  
429 516-521.
- 430 • Teuten, E.L., Rowland, S.J., Galloway, T.S., Thompson, R.S., 2007. Potential for plastics  
431 to transport hydrophobic contaminants. *Environmental Science and Technology* 41, 7759–  
432 7764.
- 433 • Turner, A., 2010. Marine pollution from antifouling paint particles. *Marine Pollution*  
434 *Bulletin* 60, 159-171.
- 435 • Turner, A., 2017. In situ elemental characterisation of marine microplastics by portable  
436 XRF. *Marine Pollution Bulletin* 124, 286-291.
- 437 • Van Cauwenberghe, L., Janssen, C.R., 2014. Microplastics in bivalves cultured for human  
438 consumption. *Environmental Pollution* 193, 65–70.

- 439 • Van Cauwenberghe, L., Claessens, M., Vandegehuchte, M.B., Janssen, C.R., 2015.  
440 Microplastics are taken up by mussels (*Mytilus edulis*) and lugworms (*Arenicola marina*)  
441 living in natural habitats. *Environmental Pollution* 199, 10-17.
- 442 • von Moos, N., Burkhardt-Holm, P., Koehler, A., 2012. Uptake and effects of microplastics  
443 on cells and tissue of the blue mussel *Mytilus edulis* L. after an experimental exposure.  
444 *Environmental Science and Technology* 46, 11327-11335.
- 445 • Wang, J., Tan, Z., Peng, J., Qiu, Q., Li, M., 2016. The behaviors of microplastics in the  
446 marine environment. *Marine Environmental Research* 113, 7–17.
- 447 • Woodall, L.C., Gwinnett, C., Packer, M., Thompson, R.C., Robinson, L.F., Paterson,  
448 G.L.J., 2015. Using a forensic science approach to minimize environmental contamination  
449 and to identify microfibrils in marine sediments. *Marine Pollution Bulletin* 95, 40–46.
- 450 • Wright, S.L., Thompson, R.C., Galloway, T.S., 2013. The physical impacts of  
451 microplastics on marine organisms: A review. *Environmental Pollution* 178, 483–492.
- 452 • Ziccardi, L.M., Edgington, A., Hentz, K., Kulacki, K.J., Driscoll, S.K., 2016. Microplastics  
453 as vectors for bioaccumulation of hydrophobic organic chemicals in the marine  
454 environment: a state-of-the-science review. *Environmental Toxicology and Chemistry* 35,  
455 1667–1676.

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457 **Figure captions:**

458 **Figure 1:** Locations of the five sampling sites along the coast of the Musa estuary.

459 **Figure 2:** Examples of MPs encountered in fish and prawn tissues and as captured by binocular  
460 microscope. Note that fibres in panels (b) and (g) are extremely thin and, therefore, have a  
461 relatively high propensity to penetrate tissue, and that fibres in panels (c), (d), (f), (h), (i) and (j)  
462 exhibit partial entrapment in half-digested tissues.

463 **Figure 3:** An image and the composition of a fibre obtained by SEM/EDS (W% = weight percent  
464 and A% = atomic percent) (a); fibre images obtained using upper-light fluorescence microscopy  
465 (b,c); fibre images obtained by polarized downward projecting light microscopy (e,g) and  
466 corresponding images obtained without polarized light (d,f).

467 **Figure 4:** The net distribution of MPs among different size categories (in  $\mu\text{m}$ ) in the five species.

468 **Figure 5:** Overall colour distribution of the MPs observed in the samples.

469 **Figure 6:** SEM/EDS image and composition of a particle of a relatively brittle and non-fibrous  
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472 **Table 1:** Number of species caught from each site (*n*) together with the mean (and minimum and  
 473 maximum) lengths (cm) and weights (g).

		<i>S. sihama</i>	<i>P. indicus</i>	<i>C. abbreviatus</i>	<i>S. tumbil</i>	<i>p. semisulcatus</i>
S1	<i>n</i>	4	3	4		
	length	20.1 (17.2-20.1)	17.3(16.5-18.5)	17.7 (14.2-20.5)		
	weight	67.8 (49.3-95.1)	23.8 (18.3-32.7)	33.1 (14.2-56.6)		
S2	<i>n</i>	4	1	4		
	length	16.6 (13.0-20.0)	16.0	17.7 (14.2-20.5)		
	weight	39.4 (14.2-62.4)	16.8	33.1 (14.2-56.6)		
S3	<i>n</i>					5
	length					7.8 (5.5-10.0)
	weight					5.4 (2.3-10.6)
S4	<i>n</i>	4	4	3	4	3
	length	16.6 (15.5-18.0)	19.5 (18.5-21.5)	23.7 (23.0-24.0)	15.7 (13.0-18.0)	7.3 (6.0-8.5)
	weight	45.6 (36.4-53.3)	46.7 (35.7-63.6)	115.9 (109.3-123.0)	36.1 (18.6-50.4)	5.2 (2.5-8.0)
S5	<i>n</i>	5	4	4		4
	length	20.5 (18.5-24.5)	20.5 (20.0-22.0)	23.8 (22.5-26.0)		7.6 (4.5-10.5)
	weight	72.2 (51.8-119.9)	41.7 (39.1-46.5)	88.4 (75.2-115.4)		4.9 (1.4-8.7)
total	<i>n</i>	17	12	15	4	12
	length	18.6 (13.0-24.5)	19.0 (16.0-22.0)	24.6 (14.2-21.7)	15.7 (13.0-18.0)	7.6 (4.5-10.5)
	weight	57.2 (14.2-119.9)	36.8 (16.8-63.6)	75.8 (14.2-123.0)	36.1 (18.6-50.4)	5.2 (1.4-10.6)

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480 **Table 2:** Number of MPs detected in the five species pooled from each site (with the number of  
 481 species given in Table 1). Also shown is the total number of MPs in each species, the mean number  
 482 when normalized for the number of individuals analysed and the average mass of individuals, and  
 483 the mean number per individual when only the gut was considered.

		<i>S. sihama</i>	<i>P. indicus</i>	<i>C. abbreviatus</i>	<i>S. tumbil</i>	<i>P. semisulcatus</i>
S1	skin	7	27			
	muscle	14	7			
	gut	1	11			
	gills	15	22			
	liver	6	0			
S2	skin	29	14	8		
	muscle	20	21	10		
	gut	9	4	11		
	gills	12	12	12		
	liver	4	0	5		
S3	skin					23
	muscle					12
	gut					
	gills					
	liver					
S4	skin	14	14	8	6	21
	muscle	19	14	12	12	14
	gut	12	12	18	11	
	gills	20	27	13	8	
	liver	11	13	24	17	
S5	skin	11	27	13		14
	muscle	11	13	12		10
	gut	4	0	15		
	gills	8	23	8		
	liver	12	0	11		
total		239	261	180	54	94
mean/individual		14.1	21.8	12.0	13.5	7.8
mean/g		0.25	0.59	0.16	0.37	1.51
mean/gut		1.5	2.3	2.9	2.8	

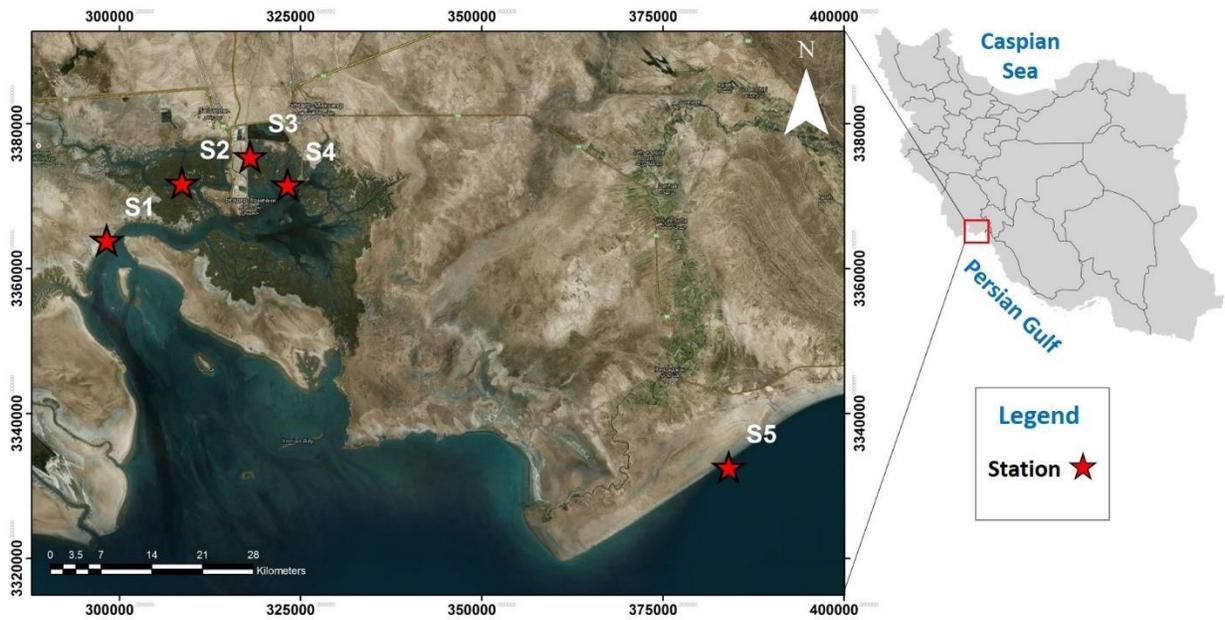
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487 Fig 1.

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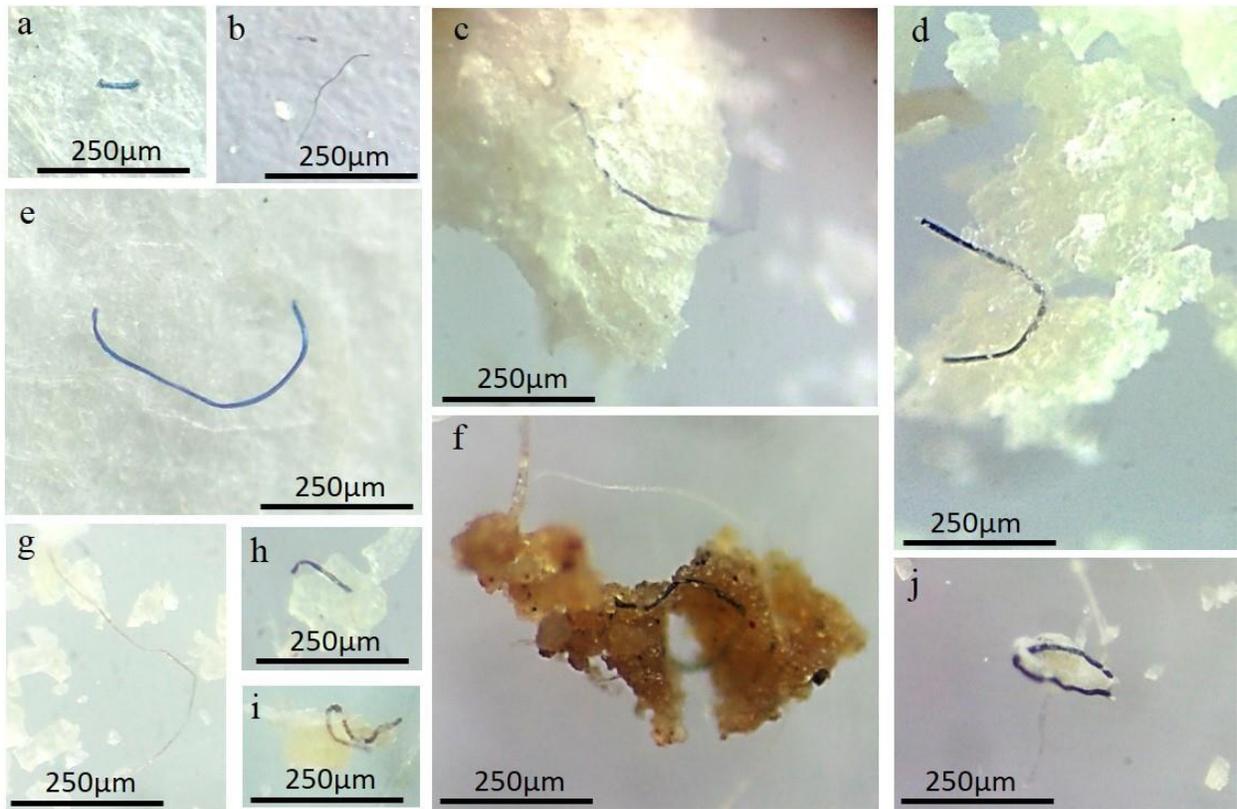
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495 Fig 2

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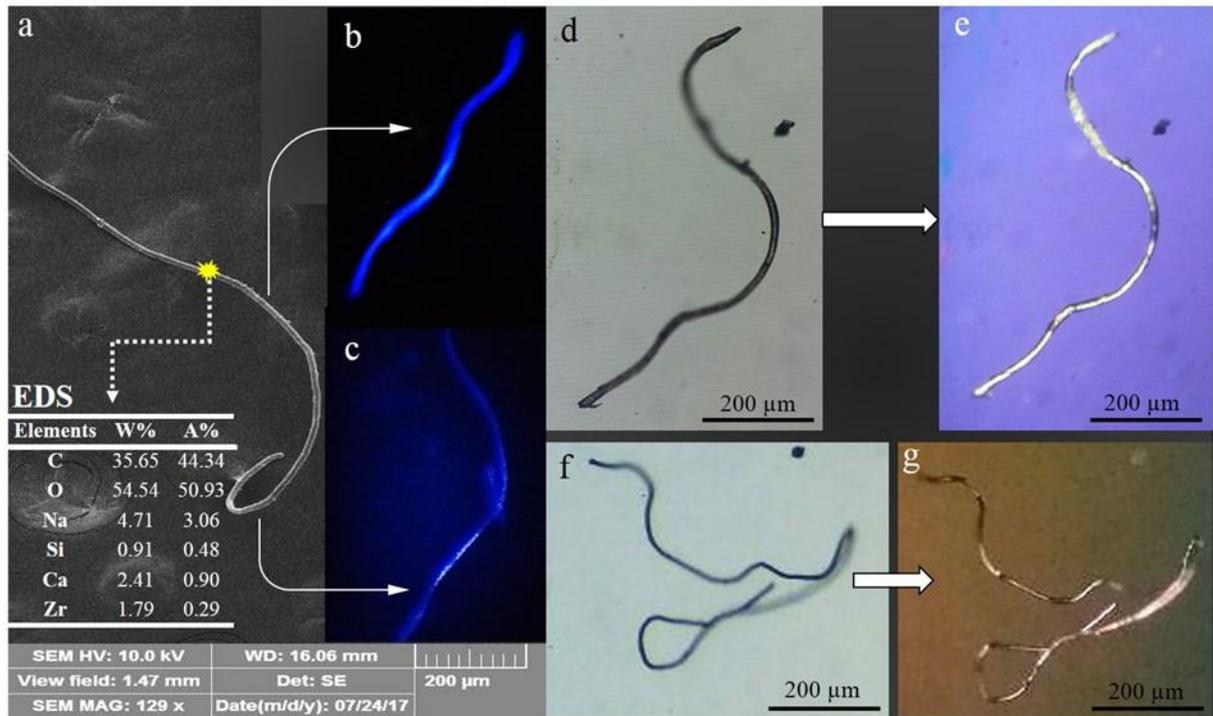


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499 Fig 3

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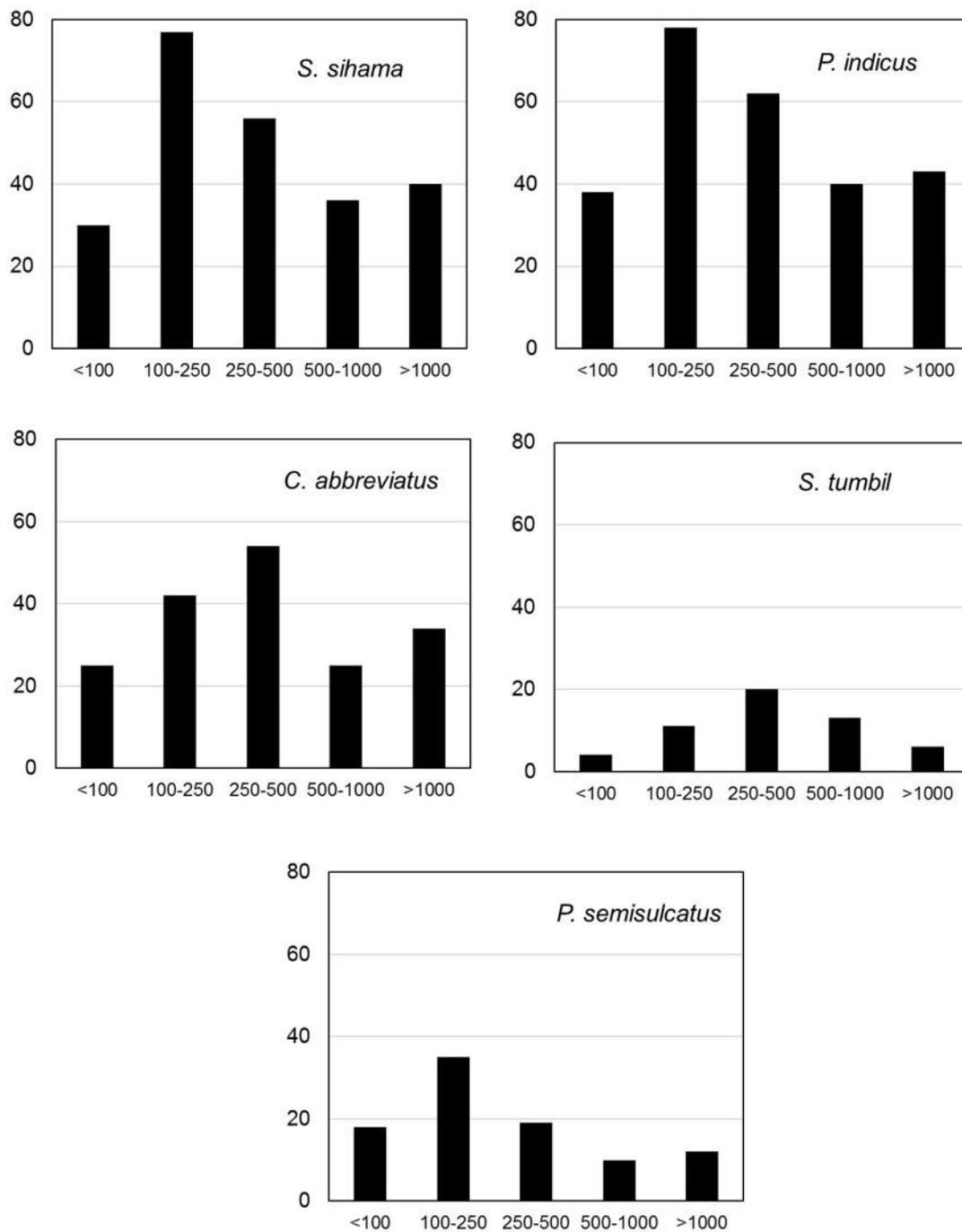
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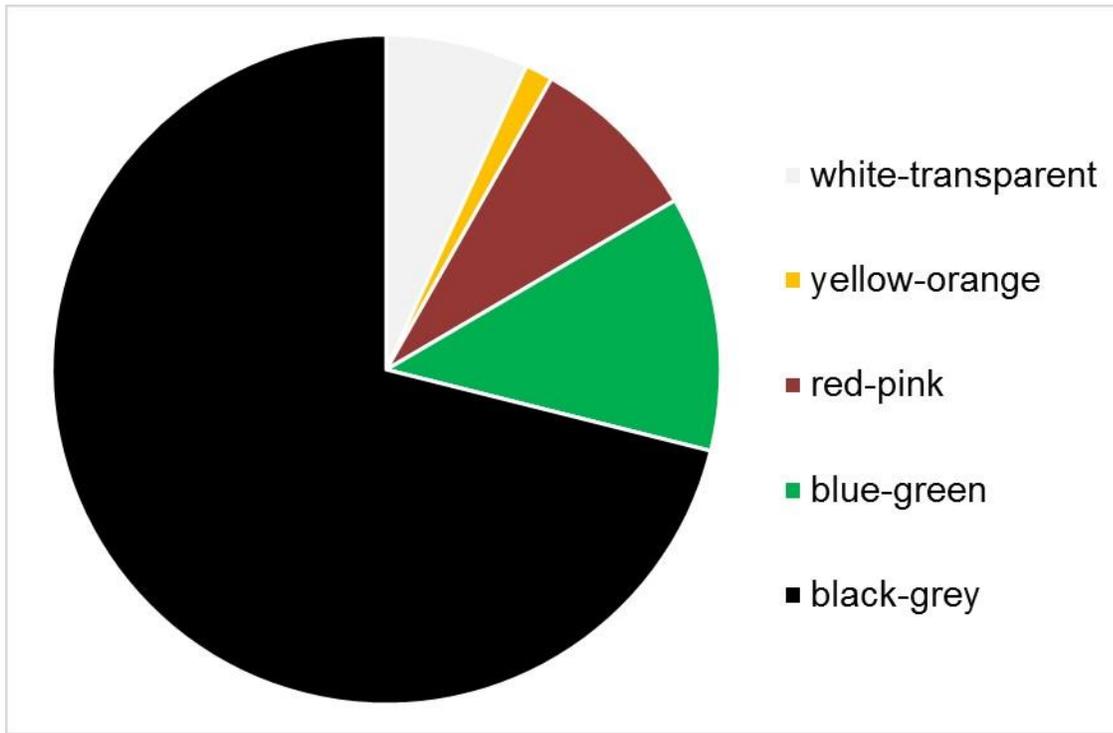


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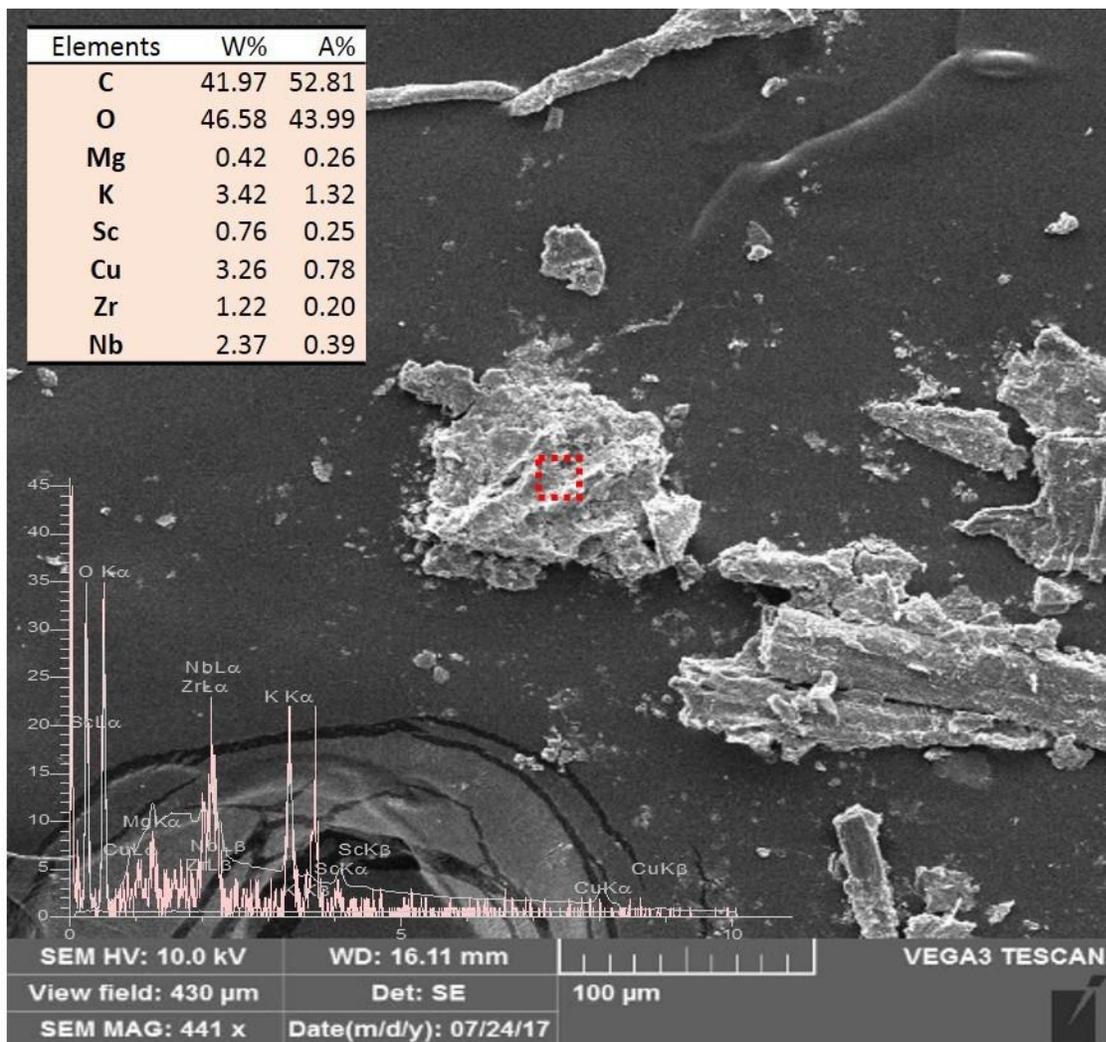
512 Fig 5



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515 Fig 6



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519 re (W% = weight percent and A% = atomic percent).