Faculty of Science and Engineering

School of Geography, Earth and Environmental Sciences

2018-06

Changes to polychlorinated biphenyl (PCB) signatures and enantiomer fractions across different tissue types in Guillemots

Megson, D

http://hdl.handle.net/10026.1/11282

10.1016/j.marpolbul.2018.04.014 Marine Pollution Bulletin Elsevier

All content in PEARL is protected by copyright law. Author manuscripts are made available in accordance with publisher policies. Please cite only the published version using the details provided on the item record or document. In the absence of an open licence (e.g. Creative Commons), permissions for further reuse of content should be sought from the publisher or author.

1 NOTE: This is a preproduction accepted draft – published copies are available

- 2 **from Marine Pollution Bulletin:**
- 3 https://doi.org/10.1016/j.marpolbul.2018.04.014
- 4

5 Changes to polychlorinated biphenyl (PCB) signatures and enantiomer 6 fractions across different tissue types in Guillemots

- David Megson ^{a b*}, Thomas A. Brown ^{b c}, Gwen O'Sullivan ^d, Matthew Robson ^e, Xavier Ortiz
 ^e Paul J. Worsfold ^b, Sean Comber ^b, Maeve C. Lohan ^b, Eric J. Reiner ^e.
- ^{a.} School of Science and the Environment, Manchester Metropolitan University, Manchester
 M1 5GD, UK.
- ^{b.} Biogeochemistry Research Centre, SoGEES, Plymouth University, Plymouth, Devon, PL4
 8AA, UK
- ^c Marine Ecology and Chemistry, Scottish Association for Marine Science, Oban, PA37 1QA,
 UK
- 15 ^{d.} Department of Earth & Environmental Science, Mount Royal University, 4825 Mount Royal
- 16 Gate SW, Calgary, Alberta, T3E 6K6, Canada
- 17 ^{e.} Ontario Ministry of the Environment and Climate Change, 125 Resources Road, Toronto,
- 18 Ontario, M9P 3V6, Canada
- 19

20 *Corresponding Author

- 21 David Megson
- 22 School of Science and the Environment,
- 23 Manchester Metropolitan University,
- 24 Manchester, M1 5GD, UK
- 25 E-mail: <u>dpmegson@hotmail.co.uk</u>
- 26

27 Abstract

28 Two Guillemot carcases were dissected, each providing 12 discrete tissue samples and 3 29 samples of partially digested food. One hundred and five PCBs from the 209 PCBs determined by GCxGC-ToFMS were detected. The relative proportions of individual PCBs did not vary 30 31 greatly within tissue types, although the PCB profile from undigested food could be 32 distinguished. Enantiomer fractions (EFs) were determined for CB-95, CB-136 and CB-149 by 33 GC-HRqToFMS. EFs in the partially digested food were near racemic, with high levels of enrichment for E1 CB-95 in the kidneys and liver (EF of 0.80 and 0.84 respectively). This 34 35 provides some of the clearest evidence to date that fractionation takes place in the organs 36 where metabolic biotransformation and elimination of PCBs occurs. Our findings also confirm

- 37 the ability of non-lethal sampling techniques, such as collection of small (<1 g) blood samples,
- 38 to provide PCB signatures that are representative of an individual organism.

39 Key Words

40 Polychlorinated biphenyl (PCBs); Guillemot (*Uria aalge*); North Atlantic; Tissue;
41 Comprehensive two-dimensional gas chromatography (GCxGC); Enantiomer fraction

42 1 Introduction

Polychlorinated biphenyls (PCBs) predominantly enter animals through ingestion of 43 contaminated food. This can result in the accumulation of PCBs, with higher PCB 44 45 concentrations usually associated with lipid rich tissues (Maervoet et al., 2005, Karjalainen et 46 al., 2006). Biomagnification can lead to elevated PCB concentrations in top predators (Hansen, 47 1999, Muir et al., 1988). In most cases the PCB signature in animals can be largely explained 48 by their food source (Jaspers et al., 2013), although human induced changes in land use can also influence the signature (Fernie et al., 2008). Once incorporated, biotransformation and 49 50 elimination of PCBs can vary from species to species. For example, animals such as bears 51 and humans have been shown to be capable of metabolising some PCBs, while the equivalent 52 capacity has not been observed in predatory birds (Jaspers et al., 2013).

53 Some studies have provided evidence that the relative proportions of PCBs varies between 54 different tissue types as a result of the preferential accumulation of, for example, ortho-55 chlorinated PCBs in the brain of rats (Kodavanti et al., 1998). PCBs are highly soluble in lipids 56 and therefore accumulate in tissues and organs according to their respective Kow-dependent 57 release rates (Karjalainen et al., 2006). The accumulation of PCBs from digested food occurs 58 as PCBs partition across the membrane lining the gastrointestinal tract into the bloodstream. The blood flow in different tissues initially drives the distribution of PCBs until an equilibrium 59 60 is reached which is primarily driven by the tissue lipid content (Karjalainen et al., 2006). 61 However, despite the variable relative accumulation rates of PCBs, highly similar PCB 62 distributions have been previously measured in birds (Boumphrey et al., 1993), specifically in 63 the gut, heart, liver and stomach tissues from a Leach's storm petrel (Megson et al., 2014). 64 Relatively little is known about the relative proportions of individual PCBs within an organism 65 as a function of the observed non-uniform accumulation, indicating the necessity for further 66 work.

67 There are 19 out of 209 PCBs that are predicted to exist as stable atropisomers (Oki, 1983). 68 In commercial mixtures both enantiomers are produced in equal proportions and thus are 69 racemic. In animals, metabolic processes such as enzyme mediated oxidation have been 70 proven to preferentially target one stereoisomer, resulting in atropisomeric enrichment (Harrad 71 et al., 2006, Wong et al., 2002, Wu et al., 2014). The degree of enrichment is species specific 72 and can vary for the different enantiomers measured. There is currently little information on 73 how the enantiomer fractions vary in different organs in animals. Chu et al. (2003) showed 74 that PCBs 95, 132 and 149 are near racemic in human muscle, brain and kidney tissue, 75 whereas Kania-Korwel et al. (2010) identified enrichment of CB-95 in the blood, adipose tissue, 76 brain and kidneys of mice.

Here we present the results for the determination of all 209 PCB congeners and three atropisomers in 30 tissue samples obtained from two common guillemot (*Uria aalge*) birds. This study examines potential changes to the PCB signature and enantiomer fraction that may occur in different organs. The results are discussed in the context of providing evidence for the adoption of ethical, non-fatal sampling techniques, such as blood collection, to provide a reliable indicator of the PCB signature in future studies.

83 2 Materials and Methods

For this study we use the common guillemot (Uria aalge) to investigate if the relative 84 85 proportions of PCBs and enantiomer fractions change within an individual organism. The 86 common guillemot is the most abundant seabird breeding in the UK, with an estimated 1 million breeding pairs (Harris and Wanless, 2004). Guillemots are colonial, cliff-nesting 87 seabirds that spend a large proportion of their time at sea foraging for food. The guillemot's 88 89 diet is primarily comprised of benthic fish from the Ammodytidae, Gadidae and Clupeidae 90 families but also includes a wide variety of invertebrates such as crustaceans, annelids and 91 molluscs (Anderson et al., 2014, Bradstreet and Brown, 1985). The guillemot's diet is known 92 to vary considerably due to the availability of prey.

93 2.1 Sample collection and preparation

94 Two guillemot carcases were collected from the south coast of the UK. The carcases were in 95 good condition, with minimal degradation or damage, which enabled detailed dissections 96 yielding samples from 12 different tissue types including: the kidney, heart, breast muscle, 97 intestines, leg muscle, liver, blood, brain, pancreas, proventriculus, duodenum and gonads 98 from each bird. Three samples were also obtained from partially digested food from within the 99 gastrointestinal tract including the proventriculus contents, duodenum contents and intestine 100 contents.

PCB signatures were determined by comprehensive two-dimensional gas chromatography coupled with time of flight mass spectrometry (GCxGC-ToFMS) and interpreted using principal component analysis to compare the relative proportions of 74 common PCB congeners in the different tissue types. Enantiomer fractions were determined using gas chromatography coupled with high resolution time of flight mass spectrometry.

106 2.2 Extraction procedure

107 Sample extraction was undertaken following the established method for PCB extraction in 108 tissues reported by Megson et al. (2013) and outlined in Brown et al. (2013). All samples were 109 freeze-dried (-45 °C; 0.2 mbar; 72 h) and ground into a powder. Samples were saponified in a 110 methanolic potassium hydroxide solution (~ 4 mL H₂O:MeOH, 1:9; 20% KOH) for 60 min (80 °C). Hexane (3 x 4 mL) was added to the saponified solutions, which were then vortexed (1 111 112 min) and centrifuged (1 min; 2,000 rpm). Supernatant solutions containing non-saponifiable 113 lipids (NSLs) were transferred to clean vials with glass pipettes and dried using nitrogen to 114 remove traces of H₂O/MeOH. NSLs were then re-suspended in hexane (0.5 mL) and 115 fractionated (5 mL hexane) using column chromatography (SiO₂; 0.5 g). Samples were 116 evaporated to incipient dryness and reconstituted with 10 µL of an internal standard 117 comprising ¹³C₁₂ PCBs 60, 127 and 159 at a concentration of 10 ng mL⁻¹ (CIL-EC-5370 EN-118 1948-4 PCB sampling standard, LGC) and 90 µL of hexane prior to analysis.

119 2.3 PCB signature analysis (GCxGC-ToFMS)

120 2.3.1 Analytical procedure

121 Samples were analysed to determine the presence of all 209 PCBs using the methods 122 described by Megson et al. (2013) using a time-of-flight mass spectrometer (LECO, St. Joseph, 123 MI Pegasus 4D) coupled to a two dimensional gas chromatograph (Agilent Technologies 124 7890A) equipped with a thermal modulator (LECO, St. Joseph, MI). The gas chromatograph was installed with a Rtx-PCB (60 m x 0.18 mm x 0.18 μ m) ¹D column and a Rxi-17 (1.5 m x 125 126 0.1 mm x 0.1 µm) ²D column. A sample volume of 1 µL was injected in splitless mode. All data files were processed using ChromaTOF software set to identify 10,000 peaks with a signal-to-127 128 noise ratio of > 10:1.

129 2.3.2 Data Quality

130 Analytical blanks were run with each batch of approximately 10 samples. All samples were spiked with a ¹³C₁₂ internal standard (CB-60, CB-127, CB-159) which was used to quantify 131 PCB concentrations by isotope dilution. Concentrations were normalised to dry weight tissue 132 133 mass and are therefore reported as µg g⁻¹. As samples were originally extracted for the 134 analysis of other lipids, PCB recovery could not be accurately determined for each sample; 135 therefore reported concentrations were not corrected based on sample recovery or lipid 136 content. PCBs are located within the lipid fraction, which was quantitatively extracted, 137 therefore any bias should not be significant. Furthermore, because results for enantiomer 138 fractions are relative these values are not biased and recovery correction is not necessary. Limits of detection (LOD) for individual PCBs were in the range $0.1 - 5 \text{ ng g}^{-1}$ (dry weight). 139 140 Accuracy and precision were measured for the sum of the European Union 7 indicator 141 congeners (EC7) (CB-28, CB-52, CB-101, CB-118, CB-138, CB-153, CB-180) by analysing a 10 mg L⁻¹ Aroclor 1248 standard three times. The sum of the EC7 congeners for the three 142 143 samples was $105 \pm 0.9 \% (1 \sigma)$.

144 2.4 Chiral analysis (GC-HRqToFMS)

145 2.4.1 Analytical procedure

146 The Enantiomeric Fractions (EFs) of CBs 95, 136 and 149 were analysed based on the gas chromatography conditions specified by Robson and Harrad (2004). Samples were analysed 147 using an Agilent 7890 Gas Chromatograph coupled to a Waters Xevo G2-XS qTOF based on 148 149 the conditions specified in Megson et al. (2016) The corona voltage was set at 5 mAu, the cone gas at a flow rate of 175 L h⁻¹, and the desolvation gas flow set at 175 L h⁻¹. Ionization 150 was undertaken using an atmospheric pressure chemical ionization source at 150 °C with the 151 152 detector run in full scan mode using two target enhanced functions on masses 326 and 360. 153 The two most abundant isotopes of each enantiomer were recorded with a mass accuracy of 154 <1ppm.

155 2.4.2 Data Quality

The chromatographic performance of the method was assessed prior to each run of 10 samples by analysing a 1:1:1 mixture of Aroclors 1248, 1252 and 1260. Enantiomeric fractions 158 were calculated as per Harner et al. (2000), whereby EF = E1/(E1+E2), E1 is the first eluting 159 or the (+) enantiomer and E2 is the second eluting enantiomer. Samples where only accepted for quantitation if; the enantiomeric fractions of the three atropisomers studied were 0.50 160 161 (±0.01) in the Aroclor mixture; the least abundant enantiomer of the pair had a signal to noise (S:N) ratio greater than 10:1; and the isotope ratios were within 20% of their theoretical values. 162 The instrumental LODs were calculated by analysing a standard mixture of CB-95 and CB-163 149; LODs were established at a concentration of 0.1 pg μ L⁻¹ per enantiomer. Procedural 164 165 blanks were prepared for each batch of 10 samples; no chiral PCBs were detected in the 166 blanks above the LODs.

167 **2.5 Statistical analysis**

168 Exploratory data analysis was undertaken using principal component analysis (PCA) 169 performed using PRIMER 6 software. PCA is a statistical technique that is often used to 170 simplify complex datasets as it reduces the dimensionality of the dataset by transforming it to 171 a set of new uncorrelated eigenvectors called principal components (Johnson et al., 2002).

172 Where a PCB was not detected it was included in the dataset as a '0'. As part of the data 173 quality check, other values were substituted for '0', including the smallest integrated peak area 174 and the smallest integrated peak area divided by 2, but these had no observable effect on the 175 data output and so the '0's were retained. To reduce any bias from a high proportion of non-176 detects for a specific congener, PCBs that were not detected in over 60% of samples (i.e. 177 PCBs present in less than 18 out of the 30 samples) were removed from the analysis following the guidance of Helsel (2006). This resulted in a data set containing 30 samples and 74 PCBs. 178 179 Before performing PCA the data were normalised by transformation to a percent metric to 180 remove concentration/dilution effects. The data were then mean centred and scaled using a 181 Z-transform (autoscale transform) to prevent high concentration variables from dominating the 182 analysis (Johnson et al., 2007).

183 3 Results and discussion

184 **3.1** PCB concentrations and signatures in Guillemot tissues

185 A total of 105 different PCBs were detected in the samples. PCBs present in the highest 186 concentrations included PCBs 118, 146, 153, 163 and 187. These congeners are regularly 187 detected in the environment as they were present in high proportions in Aroclor mixtures 188 (Frame et al., 1996). However, it was interesting to note that the non-Aroclor PCBs 11 and 189 209 were also detected in the samples (estimated at approximately 0.01% and 0.05% respectively of total PCBs). Their presence in these samples provides more evidence to show 190 191 that they are now ubiquitous contaminants in the environment (Hu et al., 2011, King et al., 192 2002, Hu et al., 2008, Rodenburg et al., 2010). PCB concentrations were calculated for the 193 EC7 congeners and ranged from 0.19 μ g g⁻¹ to 69 μ g g⁻¹. These values were greater than 194 those reported in common terns (Sterna hirundo) from Ireland (0.035 μ g g⁻¹) (Acampora et al., 195 2017), and comparable to levels reported in harbour porpoises (Phocoena phocoena) in UK 196 waters (0.4 to 160 µg g⁻¹) (Jepson et al., 2016). However, concentrations were approximately 197 one order of magnitude lower than several other marine mammals from European waters where levels of over 100 µg g⁻¹ were regularly detected in bottlenose dolphins (*Tursiops* 198 199 truncates), striped dolphins (Stenella coeruleoalba) and killer whales (Orcinus orca) (Jepson

et al., 2016). While total EC7 PCB concentrations were generally similar between comparable
 tissues of the two birds, concentrations did differ greatly between particular tissue types
 (Figure 1). For example, the highest EC7 PCB concentrations were identified in the lipid-rich
 gonads, where concentrations were approximately one order of magnitude greater than those
 measured in other tissues (Figure 1).



205

Figure 1. EC7 PCB concentrations in different tissue types and partially digested food from the two Guillemot birds sampled.

In an effort to identify a more suitable means of comparing PCBs between individuals the relative proportion of PCBs (referred to here as a 'signature') obtained from the 12 tissue types and 3 samples of partially digested food were compared (individual signatures are presented in Supplementary information 1). The PCB signatures were comprised of 74 PCBs and were visually similar for each tissue type within and between birds (Figure 2). This is consistent with findings reported in Jaspers et al. (2013) who also identified that PCB signatures are highly influenced by food source.



215

Figure 2. Average signature in 12 tissue samples from each bird for 74 PCBs, error bars represent +/- 1 standard deviation.

Despite both birds operating in the same foraging area it was still possible to identify subtle differences between samples from each bird using principal component analysis (Figure 3). Within Figure 3 it can be observed that organs from each bird can be differentiated based on the influence of principal component 2. We suspect that this is likely to be due to the different sexes of the two birds, rather than geographical differences in foraging. However, it could also be explained by an age difference since Bird A had slightly higher total PCB concentrations with higher proportions of the more chlorinated PCBs which are generally more resistant to biotransformation and elimination (Hansen, 1999). Although samples were obtained from only two individuals, the results clearly indicate that PCB signatures remain constant between the different tissue types analysed in each bird, which is an important finding that is consistent with previous studies undertaken on fewer organs (Boumphrey et al., 1993, Megson et al., 2014)

Figure 3 also highlights the variation in PCB signature in partially digested food and tissue. 230 231 The PCB signature derived from partially digested food items, in particular from the 232 proventriculus (first part of the birds stomach), differed from those measured in the respective 233 host bird (Supplementary Information 1) and strongly influenced principal component 1 of the 234 scores plot (Figure 3). Closer examination of the individual PCBs revealed that this was 235 primarily due to higher proportions of the less chlorinated PCBs, along with CB-118, CB-99 & 236 101, CB-105 and CB-147 & 149, in the partially digested food (Figure 4; Supplementary Information 1). Uniquely, eight PCBs that were absent from the tissue samples were recorded 237 in all of the partially digested food samples (CB-44, CB-49, CB-64, CB-71, CB-87, CB-88&95, 238 239 and CB-179). The majority of these PCBs contained either -25 or -236 substitution patterns. 240 Many of the congeners present in high proportions in the undigested food have also previously been reported in high proportions in members of the Gadidae family (Megson et al., 2013) 241 from the southwest coast of Great Britain, which guillemots are known to feed on (Anderson 242 et al., 2014). Since both samples from the proventriculus contents grouped in a similar area 243 244 in the scores plot (Figure 3), we suggest that both birds had recently consumed a similar prey 245 type.



Figure 3. Scores plot of PC1 and PC2 showing the difference in PCB signature between the two birds, along with a different PCB signature in the partially digested food samples.

246



249

Figure 4. Loadings plot of PC1 and PC2 showing higher proportions of the less chlorinated biphenyls in the partially digested food. PCBs with 1 to 4 chlorines are coloured in **light grey**, PCBs with 5 and 6 chlorines are in **dark grey**, and PCBs with 7 to 10 chlorines are in **black**.

253 The ability of principal component analysis to distinguish samples recovered from the two birds 254 indicates a strong degree of perpetuation of the PCB signature in different tissues. This finding is consistent with previous studies undertaken on fewer tissue types, which have shown that 255 blood flow in different tissues is effective at distributing PCBs around the body until an 256 257 equilibrium is reached which is primarily driven by the tissue lipid content (Karjalainen et al., 258 2006). Here we show that in the guillemot this redistribution does not appear to cause fractionation of more lipophilic PCBs to more lipid rich tissues. Instead the blood flow appears 259 260 to distribute PCBs relatively evenly, resulting in comparable PCB signatures for blood and all 261 analysed tissues. The consistency of PCB signatures in the different tissue types within each 262 bird indicates that a small mass of blood (<1g) can be used to represent the PCB signature of 263 the organism as a whole. This would correspond to approximately 1 mL of blood which could 264 be sourced from most birds without any detrimental effects.

265 3.2 Chiral PCBs in different tissue types

Chiral PCBs have previously been monitored in wildlife and results to date indicate that enantioselective processing can occur for several species (Wong et al., 2002, Buckman et al., 2006, Warner et al., 2009, Kania-Korwel et al., 2010, Kania-Korwel et al., 2008a, Kania-Korwel et al., 2007, Kania-Korwel et al., 2008b, Kania-Korwel and Lehmler, 2016). Despite the relatively large number of studies on chiral PCBs in animals it is currently unclear if atropselective metabolism at the site of absorption or in other, extrahepatic tissues contributes to the atropisomeric enrichment of chiral PCBs (Kania-Korwel and Lehmler, 2016). This study aims to help address this knowledge gap through the analysis of 12 different tissue samplesand 3 partially digested food samples in two birds (Supplementary Information 2).

275 Concentrations of chiral PCBs were below the analytical limits of detection (0.1 pg μ L⁻¹) in many of the tissue samples analysed. CB-149 was not detected during this study, although 276 277 both CB-95 and CB-136 were identified in several samples. The available results indicate 278 some degree of fractionation in the different tissue samples. This differed from the PCB signature data which showed a strong degree of perpetuation in the different tissue types. The 279 280 highest levels of enantiomer enrichment for E1 were recorded for CB-95 in the liver of Bird A 281 (0.84) and kidney of Bird B (0.80). High levels of enrichment were also recorded for CB-95 in 282 the breast tissue of Bird A (0.69) and Bird B (0.76).

283 The contents of the proventriculus, duodenum and intestine all had a near racemic enantiomer 284 fraction (mean = 0.50 ± 0.03 (1 standard deviation)). Without wanting to over interpret this dataset, the results provide some insight as to where enantioselective processing occurs. The 285 286 food ingested contained near racemic enantiomer fractions which can be used as a baseline 287 for comparison. The process of absorption of PCBs into the proventiculus, duodenum and 288 intestines does not appear to result in significant fractionation. However this is not surprising considering that absorption of PCBs in the gastrointestinal tract is a passive transport process 289 290 so makes no contribution to the atropisomeric enrichment of chiral PCBs (Kania-Korwel and 291 Lehmler, 2016). The data reported here provides some of the clearest evidence to date to 292 suggest that fractionation occurs in the organs where metabolic biotransformation and 293 elimination of PCBs occurs (kidney and liver). This fractionation is also recorded in muscle tissue in the breast. The results also indicate that the enantiomer fraction is not consistent 294 295 within different organs from the same individual. This finding warrants further investigation to 296 establish how and where enantioselective fractionation occurs.

The results indicate that CB-95 is much more susceptible to fractionation than CB-136. This trend has also been identified by Megson et al. (2015) in humans. There are currently few studies that assess changes to the enantiomer fractions in different tissues. Chu et al. (2003) identified that PCBs 95, 132 and 149 are near racemic in human muscle, brain and kidney tissue whereas enrichment of CB-136 was recorded in rainbow trout (Wong et al., 2002, Buckman et al., 2006), mice (Warner et al., 2009) and rats (Kania-Korwel et al., 2010, Kania-Korwel et al., 2008a, Kania-Korwel et al., 2007, Kania-Korwel et al., 2008b).

304 4 Conclusions

305 Over 100 different PCBs were detected in guillemot tissue, with concentrations of the i7 PCBs in the low µg g⁻¹ range. Samples were obtained from 12 different tissue types and partially 306 digested food samples obtained from three different points within the gastrointestinal tract. 307 308 The highest PCB concentrations were identified in the lipid-rich gonads at concentrations of approximately one order of magnitude greater than those measured in other tissues. Whilst 309 310 PCB concentrations varied in different organs the relative proportions of PCBs were consistent 311 between the different tissue types. This represents a significant development in PCB signature 312 analysis in animals since it negates the inherent difficulties associated with comparing 313 concentration data. The results show that the collection of only 1 mL of blood represents a 314 useful, ethically sound, analytical method.

This study also highlights the ability to measure individual enantiomers with femtogram detection limits. The results provide some of the clearest evidence to date to suggest that enantioselective metabolism does occur in the kidneys and liver which results in enrichment of the E1 stereoisomer of CB-95 in these organs.

Finally, with the growing demand for ethical, non-lethal sampling strategies, this study illustrates that 1 mL samples of blood (<1 g) can be used to provide representative and detailed congener specific PCB data.

322 Acknowledgements

The authors would like to thank Peter Venn at RSPCA West Hatch for supplying guillemot carcases and Alec Kettle, Alan Griffiths and Tony Hadley (LECO) for their help and support with the research. Finally, David Megson would also like to thank Plymouth University for funding the initial part of this project which formed part of his PhD research.

327

328 References

329

330	Acampora, H., White, P., Lyashevska, O. and O'Connor, I. (2017) 'Presence of persistent
331	organic pollutants in a breeding common tern (Sterna hirundo) population in Ireland',
332	Environmental Science and Pollution Research, 24(14), pp. 13025-13035.

- Anderson, H. B., Evans, P. G. H., Potts, J. M., Harris, M. P. and Wanless, S. (2014) 'The diet
 of Common Guillemot Uria aalge chicks provides evidence of changing prey
 communities in the North Sea', *Ibis*, 156(1), pp. 23-34.
- Boumphrey, R. S., Harrad, S. J., Jones, K. C. and Osborn, D. (1993) 'Polychlorinated
 biphenyl congener patterns in tissues from a selection of British birds', *Archives of Environmental Contamination and Toxicology*, 25(3), pp. 346-352.
- Bradstreet, M. S. W. and Brown, R. G. B. (1985) 'Feeding ecology of the Atlantic Alcidae', in
 Nettleship, D.N. & Birkhead, T.R. (eds.) *The Atlantic Alcidae*. London, UK: Academic
 Press, pp. 263-318.
- Brown, T. A., Belt, S. T., Ferguson, S. H., Yurkowski, D. J., Davison, N. J., Barnett, J. E. F.
 and Jepson, P. D. (2013) 'Identification of the sea ice diatom biomarker IP25 and
 related lipids in marine mammals: A potential method for investigating regional
 variations in dietary sources within higher trophic level marine systems', *Journal of Experimental Marine Biology and Ecology*, 441, pp. 99-104.
- Buckman, A. H., Wong, C. S., Chow, E. A., Brown, S. B., Solomon, K. R. and Fisk, A. T.
 (2006) 'Biotransformation of polychlorinated biphenyls (PCBs) and bioformation of hydroxylated PCBs in fish', *Aquatic Toxicology*, 78(2), pp. 176-185.
- Chu, S. G., Covaci, A. and Schepens, P. (2003) 'Levels and chiral signatures of persistent
 organochlorine pollutants in human tissues from Belgium', *Environmental Research*,
 93(2), pp. 167-176.
- Fernie, K. J., King, R. B., Drouillard, K. G. and Stanford, K. M. (2008) 'Temporal and spatial patterns of contaminants in Lake Erie watersnakes (Nerodia sipedon insularum)
 before and after the round goby (Apollonia melanostomus) invasion', *Science of the Total Environment*, 406(1-2), pp. 344-351.
- Frame, G. M., Cochran, J. W. and Bowadt, S. S. (1996) 'Complete PCB congener
 distributions for 17 aroclor mixtures determined by 3 HRGC systems optimized for
 comprehensive, quantitative, congener-specific analysis', *Hrc-Journal of High Resolution Chromatography*, 19(12), pp. 657-668.
- 361 Hansen, L. G. (1999) The ortho side of PCBs. Kluwer Academic Publishers.

- Harner, T., Wiberg, K. and Norstrom, R. (2000) 'Enantiomer fractions are preferred to
 enantiomer ratios for describing chiral signatures in environmental analysis',
 Environmental Science & Technology, 34(1), pp. 218-220.
- Harrad, S., Ren, J. and Hazrati, S. (2006) 'Chiral signatures of PCB#s 95 and 149 in indoor air, grass, duplicate diets and human feces', *Chemosphere*, 63, pp. 1368-1376.
- Harris, M. P. and Wanless, S. (2004) 'Common Guillemot (Uria aalge)', in Mitchell, P.I.,
 Stephen, F.N., Ratcliffe, N. & Dunn, T.E. (eds.) Seabird Populations of Britain and *Ireland*: T. & A.D. Poyser, London, UK, pp. 350-363.
- Helsel, D. R. (2006) 'Fabricating data: How substituting values for nondetects can ruin results, and what can be done about it', *Chemosphere*, 65(11), pp. 2434-2439.
- Hu, D., Martinez, A. and Hornbuckle, K. C. (2011) 'Sedimentary Records of Non-Aroclor and
 Aroclor PCB mixtures in the Great Lakes', *J Great Lakes Res*, 37(2), pp. 359-364.
- Hu, D. F., Martinez, A. and Hornbuckle, K. C. (2008) 'Discovery of Non-Aroclor PCB (3,3 'Dichlorobiphenyl) in Chicago Air', *Environmental Science & Technology*, 42(21), pp.
 7873-7877.
- Jaspers, V. L. B., Megson, D. and O'Sullivan, G. (2013) 'Chapter 7, POPs in the Terrestrial
 Environment', in O'Sullivan, G. & Sandau, C.D. (eds.) *Environmental Forensics for Persistent Organic Pollutants*: Elsevier.
- Jepson, P. D., Deaville, R., Barber, J. L., Aguilar, A., Borrell, A., Murphy, S., Barry, J.,
 Brownlow, A., Barnett, J., Berrow, S., Cunningham, A. A., Davison, N. J., ten
 Doeschate, M., Esteban, R., Ferreira, M., Foote, A. D., Genov, T., Gimenez, J.,
 Loveridge, J., Llavona, A., Martin, V., Maxwell, D. L., Papachlimitzou, A., Penrose,
 R., Perkins, M. W., Smith, B., de Stephanis, R., Tregenza, N., Verborgh, P.,
 Fernandez, A. and Law, R. J. (2016) 'PCB pollution continues to impact populations
 of orcas and other dolphins in European waters', *Scientific Reports*, 6.
- Johnson, G. W., Ehrlich, R., Full, W. and Ramos, S. (2007) 'Chapter 7: Principal
 components analysis and receptor models in environmental forensics', in Morrison,
 R. & Murphy, B.L. (eds.) *An Introduction to Environmental Forensics. 2nd Edition.*Amsterdam: Elsevier, pp. 207-272.
- Kania-Korwel, I., El-Komy, M., Veng-Pedersen, P. and Lehmler, H. J. (2010) 'Clearance of
 Polychlorinated Biphenyl Atropisomers is Enantioselective in Female C57Bl/6 Mice',
 Environmental Science & Technology, 44(8), pp. 2828-2835.
- Kania-Korwel, I., Hornbuckle, K. C., Robertson, L. W. and Lehmler, H. J. (2008a) 'Dosedependent enantiomeric enrichment of 2,2 ',3,3 ',6,6 '-hexachlorobiphenyl in female
 mice', *Environmental Toxicology and Chemistry*, 27(2), pp. 299-305.
- Kania-Korwel, I. and Lehmler, H. J. (2016) 'Chiral polychlorinated biphenyls: absorption,
 metabolism and excretion-a review', *Environmental Science and Pollution Research*,
 23(3), pp. 2042-2057.
- Kania-Korwel, I., Shaikh, N. S., Hornbuckle, K. C., Robertson, L. W. and Lehmler, H. J.
 (2007) 'Enantioselective disposition of PCB 136 (2,2 ',3,3 ',6,6 '-hexachlorobiphenyl)
 in C57BL/6 mice after oral and intraperitoneal administration', *Chirality*, 19(1), pp. 56-66.
- Kania-Korwel, I., Xie, W., Hornbuckle, K. C., Robertson, L. W. and Lehmler, H. J. (2008b)
 'Enantiomeric enrichment of 2,2 ',3,3 ',6,6 '-hexachlorobiphenyl (PCB 136) in mice
 after induction of CYP enzymes', *Archives of Environmental Contamination and Toxicology*, 55(3), pp. 510-517.
- Karjalainen, A., Paakkonen, J. P. J. and Karjalainen, J. (2006) 'Tissue-specific and whole fish accumulation of polychlorinated biphenyls by juvenile Baltic salmon (Salmo salar
 L.) after oral gavage exposure', *Boreal Environment Research*, 11(6), pp. 421-430.
- King, T. L., Yeats, P., Hellou, J. and Niven, S. (2002) 'Tracing the source of 3,3 'dichlorobiphenyl found in samples collected in and around Halifax Harbour', *Marine Pollution Bulletin*, 44(7), pp. 590-596.
- Kodavanti, P. R., Ward, T. R., Derr-Yellin, E. C., Mundy, W. R., Casey, A. C., Bush, B. and
 Tilson, H. A. (1998) 'Congener-specific distribution of polychlorinated biphenyls in

416 brain regions, blood, liver, and fat of adult rats following repeated exposure to Aroclor 417 1254', Toxicology and Applied Pharmacology, 153, pp. 199-210. 418 Maervoet, J., Beck, V., Roelens, S. A., Covaci, A., Voorspoels, S., Geuns, J. M. C., Darras, V. M. and Schepens, P. (2005) 'Uptake and tissue-specific distribution of selected 419 420 polychlorinated biphenyls in developing chicken embryos', Environmental Toxicology 421 and Chemistry, 24(3), pp. 597-602. 422 Megson, D., Brown, T. A., Johnson, G. W., O'Sullivan, G., Bicknell, A. W. J., Votier, S. C., 423 Lohan, M. C., Comber, S., Kalin, R. B. and Worsfold, P. J. (2014) 'Identifying the 424 provenance of Leach's storm petrels in the North Atlantic using polychlorinated 425 biphenyl signatures derived from comprehensive two-dimensional gas 426 chromatography with time-of-flight mass spectrometry', Chemosphere, 114, pp. 195-427 202. Megson, D., Focant, J. F., Patterson, D. G., Robson, M., Lohan, M. C., Worsfold, P. J., 428 429 Comber, S., Kalin, R., Reiner, E. and O'Sullivan, G. (2015) 'Can polychlorinated 430 biphenyl (PCB) signatures and enantiomer fractions be used for source identification 431 and to age date occupational exposure?', Environment International, 81, pp. 56-63. 432 Megson, D., Kalin, R. B., Worsfold, P., Gauchotte-Lindsay, C., Patterson Jr, D. G., Lohan, M. 433 C., Comber, S., Brown, T. A. and O'Sullivan, G. (2013) 'Fingerprinting polychlorinated biphenyls in environmental samples using comprehensive two-434 435 dimensional gas chromatography with time-of-flight mass', Journal of 436 Chromatography A, 1318, pp. 276-283. 437 Megson, D., Robson, M., Jobst, K. J., Helm, P. A. and Reiner, E. J. (2016) 'Determination of 438 Halogenated Flame Retardants Using Gas Chromatography with Atmospheric 439 Pressure Chemical Ionization (APCI) and a High-Resolution Quadrupole Time-of-440 Flight Mass Spectrometer (HRqTOFMS)', Analytical Chemistry, 88(23), pp. 11406-441 11411. 442 Muir, D. C. G., Norstrom, R. J. and Simon, M. (1988) 'Organochlorine contaminants in arctic 443 marine food-chains - accumulation of specific polychlorinated-biphenyls and 444 chlordane-related compounds', Environmental Science & Technology, 22(9), pp. 445 1071-1079. 446 Oki, M. (1983) 'Recent Advances in Atropisomerism', Topics in Stereochemistry, 14, pp. 1-447 81. 448 Robson, M. and Harrad, S. (2004) 'Chiral PCB signatures in air and soil: Implications for atmospheric source apportionment', Environmental Science & Technology, 38(6), pp. 449 450 1662-1666. 451 Rodenburg, L. A., Guo, J., Du, S. and Cavallo, G. J. (2010) 'Evidence for unique and 452 ubiquitous environmental sources of 3,3'-dichlorobiphenyl (PCB 11)', Environ Sci 453 Technol, 44(8), pp. 2816-21. 454 Warner, N. A., Martin, J. W. and Wong, C. S. (2009) 'Chiral Polychlorinated Biphenyls Are Biotransformed Enantioselectively by Mammalian Cytochrome P-450 Isozymes to 455 456 Form Hydroxylated Metabolites', Environmental Science & Technology, 43(1), pp. 457 114-121. 458 Wong, C. S., Lau, F., Clark, M., Mabury, S. A. and Muir, D. C. G. (2002) 'Rainbow trout 459 (Oncorhynchus mykiss) can eliminate chiral organochlorine compounds 460 enantioselectively', Environmental Science & Technology, 36(6), pp. 1257-1262. Wu, X. A., Kammerer, A. and Lehmler, H. J. (2014) 'Microsomal Oxidation of 2,2 ',3,3 ',6,6 '-461 Hexachlorobiphenyl (PCB 136) Results in Species-Dependent Chiral Signatures of 462 the Hydroxylated Metabolites', Environmental Science & Technology, 48(4), pp. 463 2436-2444. 464

465